A Final Author Comment.

The manuscript HESSD-11-183-2014 "Quantifying groundwater dependence of a sub-polar lake cluster in Finland using isotope mass balance approach" by E. Isokangas et al.

We very much appreciate thorough reviews by two anonymous referees. Their insightful and comprehensive comments helped us to make numerous changes in the way how our data are interpreted, presented and discussed. Isotope mass balance calculations have been refined through including isotopic composition of local atmospheric moisture derived from isotope signal of local precipitation and through assessing the isotopic composition of total inflow individually for each of the studied lakes. Moreover, isotope mass balance calculations were run also deuterium data. The manuscript is in much better shape now. We followed large majority of the reviewer's comments and suggestions when preparing the revised version of the manuscript. In some points, however, we do maintain our views and opinions. Detailed comments addressing all questions/comments of the reviewer #1 are listed below.

Anonymous Referee #1

General comment:

1. C3857, line 1-9: *I have several fairly substantial reservations to publishing the manuscript as it is currently written. The most serious of these are the simplified assumptions embedded in the calculation of the total inflow to evaporation ratio (Itot / E). The assumptions are logical and necessary in this study, but they should be investigated more fully before the manuscript can be published. They include 1) assumptions of steady-state isotopic signature in the lakes; 2) implicit assumptions that all of the lakes are completely mixed; 3) assumptions that all inflowing water has the same isotopic signature. All of these assumptions should be discussed in more detail and in the case of number 2 & 3, the sensitivity of the outcomes should be tested against these assumptions. Beyond that, the paper is well-written and well-presented. A minor point where the manuscript could be improved is that the motivation for the study is described very clearly at the end of the Introduction, but it is never revisited later in the paper. The authors have some important results which could help inform management of these lakes and a few sentences placing the results in that context would improve the overall manuscript.*

We appreciate very much this comment. Our original idea was to present a simplified, operational approach to derive the parameters characterizing the dependence of Rokua lakes on groundwater, based on the results of isotope survey made between July 29 and August 27, 2013. We have now modified the text and discussed especially the assumptions made in more detail.

1. <u>Assumption: Isotope steady-state of the studied lakes</u>

This assumption was behind all our calculations, although it was not discussed in the text. In the revised manuscript we address this problem. We think that this assumption is in fact well-justified. The reason is that natural systems such as Rokua lakes had a lot of time in the past to come up to an isotopic steady-state,

reflecting their mean hydrological status and local climate. What we have observed in our study were seasonal fluctuations around these steady-state values induced mostly by seasonal ice cover. Three months following disappearance of ice cover in mid of May 2013 apparently represent sufficiently long time period to bring the lakes close enough to the steady-state values, to justify the steady-state assumption used in the calculations.

2. Assumption: Completely mixed lakes

We were aware of the problem related to the isotopic homogeneity of lakes. Therefore, for larger and/or deeper lakes several water samples were collected for isotope analyses (cf. section 3.1., lines 13-18 of the original manuscript). Deeper lakes showed expected enrichment of the shallow layer with respect to the bottom waters (0.7‰ on the average for $\delta^{18}O$). For balance calculations presented in the manuscript we used average δ values obtained on the basis of all water samples collected in the given lake including depth profiles. We consider this simplified approach justified in the light of methodology presented in the manuscript, tailored to single sampling campaign of a large number of lakes. See also comment 5. p. 9191, lines 8-11 about completely mixed lakes.

3. Assumption: Isotopic composition of the total inflow to the studied lakes

We admit that our assumption setting identical isotopic composition of the total inflow for all studied lakes was perhaps a too-simplistic approach. This was corrected in the revised version. Iterative procedure of calculating I_{TOT}/E ratios was adopted. In the first step groundwater fluxes to the lakes without visible surface inflow were calculated assuming that the isotopic composition of the total inflow equals the coordinates of the intercept of LMWL and LEL (–14.1‰, –100.0‰). In the second step, the calculations of I_{TOT}/E ratios were repeated, this time with δ_I calculated individually for each lake as a flux-weighted average of local precipitation ($\delta^{18}O_P$= –14.1 ‰ and $\delta^{2}H_P$ = –100 ‰) and groundwater input ($\delta^{18}O_{GW}$= –13.1 ‰ and $\delta^{2}H_{GW}$ = –95 ‰) calculated in the preceding step. The procedure was repeated until the change of δ_I was smaller than the adopted uncertainty of isotope measurements.

For lakes with visible surface inflow a slightly different approach was adopted to calculate δ_I. The first step was identical as in the case of lakes without visible surface inflow. In the second step, the total outflow (groundwater plus surface water) from the upstream lake was determined as a difference $I_{TOT} – E$. Then, the isotopic composition of the total inflow to the downstream lake was determined as a weighted average of three components: (i) local precipitation, (ii) surface inflow from the upstream lake, and (iii) groundwater inflow of constant isotopic composition (–13.1 ‰). Since discharges of surface inflows were not measured, it was assumed arbitrarily that they constitute 25 % of total outflows from upstream lakes. Surface inflows determined in the previous step were assumed to have the isotopic composition of upstream lakes. The calculations were repeated until the change of δ_I was smaller than the adopted uncertainty of isotope measurements.

2. C3857, lines 11-13: *A minor point where the manuscript could be improved is that the motivation for the study is described very clearly at the end of the Introduction, but it is never revisited later in the paper.*

Accepted. Appropriate corrections/additions were made in the revised version.

Detailed comments:

3. p. 9186, line 11: *"a. . . one-off survey" – unscientific language.*

Has been corrected.

4. p. 9816, line 15: *"more trophic" – change to "more productive".*

Done.

5. p. 9191, lines 8-11: *This approach assumes that all of the lakes mix completely. How does changing this assumption change the outcome of the study? (see Stets et al. 2010).*

Stets et al. (2010) used Priestley-Taylor method to calculate evaporation flux. In the framework of this approach the average temperature of the lake is needed. In our calculations of *E* we adopted mass transfer method which exploits the vapor-pressure gradients above the evaporating surface. Therefore, the surface water temperature is of importance here. However, we were concerned with the mixing of the studied lakes from the isotope perspective. This is the reason why we haven't even considered the mixing of the lake in this context.

6. p. 9192, line 3: *dLS assumes that the lakes are in isotopic steady state, but this is almost certainly not the case (see figure 6). How does this assumption affect the study?*

See assumption 1 above.

7. p. 9192, line 11: *"If . . .the isotopic composition of atmospheric water vapour is in isotopic equilibrium with the total inflow" – This seems like a poor assumption because most of the groundwater wells in figure 4 are close to the average recharge, not the current atmospheric conditions. Why not set the isotopic composition of atmospheric water vapor equivalent to precipitation at or near the time of the survey? (see Gibson et al. 2002)*

Correct. In the revised manuscript we refrain from this simplifying assumption and calculate δ_A assuming isotopic equilibrium between local precipitation and atmospheric water vapor at ground level temperature for the period in question (June-August 2013).

8. p. 9192: *It is worth pointing out that isotopic enrichment (delta d) was calculated using measured dLS and assumed dTI.*

Done.

9. p. 9192: *I didn't see anywhere in the manuscript where lake water budgets were calculated using 2H. It may offer some insight. If the authors decide not to use 2H for that purpose, the references to 2H in the methods section should be removed.*

In the revised version we present the results of isotope balance calculations also for deuterium. The total inflow-to-evaporation ratios derived from ^{2}H-based balance turned out to be ca. 5 % higher on the average than those derived from ^{18}O-based balance. For the *G* index this difference is approximately 0.8 %. The MTT values were ca. 4.3 % higher for ^{18}O-based balance.

10. p. 9194, lines 13-16: *The most direct evidence for evaporation in the groundwater flowpath is the fact that a number of wells lie along the LEL.*

Yes, this is true. The revised text has been modified to reflect that.

11. p. 9196, line 4: *"Steady-state isotope enrichment is primarily controlled by water balance...." Again, I don't see any evidence that the lakes are in steady state. The ones most likely to meet this condition are the very low Itot/E lakes. It would be worth comparing the results of lakes with very low Itot/E to those with high Itot/E.*

See point 1 above and the revised text.

12. p. 9196, line 23-24: *"Introducing the assumption that the total inflow to each lake also contains a groundwater component" – I think the authors have good evidence that these lakes do have groundwater inputs. In which case, the governing assumptions used to calculate h_N are not correct. The authors rightly include a sensitivity analysis of this assumption, but it should be a more fundamental part of the overall modeling effort.*

Not relevant anymore. See point 1 above and the revised text.

13. p. 9196, line 26-27: *"d18O value of the total inflow -14.1 per mil". The authors show that the isotopic composition of groundwater can evolve in this landscape. Why not incorporate some of the spatial heterogeneity into assumptions about the d18O of groundwater? Or at least provide sensitivity analysis of how this assumption affects the results.*

Corrected. In the revised version we provide also the sensitivity analysis with respect to possible changes of $\delta^{18}O$ of groundwater.

14. p. 9196, line 26-27: *A quick sensitivity analysis using equation 5 and five random lakes from Table 2 shows that assuming that the d18O of groundwater is -13.1 per mil instead of -14.1 per mil introduces changes ranging from 60 to 143 % in estimated Itot/E. Maybe a 1 per mil error on dTI is too high, but the authors should address this assumption more thoroughly. A slight reformulation of the LMWL & LEL lines (Figure 4) could allow the error on average recharge to be approached statistically (i.e. as std err on the intercept). Although, just looking at figures 4 and 5, I suspect that the groundwater flow system is fairly heterogeneous in its isotopic composition.*

In the revised version δ_T is calculated individually for each lake (cf. point 1 above). Still, δ_{GW} was maintained constant (–13.1 ‰, –95 ‰), simply because we do not have sufficiently detailed knowledge of this parameter with respect to each studied lake. Sensitivity analysis included in the revised version highlights potential impact of changes of this parameter on the *G* index.

15. p. 9197, line 13: – *"The link between MTT and the Itot/E ratio is much weaker" – I wonder if E/Itot vs. MTT would provide more insight?*

Not really.

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Anonymous Referee #2

Specific comments

Section 3.3. Evaporation estimation

1. The authors should indicate the time step used in the evaporation estimation using the mass transfer approach (hourly? daily? monthly?) and as to why this approach is chosen (eg availability of hourly data?). Normally, mass transfer approach works very well for instantaneous evaporation estimation where the dependence of vapor pressure gradient (es-ea) and wind speed in equation 1 is assumed to be nonexistent or at least low. Over long time integrated data the two parameters are interdependent and may not give good evaporation result (as in Dingman, 1994).

This is a good point. The averaging period used in the evaporation calculations was one day. This was chosen because averaging over this period does not introduce large error to the calculations (Dingman, 2008). Averaging over one month period would likely induce significant bias. As the Referee pointed out, vapour pressure differences and wind speed may correlate significantly in this case. This comment will be added to the revised manuscript.

2. What is the relation between the empirical constant n in equation 2 (which follows a power law) and the turbulent parameter 'n' implied in the Ck estimation-equation 3. Ck in the Isotope equation in this work is based on n = 0.5. For this reason the equation used in estimation of Ck needs to be elaborated (incorporating the turbulent parameter n) later in section 4.4- line 18 in addition to referring to Gonfiantini, 1986. The Ck (for d18O) values used in this work would be convincing if d2H has been used for the water budget computations using the Ck 2H proposed by Gonfiantini, 1986 and get comparable water budget estimation.

This is not the same "n". The term "n" (also "α" is used in the literature) used in eq. (2) is the friction coefficient linked to the vertical profile of wind speed within the boundary layer between the surface and the free atmosphere. In our case we calculated wind speed at 2 m height using the meteo data from 10 meters above the surface. We used in our calculations 0.15 which is the friction coefficient for grassland

(Bañuelos-Ruedas et al., 2010). Friction coefficient for lakes and oceans is 0.10, but we estimated that 0.15 is more suitable for our situation since the lakes have relatively small surface areas and they are surrounded by forests which lowers the wind speed. If 0.10 is used instead, the calculated E values are ca. 8 % higher.

In the framework of the Craig-Gordon model describing isotope effects during evaporation of water the term "*n*" is introduced to quantify the impact of dynamic conditions at the interface on transport of different isotopic molecules of water through the layer of air above the water-air interface (see e.g. Craig and Gordon, 1965; Gat, 2010; Horita et al., 2008). It can be shown that kinetic fractionation associated with this transport can be expressed by the following formula:

$$\Delta\varepsilon \cong n \cdot (1-\frac{D_i}{D})(1-h_N) = C_k \cdot (1-h_N)$$

Where D and D_i stand for molecular diffusivities of normal and heavy water molecules, respectively, in air, and h_N is relative humidity of the atmosphere normalized to the temperature of the evaporating water. The term "*n*" changes from 0.5 (rough evaporating surfaces, turbulent conditions) to 1 (maximum kinetic fractionation - transport accomplished only through molecular diffusion). The ratio D_i/D was measured in the laboratory (Merlivat, 1978). Wind tunnel experiments (Vogt, 1976) have shown that C_k is equal 14.2 ‰ for ^{18}O and 12.5 ‰ for ^{2}H. This corresponds to n = 0.5 i.e. rough evaporating surface and dominance of turbulent transport. Numerous field studies using isotopes to quantify lake balance have shown that these values are well suited for describing kinetic fractionation during evaporation of surface water bodies in continental settings.

