

the atmospheric input of chloride in precipitation and dry deposition, concentrates in residual soil water via evapotranspiration processes (Allison and Hughes, 1978; Allison, 1988; Guan et al., 2010). Hence, the conventional CMB essentially estimates diffuse recharge through the soil profile. The CMB method can be applied to a saturated zone (Wood and Sanford, 1995) by measuring groundwater chloride or an unsaturated zone (Walker et al., 1991; Cook et al., 1992) by measuring chloride in soil water. Wood and Sanford (1995) and Wood (1999) highlight the necessary conditions for application of the conventional CMB method, as well as some of its limitations.

Accurate estimation of recharge to aquifers in arid and semi-arid areas remains challenging, especially where point recharge (recharge from discrete locations) dominates. The main distinctions between karstic and non-karstic aquifers are the duality in porosity, infiltration and recharge. It is these properties of karst systems that generate point recharge and lead to uncertainty in estimated recharge using the conventional CMB method (Somaratne et al., 2013).

In South Australia, the Uley South (113 km²) and Robinson basins (4.7 km²) on the Eyre Peninsula are characterized by significant karst developments in the form of dissolution features, which represent natural discrete points of recharge. In the city of Mount Gambier (26.5 km²) within the Blue Lake capture zone, stormwater is discharged directly to an unconfined aquifer through large sinkholes and 400 drainage wells (EPA, 2007) within an area of 16.8 km². Drainage wells are discrete recharge points and may be considered as small man-made sinkholes. In the Tatiara catchment, southeast of South Australia, the total creek flow discharges to Poocher Swamp, with the majority of recharge taking place through two sinkholes located in the north-west section of the swamp forming a fresh water bubble of 20 km² downgradient (Somaratne et al., 2013). When point recharge becomes a contributing recharge mechanism, Somaratne et al. (2013) shown that conventional CMB method requires modification to include both point and diffuse recharge mechanisms.

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Several modifications to the conventional CMB equations are presented in the literature. The conventional CMB is expressed as:

$$R = \frac{P c_{p+D}}{c_g} \quad (1)$$

where R is recharge (L T⁻¹), c_{p+D} (ML⁻³) is the representative mean chloride concentration in rainwater including contributions from dry deposition (Ordens et al., 2012), which is estimated using the empirical formula developed by Hutton (1976), and c_g is chloride concentration in groundwater (ML⁻³) resulting from diffuse recharge which is in equilibrium with unsaturated zone chloride concentration (Walker et al., 1991; Cook et al., 1992). For analysis of recharge in playa basins, Wood and Sanford (1995) modified the conventional CMB equation to include chloride loading from runoff in the mass balance of Eq. (1); this may be written for the unsaturated zone in the form:

$$R = \frac{P c_{p+D}}{c_u} + \frac{r(A_b c_r)}{A_f c_u} \quad (2)$$

where r is runoff (L) from the basin, A_b is drainage area of the basin (L²), A_f is area of the playa lake floor receiving the drainage (L²), c_r is average chloride concentration of the runoff (ML⁻³), and c_u is average chloride concentration in the unsaturated zone (ML⁻³). Wood et al. (1997) classify the mode of groundwater recharge into interstitial (matrix) and macropore (fractures, cracks, solution features, natural pipes, animal burrows, root tubes and other openings) flow. For application to playa basins with macropore flow, Wood et al. (1997) use an equation of the form:

$$Q_{tb} = Q_{mb} + Q_{ib} \quad (3)$$

where Q_{tb} is total recharge through the basin floor (L³T⁻¹), Q_{mb} is recharge through macropore (L³T⁻¹), and Q_{ib} is recharge through interstitial pores (L³T⁻¹). Wood (1999)

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to the unconfined aquifer through three sinkholes and 400 storm water drainage wells (Somaratne et al., 2013). The third case study reported by Somaratne et al. (2013) is the Poocher Swamp's fresh water bubble, which is the largest of fresh water plumes that float on brackish water in the area, is a result of flows from Tatiara Creek which enter Poocher Swamp (Fig. 3). The