Budget of Rokua lakes based on ^{2}H data is presented in the revised version of the manuscript and yields the results similar to the ^{18}O-based budget (cf. point 9 in the reply to Reviewer #1).

To avoid confusion, in the revised manuscript the term "*n*" in eq. (1) is labeled as "*θ*".

Section 4.2. Local isotopic composition

1. Please replace 'intersect' by 'intercept' in section 4.2 line 11

Done.

2. Line 16 What is the source of 'evaporation signal' and how is linked to the enrichment discussed in line 21. One major issue here is what is the role of these enriched groundwaters in feeding next downstream lakes? If this enriched groundwater enters into the next downstream lake it has a bearing on the interpretation of the I/E ratio of the next lake as this affects the _IT of the next lake downstream. To avoid this confusion the authors need to briefly provide groundwater table map (if available) or at least the geographic occupation of the different lakes and discuss if such lake interconnection exists via groundwater.

The expression "evaporation signal" and "isotope enrichment" have the same meaning.

We admit that the question of groundwater input to a given lake which seeps from upstream lake and thus is enriched isotopically, is an important issue. However we do not have sufficiently detailed observations to

consider this explicitly for each individual lake. We took this into account only for lakes which have visible surface inflow originating from an upstream lake (cf. point 1 in the reply to comments of Referee #1). For the rest of the studied lakes we assumed constant isotopic composition of groundwater inflow equal the average isotope signature of groundwater measured in piezometers in the study area. The question of potential interconnections between lakes via groundwater is discussed in the revised version of the manuscript. In the revised version of the manuscript we also show the results of sensitivity study with respect to the assumed isotopic composition of groundwater inflow.

Section 4.4. Quantifying groundwater dependence of the studied lakes

1. A brief discussion about hydrology of the terminal lake Kissalampi (pond/lake) is needed so as to convince readers this lake can be used as an index terminal lake. The lake has been chosen as a terminal lake based on the assumption that it shows the highest enrichment. To substantiate this the authors need to give information about a) if the geology underneath the lake bottom allows this assumption –eg presence of clay deposits, b) if the lake occupy the lowest place in the region where all waters converge but little chance for leakage and c) compare the isotopic composition of the lake with respect to the hypothetical 'limiting isotopic composition-_ (Gonfiantini, 1986)' for the region and d) this is not a changing volume/shrinking lake during that time of the year.*

We appreciate this comment. Closer examination of the location of Kissalampi pond and calculation of the expected steady-state isotope enrichment of this lake for the case when it would indeed operate as terminal lake, revealed that the lake is in fact less enriched isotopically than expected (measured $\delta^{18}O$: –5.6 ‰, expected $\delta^{18}O$: –2.9 ‰). Therefore, we had to abandon this tempting idea of using Kissalampi lake data to derive h_N. Instead, we rely on meteorological information and measured temperature of the lakes in deriving this important parameter.

2. In section 4.4 line 26 reads "The lowest G value (27.8%) was obtained for Kissalampi pond, which is comparable to a terminal lake." This argument appears incorrect. The G reflect the inflow index not the outflow which governs whether a lake is terminal or not. This needs correction.

Correct, but not relevant anymore (see the previous comment).

3. Figure 8 and 9 add little value in the manuscript. If the authors claim this could add more value, elaboration is needed.

We agree. The figures were removed. The relations shown are described in the text.

Section 5. Conclusion

One or two lines of argument stating the presence or absence of any pattern in groundwater dependence by geographic position (upland, midland, center etc) may be interesting.

The groundwater dependence of the studied lakes does not follow any clear spatial pattern, since the existence of surface water outflow complicates the situation.

Other comments

Figure 3. In the line marks please use multiple of 5 in the vertical axis

Figure 3 has been modified.

Abstract- Line 3: please add 'and quantity' after 'role'

"and extent" was added.

In evaporation estimation, section 3.3, the authors used temperature of the surface part of the lake water body to estimate the saturated vapor pressure es (normalization has been applied to temperature of surface part of lake water body). In isotope section the humidity has been normalized to temperature of evaporating surface. The two (mass transfer and evaporative isotope fractionation) are based on similar theory and one temperature value should be used in both cases, in principle.

In fact the same temperature (temperature of the surface layer of the lakes) was used in both types of calculations.

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Quantifying groundwater dependence of a sub-polar lake cluster in Finland using an isotope mass balance approach

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Abstract

A stable isotope study of 67 kettle lakes and ponds situated on an esker aquifer (90 km^2) in northern Finland was carried out to determine the role and extent of groundwater inflow in groundwater-dependent lakes. Distinct seasonal fluctuations in the $\delta^{18}O$ and $\delta^{2}H$ values of lakes are the result of seasonal ice cover prohibiting evaporation during the winter. An iterative isotope mass balance approach was used to calculate the inflow-to-evaporation ratios (I_{TOT}/E) of all 67 lakes during the summer of 2013 when the isotopic compositions of the lakes were approaching a steady-state. The balance calculations were carried out independently for ^{2}H and ^{18}O data. Since evaporation rates were derived independently of any mass balance considerations, it was possible to determine the total inflow (I_{TOT}) and mean turnover time (MTT) of the lakes. Furthermore, the groundwater seepage rates to all studied lakes were calculated. A quantitative measure was introduced for the dependence of a lake on groundwater (*G* index) that is defined as the percentage contribution of groundwater inflow to the total inflow of water to the given lake. The *G* index values of the lakes studied ranged from ca. 39 % to 98 %, revealing generally large groundwater dependency among the studied lakes. This study shows the effectiveness of applying an isotope mass balance approach to quantify the groundwater reliance of lakes situated in a relatively small area with similar climatic conditions.

Key words

Stable isotopes of water, groundwater-surface water interaction, lake hydrology, mean turnover time (MTT), inflow-to-evaporation ratio (I_{TOT}/E), thermal imaging

1. Introduction

The characterisation of groundwater dependent ecosystems (GDEs) is a requirement of the Groundwater Directive (EC, 2006). These systems are often complex and their hydrology and contact with aquifers are not well established. Lakes can be dependent on groundwater directly or indirectly, and this dependence can vary over time (Kløve et al., 2011). Understanding groundwater and lake water interaction is important not only for water resource management (Showstack, 2004), but also for understanding the ecology and eutrophication of lakes, since groundwater may be a key element in the lake nutrient balance (Ala-aho et al., 2013; Belanger et al., 1985; Brock et al., 1982; Kidmose et al., 2013). Furthermore, the vulnerability of lakes to pollution can be controlled by their dependency on groundwater (Kløve et al., 2011). Methods such as seepage meters (Ala-aho et al., 2013; Rosenberry et al., 2008, 2013), environmental tracers (e.g. Dinçer, 1968; Shaw et al., 2013; Stets et al., 2010; Yehdegho et al., 1997; Zuber, 1983) and numerical modelling (e.g. Krabbenhoft et al., 1990; Stichler et al., 2008; Winter and Carr, 1980) can be used to determine the groundwater reliance of lakes.

Heavy stable isotopes of water (^{18}O, ^{2}H) can be considered as ideal tracers for studying the hydrological cycle (e.g. Clark and Fritz, 1997). Fractionation of isotopes of water is the very factor enabling their use in hydrological studies, as it governs the changes in isotopic abundances within the water cycle (Gat, 2010). At a global scale, the ^{2}H and ^{18}O isotope composition of meteoric waters cluster along the line called the global meteoric water line (GMWL), first determined by Craig (1961): $\delta^2 H = 8 \cdot \delta^{18} O + 10$. Locally, this linear relationship may have a slightly different form (local meteoric water line - LMWL). Evaporation from an open water body fractionates isotopes so that the remaining liquid phase is enriched in both ^{2}H and ^{18}O in proportion with their effective fractionation factors accompanying this process. Consequently, the isotopic composition of the evaporating water body evolves in the δ-space along the line known as the local evaporation line (LEL), whose slope is significantly smaller than that characterising the local or global meteoric water lines. The position of the isotopic composition of lake water along this line is strongly related to the water balance of the lake (e.g. Gat, 1996; Gibson and Edwards, 2002; Rozanski et al., 2001).

The methodology of isotope-aided studies of the water balance of lakes has been thoroughly discussed in a number of review papers and textbooks (e.g. Darling et al., 2005; Froehlich et al., 2005; Gat and Bowser, 1991; Gat, 1995; Gonfiantini, 1986; Rozanski et al., 2001). Although several authors have applied isotope techniques in studying lakes in cold climates (Gibson and Edwards, 2002; Gibson,

2002; Gibson et al., 1993; Jonsson et al., 2009; Turner et al., 2010; Yi et al., 2008), mostly in Canada and northern Sweden, these studies were generally focused on lakes spread over large areas.

The central aim of this study was to quantify the groundwater dependence of 67 kettle lakes and ponds situated across a relatively small area (90 km^2) of the Rokua esker aquifer occupying a large glaciofluvial deposit in northern Finland. To quantify the extent of the interaction between the aquifer and the lakes, a dedicated isotope study was launched in 2013. This was part of comprehensive investigations (2010-2012) aimed at understanding the hydrology of an esker aquifer area where some of the kettle lakes and ponds have suffered from water level decline or eutrophication. Since the seasonal isotopic behaviour of the selected lakes in the study area was already fairly well understood based on the data collected from 2010 to 2012, it was decided to conduct a large-scale one-time survey of the isotopic composition of all 67 lakes on the Rokua esker in order to quantify their dependence on groundwater. Ala-aho et al. (2013), who studied 11 lakes on the esker, showed that the water levels in closed-basin seepage lakes have more fluctuations than the drainage lakes, which have more stable water levels. On the other hand, the drainage lakes are more productive. Their study also showed that subsurface flow can transport phosphate to lakes. Therefore, it was important to quantify the groundwater dependence of all lakes on the esker and propose an appropriate index reflecting this dependence. The large-scale field campaign conducted in July and August 2013 comprised the sampling of water in all 67 lakes for isotope analyses, combined with continuous temperature measurements and aerial thermal imaging of the lakes.

2. The study area

The Rokua esker aquifer area, situated in Northern Finland, was formed during the transition period from Late Glacial to Holocene, between approximately 12,000 and 9,000 years ago (Tikkanen, 2002). As ice retreated, a long ridge formation, consisting mainly of fine and medium sand (Pajunen, 1995), was shaped. Ancient sea banks surrounding the esker show that the esker was originally an island that gradually rose from the sea (Aartolahti, 1973). Today, the highest elevation of the esker is 100 m above the surrounding low-lying peatlands and the layer thickness of sand ranges from 30 m to more than 100 m above the bedrock. Sea banks, dunes and kettle holes form a rolling and geologically unique terrain (Aartolahti, 1973). Kettle holes were formed when ice blocks were buried in the ground and, as they melted, left depressions in the landscape. The ground surface of the esker is mainly lichen-covered pine forests. Hydrologically, the Rokua esker is an unconfined aquifer, one of the

largest in Finland, and it has two regional groundwater mounds (Rossi et al., 2014). The recharge area of the aquifer is 90 km^2 and the discharge zones are situated in the surrounding peatlands, which partially confine the aquifer (Rossi et al., 2012).

Kettle holes – long and narrow depressions – give Rokua esker its distinct character. The sizes of these kettle holes vary. They can be 1 to 80 m deep, between 10 m and 1.5 km long, and 0.4 km wide (Aartolahti, 1973). Most of the kettle holes are now dry, but due to the influence of groundwater in the past, peat has accumulated at the bottom of them, creating kettle hole mires (Pajunen, 1995). However, the alternating topography of the area is reflected in the existence of approximately 90 lakes or ponds, referred to as kettle lakes or ponds. Peat started to accumulate in the border regions of the lakes more than 8,000 years ago, so most of the kettle lakes and ponds are partly paludified (Pajunen, 1995). Nevertheless, the majority of the lakes and ponds are characterised by their crystal clear water, which attract people; number of holiday homes and hotels are located on the lakeshores. The lakes are nowadays widely used for different recreational activities, such as swimming, fishing and scuba diving (Anttila and Heikkinen, 2007). The uniqueness of the glaciofluvial formation of Rokua, in which the actions of ice, water and wind can be seen, has been recognised in many ways. Some of the Rokua esker is protected by Natura 2000 and by the Finnish nature reserve network. Rokua was recently chosen to be part of the UNESCO GeoPark Network and is currently the northernmost region in this network.

3. Materials and methods

3.1. Hydrological measurements and thermal imaging

During 2010-2012, 11 lakes, 13 piezometers and 11 streams were sampled in the study area four times per year to analyse the stable isotopic composition of water, nutrients, water quality parameters (T, pH, E.C., O_2) and geochemical parameters (silica, major cations and anions) (Fig. 1). During the field campaign conducted in July and August 2013, a total of 67 lakes and ponds were surveyed for the same parameters and thermal images of lakes taken from the air in a helicopter using a FLIR thermal camera. In addition, composite (monthly) precipitation samples were collected during the open water season at the station on the esker during 2010-2013. Precipitation samples for winter were collected once a year before snowmelt by taking a uniform sample of the whole snowpack depth.

Water quality parameters were analysed in the field using WTW Multi 3430 or Multi 350i meters for temperature, oxygen, EC and pH. Samples of lake water were collected with a Limnos sampler,

approximately 1 m below the water surface and 1 m above the bottom of the lake. If the depth of the lake was less than 2 m, only one sample from the depth of 1 m was taken and if it was more than 20 m, samples were taken from the middle of the water profile as well. Depending on their shape and size, the lakes had between 1 and 4 sampling locations. Stream samples were collected by submerging a bottle in water, facing upstream. Piezometers were pumped for at least 10 minutes prior to taking groundwater samples or until the colour of the water was clear. The samples were collected one metre below the water table. All sampling bottles (HDPE) were rinsed with the sampled water prior to filling. Samples for isotope analyses were stored in the dark at a reduced temperature (4 °C ± 2 °C).

The isotopic composition of water samples was analysed using CRDS technology with a Picarro L2120-i analyser. Samples with visible colour or suspended matter were filtered (pore size 25 µm) prior to analysis. The measured $^{2}H/^{1}H$ and $^{18}O/^{16}O$ isotope ratios are reported as relative deviations from the VSMOW standard. Typical uncertainty of the reported $\delta^{18}O$ and $\delta^{2}H$ values are ±0.1 ‰ and ±1.0 ‰ respectively.

Lake water temperature was measured continuously during the ice-free period in 2013 for two lakes on the Rokua esker: Ahveroinen (3.3 ha) and Saarinen (15.3 ha). Hobo loggers (pendant temperature data logger UA-001-08 and conductivity logger U24-001, accuracy 0.1 °C) were installed 50 cm below the lake surface. In addition, the surface water temperature for lake Oulujärvi (92800 ha) was obtained from the database of the Finnish Environmental Institute (2013). Lake Oulujärvi is located east, next to the study site, 1 km from the easternmost lake studied. Thermal imaging of the lakes was conducted on 5 August 2013 using a Flir Thermacam P-60 thermal camera. This camera had 320 x 240 pixel sensor resolution and an opening of 24 °. It covered the electromagnetic spectrum from 7.5 to 13 µm. The imaging was taken by helicopter 150 meters above the lakes. The image data were correlated to the predominant weather conditions (temperature and relative humidity) with data from the FMI Pelso weather station measured every 10 minutes.

Depth profiling was undertaken in the lakes for which no depth contour lines are available (National Land Survey of Finland, 2010a). It was carried out with a portable depth-sounding radar (resolution 0.1 m) or with a measuring cable and GPS system. Typically, two profiles from the shore to the deepest point were defined. The number of measurement points differed between the lakes depending on their size. In total, 52 lakes were surveyed.

3.2. Lake volumes

The volumes of the lakes were determined in an ArcGIS environment using depth profiling measurements, contour lines and border lines. The water surface levels of the lakes were estimated using elevation levels presented in the basic map (National Land Survey of Finland, 2010a). Lake morphology was mostly interpolated using spline that results in a smooth surface passing by all the input points (ESRI, 2014). The Tension method with 0.1 weight and 3 input points was used to calculate the values for the interpolated cells. Interpolation rasters were extracted by surface water areas. The mean depths of these new rasters were multiplied by the water surface areas in order to calculate the volumes of the lakes.

3.3. Evaporation from lakes

Evaporation (E) was calculated individually for all the lakes surveyed using a mass transfer approach (Rosenberry et al., 2007; Dingman, 2008). This method was chosen because it yields instantaneous rates of evaporation and takes into account lake sizes (Harbeck, 1962). Our data enabled the calculations of daily mean values of evaporation flux for all the lakes studied. Averaging the parameters for periods of time longer than one day can lead to significant biases in the calculated *E* values since on these time scales the vapour pressure differences and wind speed may correlate (Jobson, 1972). The following expression was used to calculate the evaporation flux (Dingman, 2008):

$$E = K_{E}.v_{a} \cdot (e_{s} - e_{a}), \quad (1)$$

where:

E is evaporation rate (mm s^{-1})

K_E is the mass-transfer coefficient in m km^{-1} kPa^{-1} describing the impact of turbulent eddies of the wind on vertical transport of water vapour from the lake with area A_L (km^2), $K_E = 1.69 \cdot 10^{-5} \cdot A_L^{-0,05}$ (Harbeck, 1962)

v_a is wind speed (m s^{-1}) at 2 m height

e_s is the saturation vapour pressure in kPa at surface water temperature T_s (˚C), $e_s = 0.611 \cdot \exp(\frac{17.3 \cdot T_s}{T_s + 237.3})$

e_a is the vapour pressure in the air in kPa, $e_a = h \cdot 0.611 \cdot \exp(\frac{17.3 \cdot T_a}{T_a + 237.3})$, where h is relative humidity and T_a is air temperature (ºC).

Wind speed measured at 10 m height was adjusted to the corresponding speed at 2 m height using the power law profile (Justus and Mikhail, 1976):

$$v_z = v_r(\frac{z}{z_r})^{\beta}, \quad (2)$$

where v_r is the measured wind speed at the reference height z_r (10 m), z is the height for which speed is adjusted (2 m) and β is the friction coefficient. The value of 0.15 for β, characteristic for grassland, was employed in our study (0.1 characterizes oceans and lakes) (Bañuelos-Ruedas et al., 2010) since the lakes are surrounded by forests lowering the wind speed.

The meteorological parameters necessary for the calculations (relative humidity, wind speed and air temperature) were obtained from the meteorological station 5502 (Vaala-Pelso) of the Finnish Meteorological Institute (2014), located approximately 10 km from the site. A probable range for the lakes' surface water temperature was evaluated using continuous temperature measurements at 50 cm depth from one of the studied lakes, Ahveroinen (see section 4.1), and a standard deviation of this temperature determined from thermal images. Using the derived temperature range, a probable range for evaporation rates from all the lakes was calculated. The adopted method relies only on temperature difference measured on one day, but since all the lakes are in a relatively small area with almost identical weather conditions, it is highly probable that the seasonal behaviour of the surface water temperature of the studied lakes is similar.

3.4. Isotope mass balance

Instantaneous water and isotope balances for an evaporating surface water body can be formulated as follows:

$$\frac{dV}{dt} = T_{TOT} - E - O_{TOT} \quad (3)$$

$$\delta_L \frac{dV}{dt} + V\frac{d\delta_L}{dt} = \delta_{IT} \cdot I_{TOT} - \delta_E \cdot E - \delta_{OT} \cdot O_{TOT} \quad (4)$$

where V stands for the volume of the surface water body, δ_L signifies its isotopic composition and I_{TOT}, E and O_{TOT} represent the total inflow, evaporation and total outflow of water from the system, respectively, whereas δ_{IT}, δ_E and δ_{OT} stand for their respective isotopic compositions, expressed in ‰. As the total inflow may consist of several components (precipitation, underground and surface inflows), each with its specific isotopic composition, δ_{IT} should be calculated as a flux-weighted mean of the respective isotopic compositions of individual components. The total outflow may also consist

of surface and underground components. For well-mixed systems it is typically assumed that $\delta_{OT} = \delta_L$.

The isotopic composition of the evaporation flux, δ_E, cannot be measured directly. However, it can be calculated using the expression derived from the linear resistance model describing isotope effects accompanying evaporation process (Craig and Gordon, 1965; Horita et al., 2008):

$$\delta_E = \frac{(\delta_L/\alpha_{LV}) - h_N \cdot \delta_A - \varepsilon}{(1-h_N) + \Delta\varepsilon \cdot 10^{-3}} \quad (5)$$

where ε is the total effective isotope fractionation, $\varepsilon = \varepsilon^* + \Delta\varepsilon$, where ε^* stands for equilibrium isotope enrichment ($\varepsilon^* = (1-1/\alpha_{LV})\cdot 10^3$) expressed in ‰ and α_{LV} represents the equilibrium isotope fractionation factor between liquid and gaseous phases. The kinetic isotope enrichment Δε, is defined as $\Delta\varepsilon = C_k(1 - h_N)$ where C_k stands for the kinetic enrichment parameter. δ_A represents the isotopic composition of atmospheric water vapour over the evaporating water body (‰) and h_N is the relative humidity of the local atmosphere, normalized to the temperature of evaporating water.

When the evaporating water body is in hydrologic and isotopic steady-state ($dV/dt=0$ and $d\delta_L/dt=0$, respectively), the following approximate expression describing the isotope enrichment of the evaporating water body can be derived from eqs. 3-5 (see e.g. Gat and Bowser, 1991):

$$\Delta\delta = \delta_{LS} - \delta_{IT} \cong \frac{\delta_A - \delta_{IT} + \varepsilon/h_N}{1 + \frac{I_{TOT}}{E}\frac{1-h_N}{h_N}} \quad (6)$$

where Δδ stands for the evaporative enrichment and δ_{LS} is the steady-state isotopic composition of the studied system. The following expression describing the ratio of the total inflow to the evaporation rate can be derived from eq. (6):

$$\frac{I_{TOT}}{E} = \frac{\delta_A - \delta_{LS} + \varepsilon/h_N}{(\delta_{LS} - \delta_{IT})\frac{1-h_N}{h_N}} \quad (7)$$

If the evaporation flux E can be assessed independently, as is the case of this study, the total inflow of water to the given lake can be calculated from eq. (7). From this, the groundwater component of this inflow can be further inferred, provided that the other components of this inflow are known or can be independently assessed.

To quantify the water balance of a lake with the aid of eq. (7) the knowledge of the isotopic composition of atmospheric moisture interacting with the lake, δ_A, is required. Field measurements of this parameter become feasible only recently thanks to advancements of CRDS technology.

However, the combined studies of the isotopic composition of atmospheric water vapour and precipitation performed in moderate climates (Jacob and Sonntag, 1991; Schoch-Fischer et al., 1984) have shown that, on monthly basis, the isotopic composition of precipitation is generally in isotopic equilibrium with the local atmospheric moisture at ground-level temperature. This is particularly true for summer season. Therefore, the value of δ_A can be derived from the isotopic composition of local precipitation, δ_P, which is also required to quantify the isotopic composition of the total inflow to the studied lake, δ_{IT}, appearing in eq. (7). The following relation can be used to calculate δ_A (in ‰):

$$\delta_A = \frac{1}{\alpha_{LV}}(\delta_P + 10^3) - 10^3 \quad (8)$$

Equation (7) is valid under two basic assumptions: (i) the evaporating water body is in hydrologic and isotopic steady-state, and (ii) the water body is isotopically homogeneous. Natural surface water systems, such as lakes, typically operate close to their hydrologic and isotopic steady-states attained in the course of their long history. Their steady-state characteristics are defined by local climate, morphological setting and prevailing hydrological regime. Such systems usually exhibit seasonal fluctuations of varying amplitude, caused by seasonal fluctuations of local climate (surface air temperature, relative humidity, precipitation amount), superimposed on long-term trends. Seasonal ice-cover, typical for mid and high latitudes, may also contribute to the seasonal fluctuations of the steady-state characteristics of such systems. The gradual attainment of the isotopic steady-state, which is characterized by an exponential function describing the temporal evolution of δ_L, can be observed only for surface water systems which are artificially created, such as dredging lakes resulting from exploitation of gravel deposits (Zimmerman, 1979). The time constant characterizing the dynamics of this process is mainly controlled by the mean turnover time of water in this system, defined as the ratio of its volume to the total inflow, its hydrological balance (I_{TOT}/E ratio) and the normalized relative humidity (e.g. Gonfiantini, 1986; Zimmerman, 1979).

4. Results and discussion

4.1. Climate and lake water temperature data

The climate is strongly seasonal in the Rokua esker study area. The long-term monthly mean values of surface air temperature vary from −10.9 °C (January) to 13.2 °C (June) for the period between 1959 and 2013 (Fig. 2). The amount of monthly precipitation varies from 29 mm (February and April) to 79 mm (August) for the same period. The warmest months of the year are June, July and August. The

long-term (1970-2013) monthly mean relative humidity of air varies from 61 % (May) to 91 % (November).

The seasonal temperature patterns of the monitored lakes were very similar, despite significant differences in lake size (Fig. 3). Thus the surface water temperature of lake Ahveroinen 1 (mean value of 19.1 °C) during the period from 1 June 2013 to 31 August 2013 was used as a basis for estimating the water temperature of other lakes. Thermal images collected on 5 August 2013 yielded comparable surface water temperatures that ranged from 19.5 °C to 24.6 °C, with a mean of 21.3 °C and a standard deviation of 0.87 °C. Combining the results of continuous temperature measurements and thermal images, an estimate of the mean surface water temperature of all lakes for the period from 1 June to 31 August 2013 was derived to be 19.10 °C ± 0.87 °C. This temperature was used in the isotope mass balance calculations.

4.2. Local isotopic compositions

An overview of the isotopic composition of different types of water in the study area is presented on the δ^2H-$\delta^{18}O$ space in Fig. 4. It comprises precipitation data collected during the period from 18 March 2010 till 29 October 2013 at the station located on the esker, the mean isotopic compositions of the selected lakes, streams and groundwater monitored during the period 2010-2013 (Table 1), as well as the isotopic compositions of 67 lakes surveyed in July and August 2013 (Table 2).

The local meteoric water line (LMWL) of Rokua ($\delta^2H = 7.77 \cdot \delta^{18}O + 9.55$) was defined using the δ values of precipitation samples collected during the years 2010-2013 (Fig. 4). The local evaporation line (LEL), $\delta^2H = 5.09 \cdot \delta^{18}O - 28.19$, is the best fit line of the δ values representing lake water data for the year 2013. The intercept of the LMWL and LEL lines yields the estimate of the weighted annual mean $\delta^{18}O$ and δ^2H values of precipitation (−14.1 ‰ and −100 ‰ respectively). Slightly elevated mean $\delta^{18}O$ and δ^2H values of local groundwater (−13.1 ‰ and −95 ‰), combined with reduced deuterium excess ($d = \delta^2H - 8 \cdot \delta^{18}O$ = 4.8 ‰) when compared to deuterium excess of precipitation (12.8 ‰), indicate the presence of an evaporation signal in the local groundwater. Furthermore, some of the winter precipitation is most probably returned to the atmosphere via sublimation and does not contribute to groundwater recharge.

Although majority of groundwater samples cluster near the LMWL-LEL intersect, there are some data points lying along the LEL in the $\delta^{18}O$-δ^2H plot indicating the contribution of (evaporated) lake water to groundwater. However, the interconnection between the lakes via groundwater is likely to

be minor since lakes probably have groundwater table maxima between them as they are situated in deep holes in the landscape. Nevertheless, the existence of deeper flow paths from the upper elevation lakes to the lower ones cannot be excluded. Based on the estimation of elevation differences and distances between the lakes, we identified the lakes which do not have surface inflows but which might receive some groundwater input originating from upper elevation lakes. These are lakes No. 6, 14, 20, 34, 38, 41, 47, 50 and 59. The sensitivity study presented in section 4.5 considers probable changes of the isotopic composition of the total inflow to each lake caused by the presence of evaporated lake water component.

The influence of evaporated lake water seeping to groundwater is illustrated in Fig. 5, showing the isotopic composition of lake Ahveroinen 1 and adjacent groundwater. The mean $\delta^{18}O$ and $\delta^{2}H$ values of piezometers MEA 2010 and MEA 1907 situated on the south-eastern and north-western sides of the lake were −13.4 ‰ and −97 ‰ and −10.7 ‰ and −83 ‰ respectively, clearly indicating a substantial (ca. 55 %) contribution of lake water to groundwater at the north-western side of the lake. A smaller contribution (ca. 10 %) of lake water to groundwater can be seen on the eastern and western sides of the lake where the mean $\delta^{18}O$ and $\delta^{2}H$ values of groundwater were −12.9 ‰ and −93 ‰ respectively. The main direction of groundwater flow is therefore from south-east to north-west, which coincides with the results from seepage measurements conducted by Ala-aho et al. (2013). The direction of groundwater flow can also be noted from the difference in the mean isotopic composition of the lake water between points 2 and 3 of −8.7 ‰ and −73 ‰ and −8.5 ‰ and −72 ‰ respectively. The difference in isotopic compositions between points 2 and 3 was greatest during the winter: in March 2011 the difference in $\delta^{18}O$ and $\delta^{2}H$ between these points was −1.1 ‰ and −4 ‰ respectively.

4.3. Temporal variations in the isotopic composition of lake water

The seasonal variability in the isotopic composition of the lakes studied is illustrated in Fig. 6, showing changes of $\delta^{18}O$ in lake Ahveroinen 1 at two depths (1 and 4 meters). $\delta^{18}O$ of lake Ahveroinen 1 reveals distinct seasonal fluctuations with peak-to-peak amplitude in the order of 1 ‰. This lake does not have any surface inflows or outflows. After the disappearance of ice cover during the spring (April-May), the lake starts to evaporate, which results in its gradual enrichment in heavy isotopes, approaching the steady-state value sometime in September-October. Freezing of the lake in late autumn stops the evaporation flux. Systematic decline of $\delta^{18}O$ during ice-cover period seen in

Fig. 6 stems from gradual dilution of lake water with groundwater seeping into the lake. Figure 6 shows that the lake is well mixed throughout the year.

The declining parts of the $\delta^{18}O$ curve in Fig. 6 can be used to assess the intensity of groundwater inflow during ice-cover period, if the volume of the studied lake is known and the isotopic composition of groundwater is constant. The isotope balance of such lake system (eq. 4) can be then expressed as follows:

$$V \frac{d\delta_L}{dt} = \delta_{IT} \cdot I_{TOT} - \delta_{OT} \cdot O_{TOT} \quad (9)$$

Since $I_{TOT} = I_{GW}$, $\delta_{IT} = \delta_{GW}$, $O_{TOT} = O_{GW} = I_{GW}$ and $\delta_{OT} = \delta_L$, eq. (9) becomes:

$$\frac{d\delta_L}{dt} = \frac{I_{GW}}{V}(\delta_{GW} - \delta_{L)} \quad (10)$$

The solution of this differential equation reads as follows:

$$\delta_L = (\delta_{Lo} - \delta_{GW}) \cdot e^{-k \cdot t} + \delta_{GW} \quad (11)$$

where $k = I_{GW}/V$ and δ_{Lo} is the isotopic composition of the lake at the beginning of ice-cover period ($\delta^{18}O_{Lo} = -8.5$ ‰ for lake Ahveroinen 1). First derivative of eq. (11) at t = 0 is:

$$\left(\frac{d\delta_L}{dt}\right)_{t=0} = \frac{I_{GW}}{V}(\delta_{GW} - \delta_{Lo}) \quad (12)$$

Equation (12) allows to calculate the mean flux of groundwater to the lake during ice-cover period. The observed reduction of $\delta^{18}O$ of lake Ahveroinen 1 by ca. 1.2 ‰ over the six-month period (cf. Fig. 6, Table 1) results from the continuous inflow of groundwater with specific isotopic signature ($\delta^{18}O_{GW} = -13.4$ ‰). The groundwater seepage rate obtained from eq. (12) is ca. 160 m^3 day^{-1}. Identical value was obtained using corresponding ^2H data. It is worth noting that groundwater inflow to lake Ahveroinen 1, derived for the summer period of 2013 from isotope mass balance calculations (ca. 300 m^3 day^{-1}) is almost two times higher, which suggests significant seasonal variations of groundwater inflow to Rokua lakes, with high groundwater fluxes during summer and low fluxes during winter.

4.4. Quantifying groundwater dependence of the studied lakes

Out of 67 Rokua lakes sampled during July-August 2013 field campaign, 50 lakes do not reveal any surface water inflow in the form of a stream or a creek. For the remaining 17 lakes surface inflows were identified. For all but one lake those surface inflows could be linked to specific upstream lakes which were sampled during the July-August 2013 campaign.

The isotopic composition of the lakes sampled covers a wide range of δ values: from −5.6 ‰ to −12.7 ‰ and from −57 ‰ to −93 ‰ for $\delta^{18}O$ and $\delta^{2}H$ respectively. This large variability reflects a wide spectrum of the heavy isotope enrichment of the lakes studied. Since Rokua lakes are situated in a unique climatic region, their observed isotopic composition is primarily controlled by their water balance, which in turn can be characterised by the total inflow-to-evaporation ratio.

Equation (7) was used to calculate the total inflow-to-evaporation ratios (I_{TOT}/E) for all studied lakes. The isotope mass balance calculations were run for the period of June-August 2013 separately for ^{18}O and ^{2}H. Since the evaporation rates from the studied lakes were derived from eq. (1), independently of any mass balance considerations, the total inflow to each lake could also be calculated from the assessed I_{TOT}/E ratios.

The values of the parameters occurring in eq. (7) were derived as follows:

(i) The mean isotopic composition of atmospheric moisture (δ_A) was calculated with the eq. (8) using the available isotope data for local precipitation. The following mean δ_A values were used in isotope mass balance calculations: $\delta^{18}O_A$ = –20.4 ‰ and $\delta^{2}H_A$ = –149 ‰.

(ii) The mean relative humidity, normalized to lake water temperature (h_N) was calculated using the mean surface air temperature (+14.9 ^{o}C), the mean surface water temperature of the lakes (+19.1 ^{o}C) and the mean relative air-based humidity calculated on the basis of daily mean values available from meteorological station in the area (79.2 %). The resulting mean h_N value was 60.7 %.

(iii) The total effective isotope fractionation (ε) was derived as the sum of the equilibrium and kinetic isotope enrichments ($\varepsilon = \varepsilon^{*}+\Delta\varepsilon$). The equilibrium isotope enrichment (ε^{*}) values for ^{18}O and ^{2}H, were calculated for the mean surface water temperature of +19.1 ^{o}C using the known temperature dependence of the empirical equilibrium fractionation factors, α_{LV} (Horita and Wesolowski, 1994). The values of kinetic enrichment parameters (C_k) used to calculate $\Delta\varepsilon$, 14.2 ‰ for ^{18}O and 12.5 ‰ for ^{2}H, were adopted after Gonfiantini (1986). Those values were obtained in wind-tunnel experiments (Vogt, 1976) and are widely used in lake studies.

(iv) Isotopic homogeneity of the studied lakes was addressed through multiple samplings of large lakes (in both horizontal and vertical direction - cf. section 3.1). The range of the measured $\delta^{18}O$

values for the given lake was generally lower than one per mil. It was assumed that the average values calculated on the basis of individual measurements performed in each lake represent sufficiently well the studied systems.

(v) The isotopic composition of lake water obtained during the sampling campaign in July and August 2013 for each studied lake was used in the isotope mass balance calculations. As discussed in section 3.4 above, the isotopic compositions of the studied lakes fluctuate seasonally reaching their steady-state values toward the end of ice-free period (September-October). Therefore, the isotopic compositions of lake water samples collected during the late summer (July-August) may still deviate slightly from the respective steady-state values, thus creating uncertainty in the assessed components of the water (see section 4.5).

(vi) As the isotopic compositions of the surface and underground components of water inflow to each studied lake were not measured directly, an iterative approach was adopted to calculate δ_{IT} individually for each lake. In the first step it was assumed that δ_{IT} is defined by the intercept of the LMWL and LEL lines (Fig. 4) which represents the weighted annual mean $\delta^{18}O$ and $\delta^{2}H$ values of local precipitation (–14.1 ‰ and –100 ‰ respectively). With these δ_{IT} values the underground component of I_{TOT} for each lake was derived from eq. (7). The second step differed for the lakes without surface inflow and with identified surface inflow from an upstream lake. For the lakes without surface inflow the δ_{IT} values were calculated individually as flux-weighted averages of the underground inflow obtained in the first/previous step of the procedure and precipitation input to the given lake, each with their respective mean isotopic compositions representing the period of June-August 2013 (cf. section 4.2 above). For the lakes with identified surface inflow from an upstream lake, in the first instance the total outflow of water from the upstream lake was calculated using appropriate mass balance equation. Then, it was assumed that 25 % of the total outflow from the upstream lake flows to the downstream lake as surface inflow carrying the isotopic composition characteristic for the upstream lake. The δ_{IT} values were calculated individually for each lake belonging to this group of lakes as flux-weighted averages of three components: precipitation, surface inflow and groundwater inflow. For both lake groups the calculations were repeated until the change of δ_{IT} in subsequent iteration step was in the order of the analytical uncertainty of isotope measurements (0.1 ‰ for $\delta^{18}O$ and 1 ‰ for $\delta^{2}H$).

The calculated inflow-to-evaporation ratios of the studied lakes based on ^{18}O isotope mass balance are shown in Fig. 7 as a function of isotope enrichment of lake water with respect to the isotopic composition of the total inflow to the given lake. They vary in a wide range, from I_{TOT}/E values

between 2 and 3 and large ^{18}O isotope enrichments between approximately 6.5 and 8.0 ‰, indicating evaporation dominated systems, to typical through-flow lakes characterized by I_{TOT}/E ratios higher than 10 and moderate ^{18}O isotope enrichments of less than 2 ‰ (Table 2). Knowing the volume and the total inflow, the mean turnover time of water (MTT) in each lake could be quantified as the ratio of lake volume to the total inflow. The calculated MTT values range from approximately one week for the Pasko pond (V = 2×10^3 m^3, mean depth 0.2 m, maximum depth 0.7 m) to approximately five years for lake Saarijärvi 2 (V = 2.47×10^6 m^3, mean depth 11.8 m, maximum depth 26 m), with the mean in the order of ten months (Table 2). Lake Saarijärvi 2 is the deepest of all the lakes surveyed. As expected, the calculated MTT values correlate well with the mean depth of the studied lakes, expressed as the volume-to-surface area ratio. However, the link between MTT and the I_{TOT}/E ratio is much weaker; lakes with higher I_{TOT}/E ratios tend to have shorter mean turnover times.

The dependence of the studied lakes on groundwater can be quantified through an index (*G* index) defined as the percentage contribution of groundwater inflow to the total inflow of water to the given lake. Groundwater inflow was derived by subtracting the precipitation and surface water inflow (if exists) from the total inflow. Such an evaluation was undertaken for all the lakes listed in Table 2. Note that for the group of lakes with identified surface inflow from an upstream lake it was assumed arbitrarily that this surface inflow is 25 % of the total outflow from the upstream lake (discharges of surface inflows were not measured). The resulting groundwater seepage rates vary from less than 20 m^3day^{-1} for Kissalampi pond to around $14x10^3$ m^3day^{-1} for lake Nimisjärvi, the lake with the largest surface area (167.5 ha) among all studied lakes.

Figure 8 summarises the values of *G* index obtained for the lakes surveyed during the July-August 2013 sampling campaign. The mean values of *G* index obtained from ^2H and ^{18}O balance are shown. They vary from ca. 40 % to more than 95 %. The lowest value (39.4 %) was obtained for lake Etu-Salminen. The highest *G* values were derived for lakes Kiiskeroinen (97.1 %) and Levä-Soppinen (97.5 %). Interestingly, these lakes are characterised by a high degree of eutrophication induced by high loads of phosphorus brought to the lakes with groundwater (Ala-aho et al. 2013). Although the *G* index describes groundwater dependency of the studied lakes rather unambiguously, also lakes with moderate *G* index values can suffer if groundwater table in the esker aquifer would decline as a result of climate and/or land-use changes.

The isotope mass balance calculations for Rokua lakes were run independently for ^{18}O and ^2H data. Consistent results were obtained with respect to three evaluated elements of lake water balance

(I_{TOT}/E ratios, MTT values and the G index) reported in Table 2. These quantities, derived independently from ^{2}H-based and ^{18}O-based isotope mass balances are highly correlated (R^2 = 0.9681, 0.9972 and 0.9745 for I_{TOT}/E ratios, MTT values and the G index, respectively). The total inflow-to-evaporation ratios derived from ^{18}O-based balance turned out to be ca. 10.8 % higher on the average than those derived from ^{2}H-based balance. For the G index this difference is approximately 2.3 %. The MTT values were ca. 12.5 % higher for ^{2}H-based balance.

Small but significant differences in the values of the evaluated quantities (I_{TOT}/E ratios, MTT, G index), derived independently from ^{18}O- and ^{2}H-based isotope mass balance calculations, stem most probably from their different sensitivity to small changes of the measurable parameters (air and lake water temperature, relative humidity, isotopic composition of local precipitation, isotopic composition of lake water) rooted in different role of equilibrium and kinetic fractionation during the evaporation process. While for ^{18}O the ratio of equilibrium to kinetic isotope fractionation is in the order of one, for ^{2}H it is ten times higher. Since isotope mass balance method relies on isotope enrichment of lake water along the local evaporation line, controlled mostly by kinetic fractionation, the ^{18}O-based balance calculations are generally considered more reliable (e.g. Rozanski et al., 2001).

4.5. Uncertainty assessment

The above methodology for quantifying elements of water balance in the lakes studied introduces some uncertainties linked to the assumptions made and the uncertainties associated with the parameters used in the evaluation process. Sensitivity tests were performed to derive the range of uncertainties associated with the quantities being evaluated, such as mean turnover time, total inflow-to-evaporation ratio and the G index. The sensitivity analysis was focussing on eq. (7). All variables present in this equation were considered in this process. The results for ^{18}O-based calculations are summarized in Table 3.

The uncertainty with regard to lake water temperature was probed assuming the temperature change of ± 0.87 ºC (cf. section 4.1). The uncertainty of lake water temperature leads to uncertainty as regards the evaporation flux, which in turn influences the I_{TOT}/E, MTT and G values derived for each lake. Also equilibrium isotope enrichment is a function of temperature. The mean turnover time increases by ca. 15 % when the temperature of the lake is reduced by 0.87 ºC, and decreases by approximately 12 % when the temperature increases by the same amount. The G index reveals lower sensitivity (2.9

and 3.6 %, respectively). The smallest changes were obtained for I_{TOT}/E ratios (0.9 and 0.7 %, respectively).

The changes of relative humidity of the atmosphere normalised to the temperature of the lake surface have an impact on the evaporation flux, control the I_{TOT}/E ratios through eq. (7) and determine the actual value of kinetic isotope enrichment $\Delta\varepsilon$. It was assumed in the calculations that normalized relative humidity changes by ± 2 %. As seen in Table 3, the resulting changes of the derived quantities are moderate, the mean turnover time being the most sensitive parameter.

It is apparent from Table 3 that among isotope parameters occurring in eq. (7), the isotopic composition of lake water (δ_{LS}) and the isotopic composition of the total inflow (δ_{IT}), are the two most important variables in the isotope mass balance calculations. An increase of $\delta^{18}O_{LS}$ by 0.5 ‰, which may account for possible departures from the isotopic steady-state of the investigated lakes (cf. section 4.4), leads to decrease of calculated I_{TOT}/E ratios on the average by 15.8 %, increase of MTT values by 19.6 % and decrease of G index values by 3.3 %. An increase of $\delta^{18}O_{IT}$ by 0.5 ‰, which may result from the contribution of an evaporated lake water originating from an upstream lake to groundwater input, leads to substantial increase of I_{TOT}/E ratios (20.2 % on the average), comparable decrease of MTT values (14.8 % on the average) and moderate increase of the G index (4.4 % on the average). Variation of the ^{18}O isotopic composition of atmospheric water vapour by ± 1.0 ‰ introduces changes in the calculated elements of the water balance of the studied lakes in the order of several per cent (Table 3).

Figure 9 shows the percentage changes of I_{TOT}/E ratios calculated for all studied lakes using eq. (7), in response to the increase of δ_{LS} or δ_{IT} by 0.5 ‰. The sensitivity of the calculated I_{TOT}/E ratios to the given increase of δ_{LS} or δ_{IT} raises sharply with increasing value of this parameter. This is particularly true for through-flow systems characterized by high I_{TOT}/E ratios. Therefore, when isotope studies aimed at quantifying water balance of such systems are planned, it is important to characterize these two isotope quantities as good as reasonably possible.

5. Conclusions

The Rokua esker, with its numerous lakes located across a relatively small area, provided a unique opportunity to explore the possibilities offered by environmental isotope techniques in quantifying the water balances of lakes and their dependency on groundwater in a sub-polar climatic setting. The

807 quantification of groundwater seepages to lakes using conventional methods is notoriously difficult
808 and associated with considerable uncertainty. The presented study demonstrates the power of isotope
809 mass balance approach for resolving this issue. It appears that a stable isotope analysis of lake water
810 samples, collected at right time and supplemented by appropriate field observations, may lead to
811 quantitative assessment of the water balance of a large number of lakes located in a similar climatic
812 settings.

813 The presented study has demonstrated that consistent results are obtained when the isotope mass
814 balance calculations are run independently for oxygen-18 and deuterium. This strengthens the
815 position of heavy stable isotopes of water as a unique tool for quantifying elements of water balance
816 of lakes, particularly for groundwater-dominated systems. Solving three equations simultaneously
817 (one water balance equation plus two isotope balance equations) may help to quantify key balance-
818 related parameters such as evaporation and groundwater inflow and outflow rates for the studied lake
819 system, which are difficult to quantify using conventional methods.

820 The specific behaviour of lakes located in sub-polar regions, with their seasonal ice cover extending
821 over several months, offers another opportunity for quantifying groundwater seepage during ice-
822 cover periods. As shown in this study, observations of seasonal changes in the stable isotopic
823 composition of lake water, in particular during the ice-cover period, combined with the survey of
824 isotopic composition of groundwater in the vicinity of the lakes studied, allows the quantification of
825 groundwater fluxes to this lake during winter. If such an approach is combined with the isotope mass
826 balance calculations performed for ice-free summer season, important information about the seasonal
827 variability of groundwater seepage to lakes located in sub-polar and polar regions can be obtained.

828 The *G* index characterizing groundwater dependency of a lake proposed in this study, and defined as
829 a percentage contribution of groundwater inflow to the total inflow of water to the given lake, appears
830 to be a straightforward, quantitative measure of this dependency. The studied Rokua lakes appear to
831 be strongly dependent on groundwater; more than 40 % of water received by these lakes comes as
832 groundwater inflow. The quantitative evaluation of groundwater dependency of lakes via the *G* index
833 proposed in this study may assist lake restoration policies in areas where groundwater is a source of
834 nutrients to the studied lakes.

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997 Table 1. Mean isotopic composition of selected lakes, streams and groundwater, sampled four times
998 per year during the period 2010-2012.

	δ^2H (‰)	$\delta^{18}O$ (‰)	d-excess (‰)	Mean amplitude of $\delta^{18}O$ seasonal signal (‰)
Lakes:				
1. Ahveroinen 1	−72.4	−8.50	−4.42	1.2
2. Rokuanjärvi	−72.1	−8.75	−2.14	1.3
3. Jaakonjärvi	−64.1	−6.72	−10.34	0.8
4. Kolmonen 2	−66.7	−7.07	−10.14	0.9
5. Loukkojärvi	−63.0	−6.64	−9.88	0.9
6. Saarijärvi 2	−62.4	−6.81	−7.92	0.9
7. Saarinen	−72.3	−8.52	−4.14	3.3
8. Salminen	−71.4	−8.45	−3.80	0.5
9. Soppinen	−63.1	−6.84	−8.38	0.9
10. Tulijärvi	−85.6	−11.20	4.00	1.3
11. Vaulujärvi	−67.9	−7.68	−6.46	0.5
Streams:				
1. Heinäjoki	−93.0	−12.74	8.92	0.5
2. Hieto-oja	−88.9	−11.84	5.82	1.2
3. Kangasoja	−93.7	−12.90	9.50	0.7
4. Lianoja	−83.2	−10.67	2.16	2.2
5. Lohioja	−92.9	−12.93	10.54	0.8
6. Matokanava	−94.6	−13.06	9.88	0.9
7. Päiväkanava	−94.0	−13.01	10.08	0.9
8. Rokuanoja	−82.8	−10.95	4.80	3.1
9. Siirasoja	−94.6	−13.08	10.04	0.7
10. Soppisenoja	−89.9	−12.07	6.66	2.2
11. Valkiaisoja	−93.3	−12.80	9.10	1.0
Groundwater:				
1. MEA 106	−96.4	−13.32	10.16	0.6
2. MEA 206	−95.0	−13.12	9.96	0.8
3. MEA506	−93.5	−12.92	9.86	0.5
4. MEA 706	−93.3	−12.89	9.82	0.7
5. MEA 1106	−95.1	−13.05	9.30	0.8
6. MEA 1807	−93.3	−12.92	10.06	1.0
7. MEA 1907	−83.2	−10.65	2.00	1.0
8. MEA 2010	−97.3	−13.42	10.06	0.5
9. ROK 1	−95.5	−13.22	10.26	0.9
10. Siirasoja 1 esker	−94.5	−13.06	9.98	0.8
11. Siirasoja 1 slope	−94.3	−13.08	10.34	1.0
12. Siirasoja 1 sand	−97.0	−13.36	9.88	0.5
13. Siirasoja 1 peat	−93.8	−12.98	10.04	1.2

999

1000 Table 2. The results of isotope mass balance calculations for 67 lakes of Rokua esker sampled during the July-August 2013 field survey.

Lake No.	Lake name	Volume (10^3 m^3)	Surface area (ha)	Mean depth (m)	*E* [(a)] (mm)	$\delta^{18}O$ (‰)	$\delta^{2}H$ (‰)	I_{GW} (m^3/day)		I_{TOT}/E		MTT (month)		*G* index (%)	
								^{18}O	^{2}H	^{18}O	^{2}H	^{18}O	^{2}H	^{18}O	^{2}H
1	Heinälampi [(b)]	1764	0.22	0.8	286	–10.97	–84.4	71.5	67.3	14.2	13.3	0.6	0.6	72.4	73.0
2	Holma [(b)]	23778	1.89	1.3	257	–11.12	–84.8	569.3	520.7	15.8	14.2	0.9	1.0	68.4	69.5
3	Koivujärvi 1 [(b)]	39237	2.3	1.7	254	–11.38	–85.7	823.9	681.4	19.0	16.2	1.1	1.2	68.2	66.3
4	Koivujärvi 2 [(b)]	76884	2.93	2.6	251	–11.61	–87.3	1555.1	1393.4	20.6	18.5	1.5	1.7	94.1	93.9
5	Pitkäjärvi 1	36103	1.22	3	262	–6.50	–61.5	61.8	51.7	2.6	2.3	13.3	15.0	69.8	66.0
6	Nurkkajärvi	269134	3.93	6.8	247	–8.76	–73.1	407.7	362.8	4.7	4.2	17.8	19.5	82.5	80.8
7	Luontolampi	19995	0.49	4.1	275	–7.29	–64.9	34.6	28.6	3.1	2.7	14.4	16.6	76.4	72.8
8	Ahveroinen 1	119000	3.32	3.6	249	–8.80	–71.9	351.3	282.7	4.7	3.9	9.1	10.9	82.8	79.5
9	Lianjärvi [(b)]	113678	15.13	0.8	231	–10.31	–80.5	2375.0	2044.1	10.0	8.9	1.0	1.1	62.3	60.3
10	Syväjärvi 1 [(b)]	398463	11.52	3.5	234	–11.17	–85.3	3288.8	3030.3	12.7	11.7	3.5	3.8	88.3	88.1
11	Soppinen	81630	6.04	1.4	242	–7.60	–65.7	396.6	310.2	3.3	2.8	5.0	6.0	74.9	70.0
12	Salminen	681860	25.31	2.7	225	–8.59	–70.7	2177.5	1724.8	4.4	3.7	8.1	9.8	79.7	75.6
13	Saarinen	981506	15.32	6.4	231	–8.22	–69.4	1186.0	974.2	4.0	3.4	21.0	24.4	77.9	74.3
14	Kivi-Ahveroinen	216018	5.57	3.9	243	–9.02	–74.4	623.6	556.9	5.1	4.6	9.4	10.4	83.6	82.0
15	Irvi-Ahveroinen	40867	1.07	3.8	264	–7.47	–65.6	76.2	62.1	3.3	2.8	13.4	15.6	76.5	72.6
16	Loukkojärvi	126450	2.85	4.4	251	–7.10	–65.6	166.8	153.3	2.9	2.8	18.0	19.1	72.7	71.0
17	Ylimmäinen	94313	8.93	1.1	237	–9.66	–77.6	1252.1	1130.9	6.3	5.8	2.1	2.3	86.5	85.2
18	Hietajärvi	223118	7.3	3.1	240	–7.06	–63.7	394.5	318.7	2.9	2.5	13.1	15.2	71.1	66.5
19	Saarijärvi 2	2467169	20.94	11.8	228	–6.72	–62.2	924.6	747.2	2.7	2.3	58.1	66.7	66.8	61.9
20	Syväjärvi 2 [(b)]	1863446	31.95	5.8	223	–10.63	–83.1	7668.9	7328.2	12.0	11.3	6.5	6.9	82.3	83.5
21	Pasko [(b)]	2005	0.82	0.2	268	–10.62	–83.3	189.6	185.8	10.8	10.5	0.3	0.3	73.8	74.4
22	Kuikkalampi	11776	0.61	1.9	271	–7.20	–66.6	41.5	39.7	3.0	2.9	7.0	7.2	75.5	74.7
23	Soppisenlampi	25645	0.59	4.4	272	–7.32	–67.1	41.4	39.4	3.1	3.0	15.4	16.0	76.3	75.4
24	Kirvesjärvi	654618	13.47	4.9	233	–11.45	–87.6	5195.9	5601.3	18.5	19.5	3.4	3.2	82.4	84.1
25	Tulijärvi [(b)]	589617	24.81	2.4	226	–10.40	–81.7	4650.7	4428.9	10.9	10.3	2.9	3.1	69.9	70.8
26	Jaakonjärvi 2	5497	0.46	1.2	275	–7.39	–65.3	34.1	27.9	3.2	2.7	4.1	4.7	77.1	73.3
27	Jaakonjärvi 3	13594	0.66	2.1	271	–7.60	–66.4	50.7	41.8	3.4	2.9	6.8	7.9	77.9	74.3
28	Maitolampi 2	44812	2.04	2.2	256	–7.52	–65.9	142.0	115.8	3.3	2.8	7.8	9.1	76.1	72.2

29	Kotalampi	25980	2.6	1	253	–11.23	–86.3	860.9	877.7	12.9	13.1	0.9	0.9	93.8	93.9
30	Rokuanjärvi 1	4364781	164.59	2.7	205	–9.11	–71.8	15206.8	10695.0	5.1	3.9	7.6	9.9	80.8	74.7
31	Tervatienlampi	21970	0.63	3.5	271	–7.92	–69.3	54.3	49.5	3.7	3.4	10.5	11.3	79.8	78.2
32	Valkiaislampi	32073	0.7	4.6	270	–8.81	–72.5	82.5	69.6	4.8	4.1	10.7	12.3	84.3	81.9
33	Ankkalampi	103700	3.86	2.7	248	–8.57	–71.9	373.5	326.9	4.4	4.0	7.4	8.2	81.5	79.4
34	Saarilampi 1	12044	0.45	2.7	276	–10.51	–80.8	110.5	91.2	8.9	7.5	3.3	3.9	91.8	90.2
35	Kiiskeroinen	6517	0.63	1	271	–12.14	–90.2	482.7	447.8	26.6	24.8	0.4	0.5	97.2	97.0
36	Jaakonjärvi 1	93614	3.53	2.7	249	–6.82	–63.5	184.2	159.3	2.7	2.5	11.7	12.9	70.4	67.3
37	Vaulujärvi	432813	8.97	4.8	237	–7.84	–68.6	628.9	553.8	3.6	3.2	17.1	18.8	76.2	73.8
38	Levä-Soppinen	34000	2.31	1.5	254	–12.27	–91.4	1922.5	2060.2	31.0	33.1	0.6	0.5	97.4	97.6
39	Anttilanjärvi	60230	1.06	5.7	264	–7.40	–66.4	73.8	64.9	3.2	2.9	20.3	22.3	76.1	73.7
40	Hautajärvi 1	189222	2.56	7.4	253	–7.71	–68.0	187.0	165.8	3.5	3.2	25.4	27.8	76.9	74.7
41	Lepikonjärvi	191118	2.98	6.4	251	–7.77	–68.3	220.2	195.8	3.5	3.2	21.8	23.9	77.1	75.0
42	Kolmonen 1	35700	0.68	5.3	270	–7.09	–65.2	43.9	39.5	3.0	2.7	19.8	21.4	74.7	72.7
43	Kolmonen 2	21890	0.56	3.9	273	–7.25	–65.8	38.5	34.1	3.1	2.8	14.1	15.4	75.9	73.6
44	Kolmonen 3	21630	0.54	4	273	–7.26	–66.1	37.4	33.9	3.1	2.9	14.3	15.4	76.0	74.2
45	Hätäjärvi	35217	1.73	2	258	–6.50	–61.8	85.4	72.8	2.5	2.3	9.3	10.4	69.2	65.7
46	Kissalampi	800	0.36	0.2	279	–5.60	–57.1	14.5	11.8	2.1	1.8	1.2	1.3	64.9	60.2
47	Valkiajärvi	582662	8.36	7	238	–8.49	–71.2	747.9	637.5	4.3	3.8	20.4	23.1	80.3	77.6
48	Hautajärvi 2	445413	14.6	3.1	232	–7.43	–64.5	857.1	646.0	3.2	2.6	12.3	15.0	72.8	66.8
49	Keskimmäinen [b]	428515	13.07	3.3	233	–10.76	–82.8	3117.4	2686.0	10.9	9.6	3.9	4.4	86.4	84.8
50	Siirasjärvi 2	23126	0.56	4.1	273	–12.66	–93.5	977.5[d]	1315.5[d]	59.7[d]	80.1[d]	0.8[d]	0.6[d]	98.8[d]	99.1[d]
51	Siirasjärvi 1	9701	0.34	2.9	280	–10.43	–80.4	81.5	66.8	8.6	7.2	3.6	4.3	91.6	90.0
52	Telkkälampi	7284	0.25	2.9	284	–7.30	–66.3	18.7	16.9	3.1	2.9	9.8	10.6	77.2	75.4
53	Maitolampi 1	48892	1.56	3.1	259	–6.36	–61.2	73.8	63.3	2.5	2.2	14.8	16.4	68.4	64.9
54	Taka-Salminen [b]	364757	7.31	5	240	–9.63	–76.2	772.6	609.0	7.8	6.4	8.0	9.7	52.3	49.6
55	Etu-Salminen [b]	196453	5.83	3.4	243	–8.05	–68.5	311.7	259.3	5.1	4.3	8.2	9.6	39.9	38.9
56	Pikku-Salminen	39856	1.32	3	261	–7.61	–65.8	97.4	76.9	3.4	2.8	10.3	12.3	77.1	72.7
57	Kylmäjärvi	394443	7.89	5	239	–8.47	–70.3	703.3	564.6	4.3	3.6	14.7	17.4	80.2	76.5
58	Kourujärvi 1	68023	1.83	3.7	257	–8.55	–71.7	183.8	159.9	4.4	3.9	9.9	11.1	82.1	80.0
59	Kourujärvi 2	31136	1.91	1.6	256	–7.64	–67.3	139.5	120.6	3.4	3.1	5.6	6.2	76.9	74.2
60	Huttunen	54588	1.49	3.7	260	–7.50	–66.9	105.6	92.7	3.3	3.0	12.9	14.2	76.3	73.9

61	Saarijärvi 1	13590	1.36	1	261	–6.49	–61.4	68.1	56.8	2.5	2.2	4.5	5.1	69.5	65.5
62	Pitkäjärvi 2	451093	7.93	5.7	239	–8.05	–68.7	607.3	505.3	3.8	3.3	18.8	21.6	77.7	74.4
63	Pyöräinen [(b)]	76077	3.76	2	248	–7.91	–67.9	270.8	224.5	4.8	4.1	5.1	6.0	56.0	54.6
64	Likainen [(b)]	33763	8.28	0.4	238	–10.94	–83.4	2110.9	1732.2	13.2	11.2	0.4	0.5	74.8	71.9
65	Nimisjärvi [(b), (c)]	1840396	167.53	1.1	205	–8.66	–69.7	15759.0	11551.0	5.2	4.1	3.1	3.9	81.1	75.8
66	Ahveroinen 2 [(b)]	670515	16.3	4.1	230	–7.96	–67.8	1077.3	854.5	4.3	3.6	12.4	14.7	61.2	57.6
67	Tervalampi	29707	0.79	3.7	268	–7.99	–68.7	69.3	58.8	3.7	3.3	11.2	12.7	79.9	77.1

(a) - Calculated for the period June 1 - August 31, 2013.
(b) - Lakes with identified surface inflow from an upstream lake.
(c) - Since the upstream lake for lake Nimisjärvi was not sampled, the mean isotopic composition of total inflows to lakes with identified surface water inflows was used for isotope mass balance calculations of this particular lake.
(d) - As the isotopic composition of a lake becomes comparable with the isotopic composition of the total inflow, the isotope mass balance calculations become very uncertain. Therefore, for lake Siirasjärvi 2 ($\Delta\delta^{18}O$ = 0.45‰ and $\Delta\delta^{2}H$ = 1.6‰) the values of I_{TOT}/E, MTT and G index reported in the table have only an indicative character.

Table 3. Sensitivity of selected elements of ^{18}O-based water balance of the studied lakes to changes of the parameters involved.

Parameter	Parameter change	Mean change of the selected elements of water balance (%) [(a)]: I_{TOT}/E	MTT	G index
1. Lake water temperature (T_w = 19.1 ^oC)	+0.87 ^oC	–0.9	–11.7	+2.9
	–0.87 ^oC	+0.7	+15.0	–3.6
2. Normalized relative humidity (h_N = 60.7 %)	+2.0 %	–1.5	+7.0	–1.9
	–2.0 %	+1.3	-6.2	+1.6
3. Isotopic composition of atmospheric water vapour ($\delta^{18}O_A$ = –20.4 ‰)	+1.0 ‰	+7.6	-7.1	+1.7
	–1.0 ‰	–7.6	+8.2	–2.0
4. Isotopic composition of lake water ($\delta^{18}O_{LS}$)	+0.5 ‰	–15.8	+19.6	–3.3
5. Isotopic composition of the total inflow ($\delta^{18}O_{IT}$)	+0.5 ‰	+20.2	–14.8	+4.4

(a) The (+) and (-) signs signify an increase or reduction, respectively, of the derived quantity by the reported percentage.

Figures

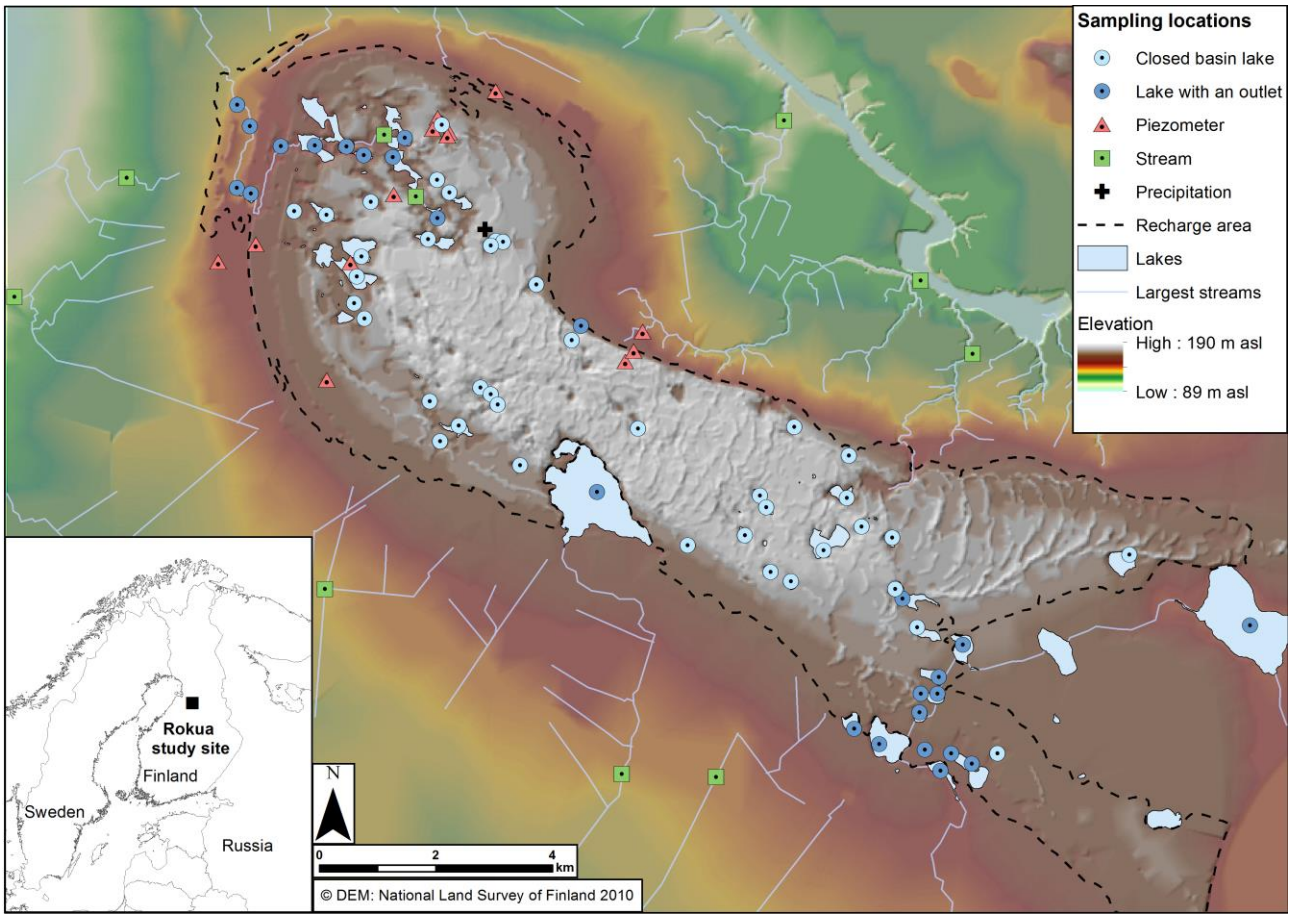


Figure 1. The study site of Rokua esker aquifer area. Digital elevation model by the National Land Survey of Finland (2010).

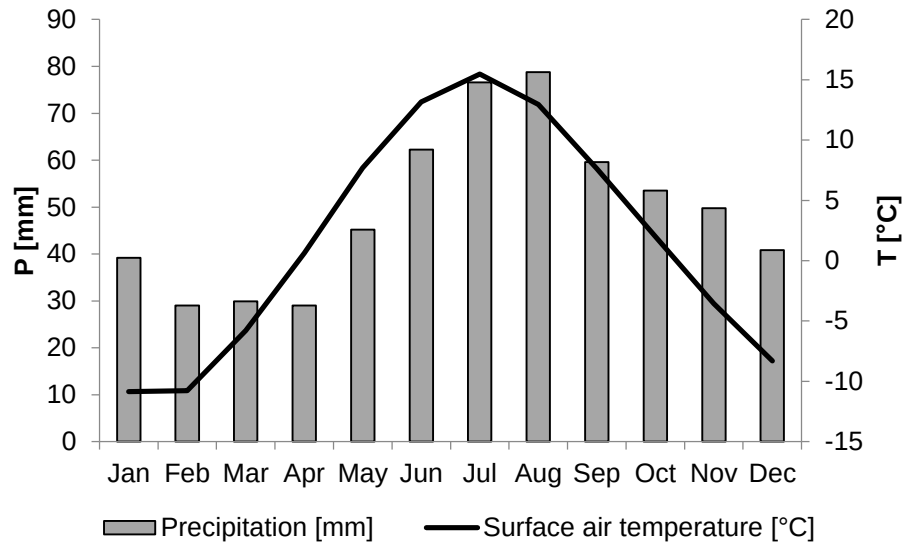


Figure 2. The long-term (1959-2013) monthly mean values of surface air temperature and the amount of precipitation recorded at the station located 10 km south-west of the study site (Finnish Meteorological Institute, 2014).

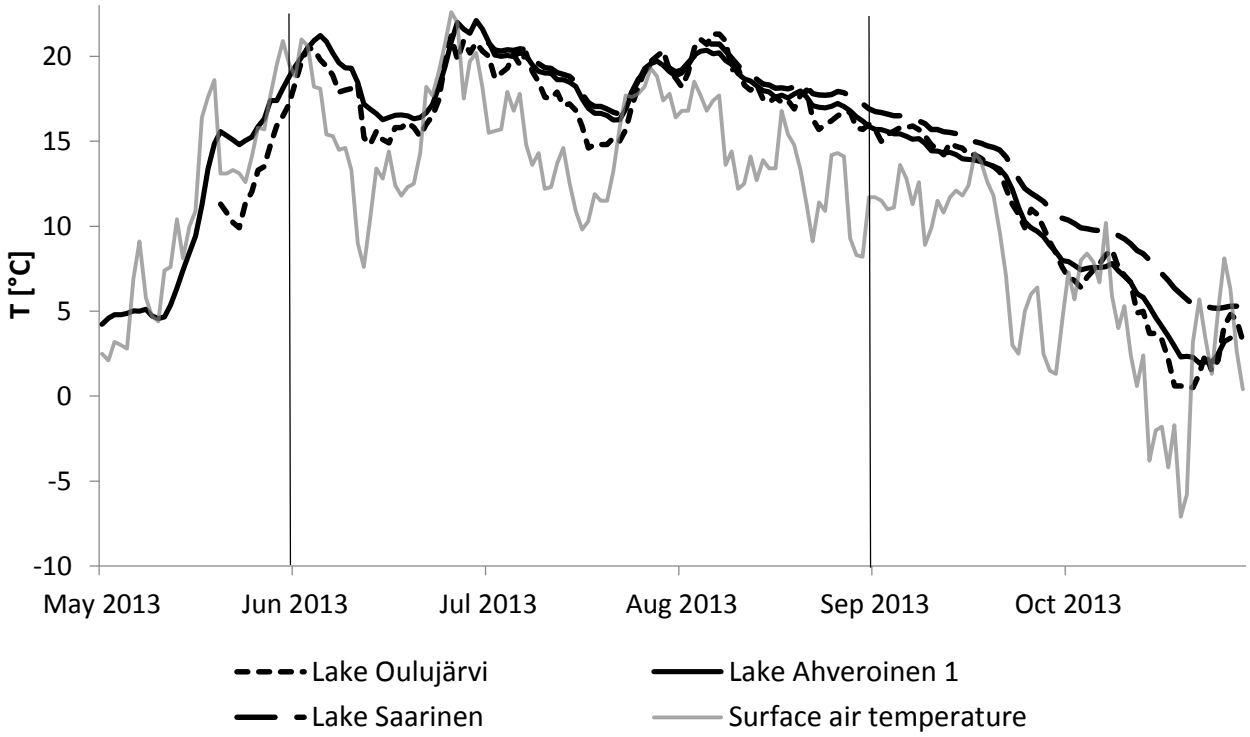


Figure 3. Daily mean surface water temperatures of lakes Oulujärvi (92,800 ha) (Finnish Environmental Institute, 2013), Ahveroinen 1 (3.3 ha) and Saarinen (15.3 ha) during the summer of 2013, compared with the surface air temperature data for the same period. Lake Oulujärvi is located in the east, next to the study site, 1 km from the easternmost lake studied. Vertical lines mark the period used in the calculations of evaporation and isotope mass balance.

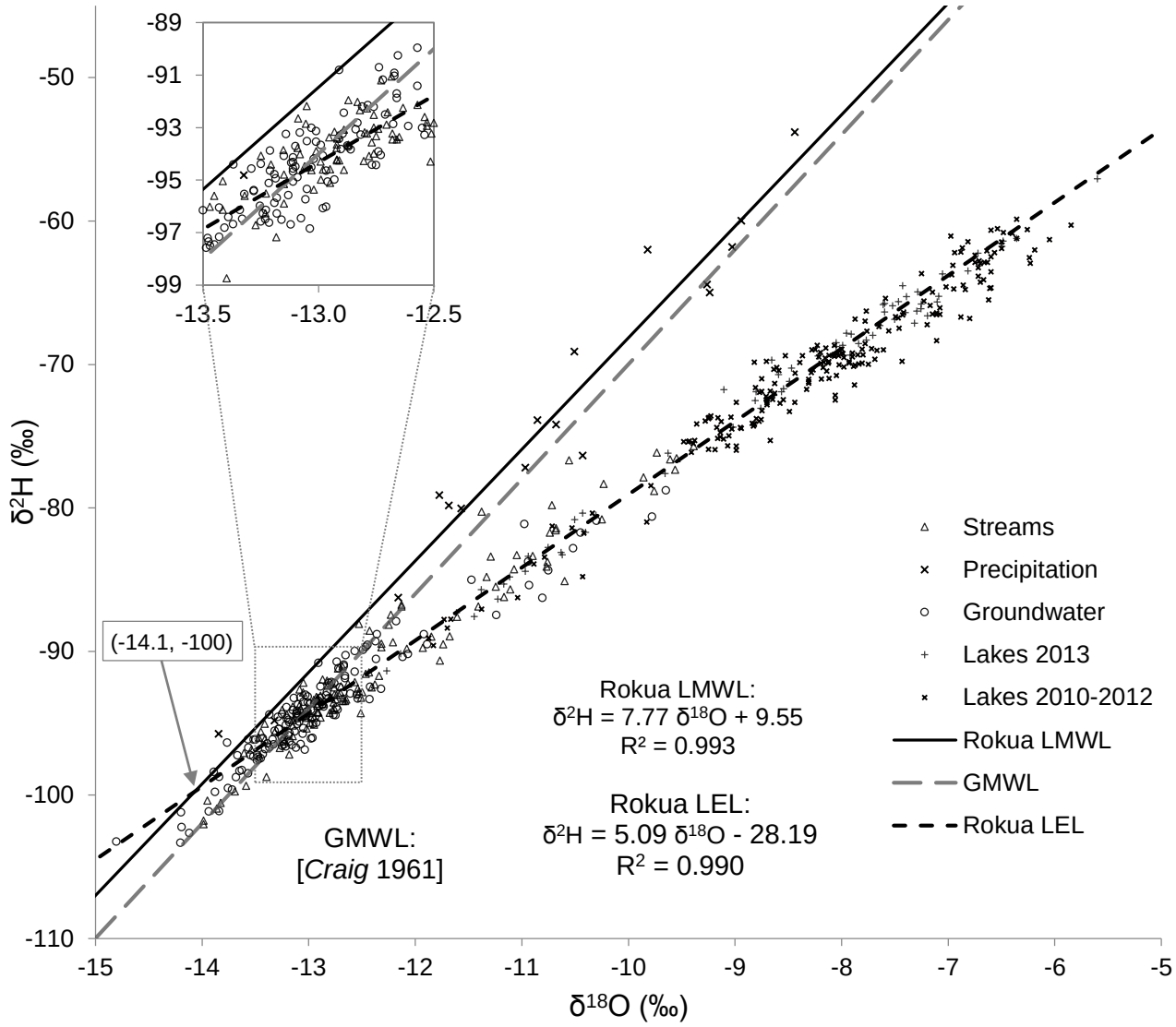


Figure 4. δ^2H - $\delta^{18}O$ relationship for different appearances of surface water (lakes, streams) and groundwater in the study area, investigated within the scope of this study. The Rokua evaporation line (local evaporation line - LEL) was defined as the best fit line of the data points representing lakes sampled during the July-August 2013 campaign.

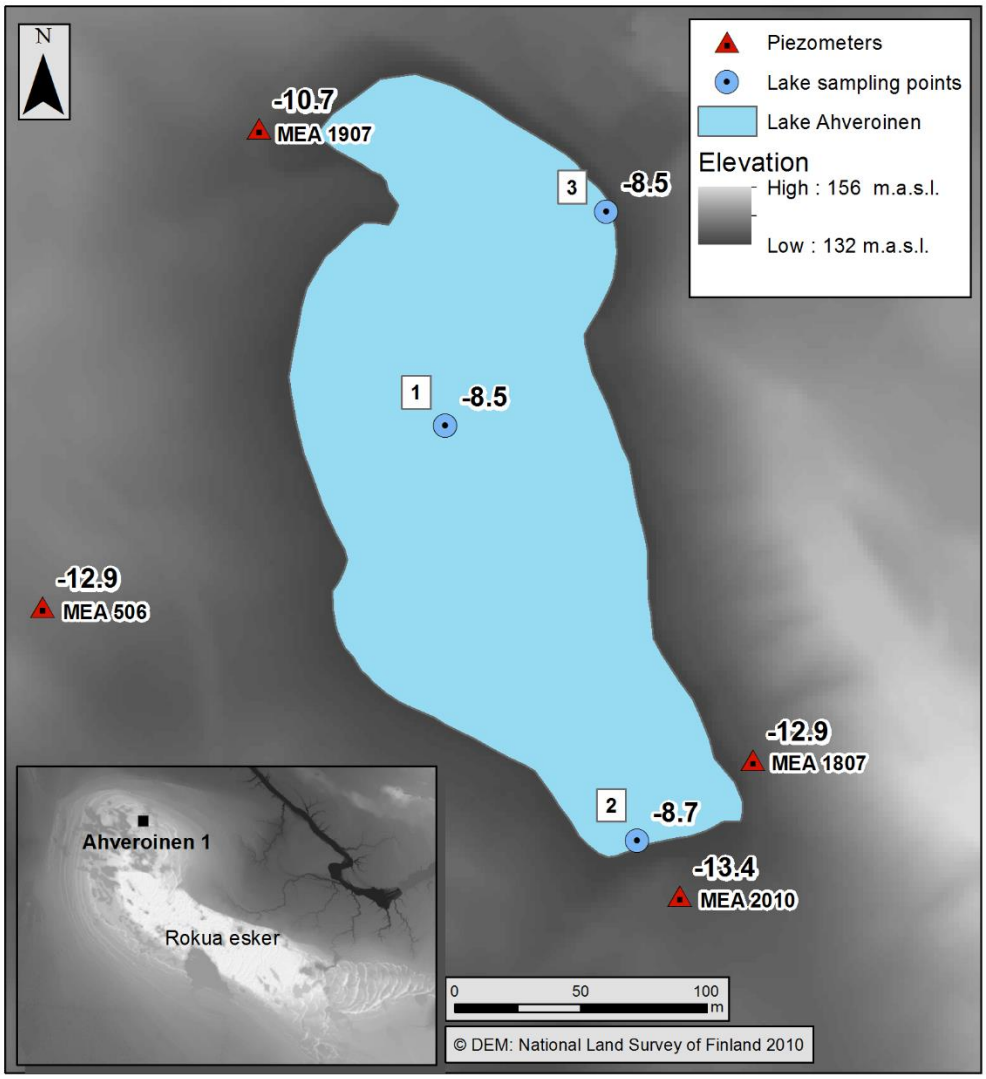
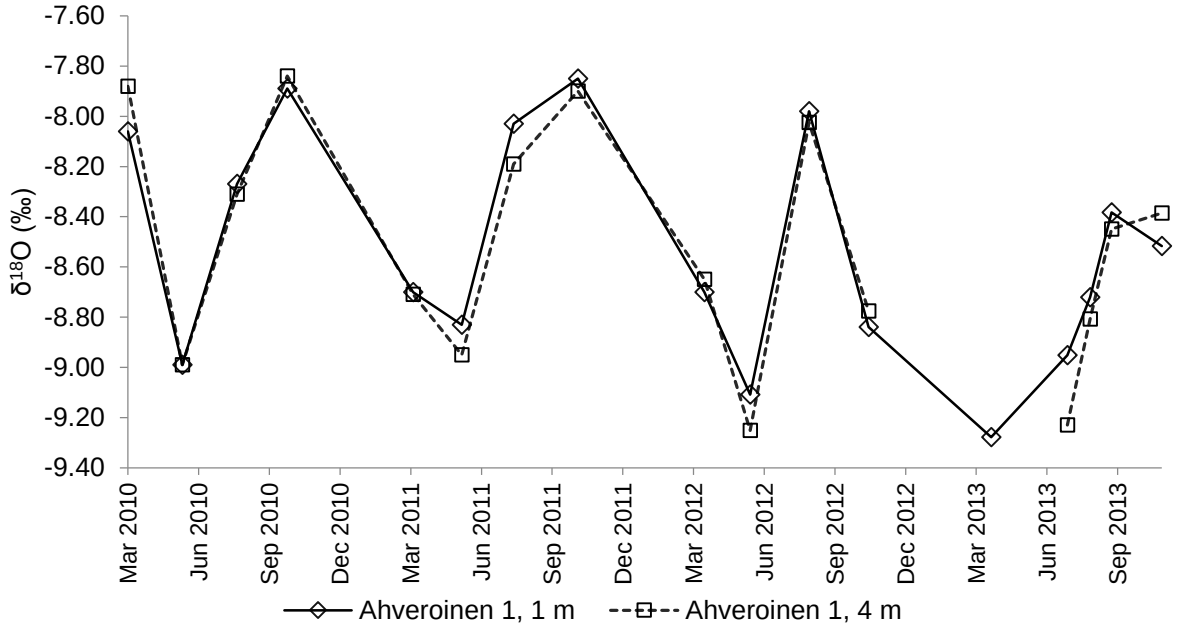
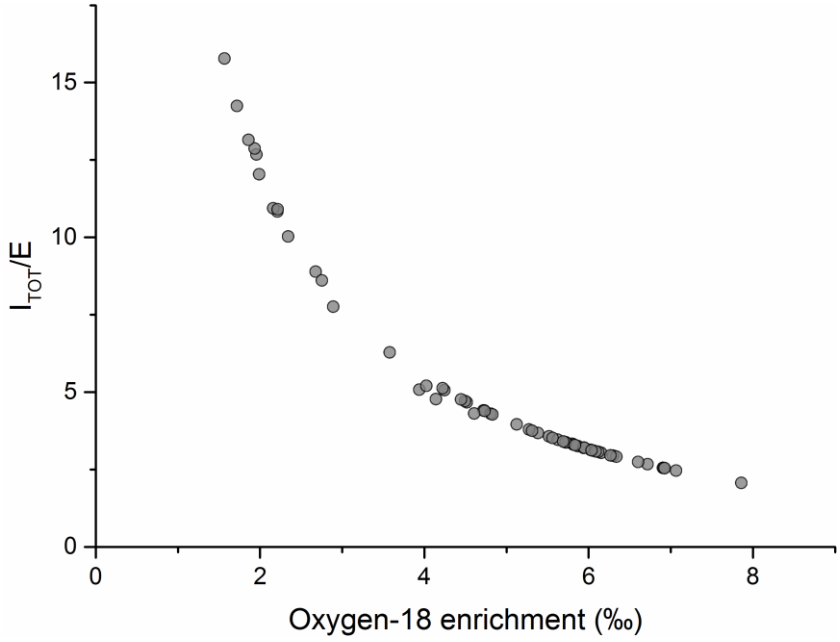


Figure 5. Mean $\delta^{18}O$ values (‰) of lake Ahveroinen and adjacent groundwater. The mean $\delta^{18}O$ value for site 1 is 8.5 ‰ at both sampled depths (1 m and 4 m).



1042

1043 Figure 6. Seasonal variations of $\delta^{18}O$ in lake Ahveroinen 1, observed at 1 and 4-meter depths.
1044 Maximum depth of the lake is 4.8 m.



1045

1046 Figure 7. The ratio of total inflow to evaporation (I_{TOT}/E) as a function of the measured ^{18}O isotope
1047 enrichment ($\Delta\delta^{18}O$) for Rokua lakes surveyed during the July-August 2013 sampling campaign.

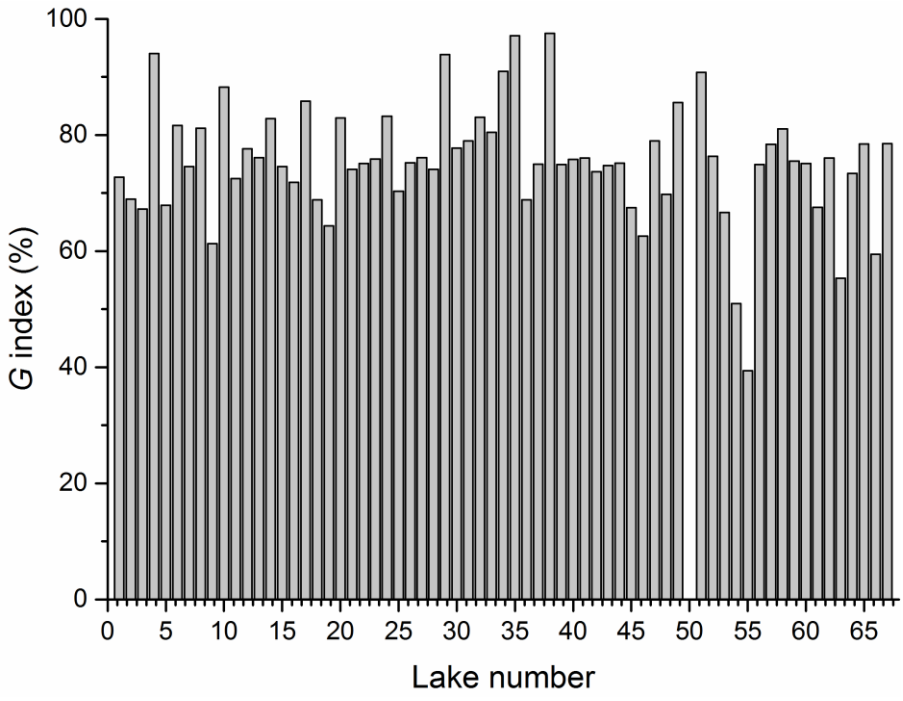


Figure 8. The *G* index quantifying the groundwater dependency of lakes in the Rokua study area. The index is defined as the percentage contribution of groundwater inflow to the total inflow of water to the given lake. Shown are the mean *G* values obtained from independent isotope mass balance calculations based on ^{2}H and ^{18}O data.

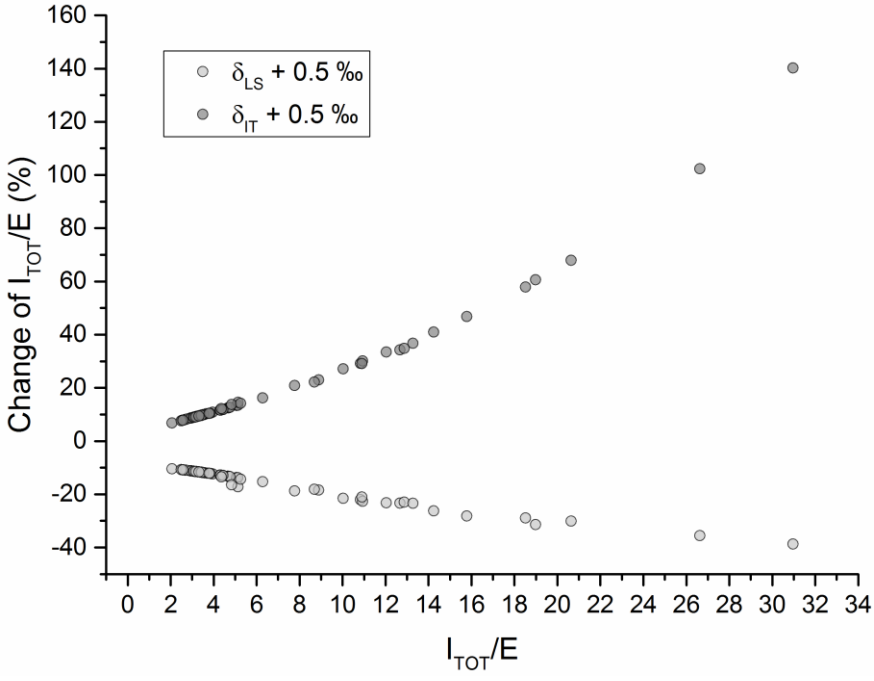


Figure 9. Changes of the total inflow-to-evaporation ratio (in %) based on ^{18}O isotope mass balance as a function of I_{TOT}/E value, in response to parameter change, calculated for the studied lakes on the Rokua esker. Two cases are considered: (a) an increase of the measured $\delta^{18}O$ of lake water by 0.5 ‰, and (b) an increase of $\delta^{18}O$ of the total inflow to the given lake by 0.5 ‰. See text for details.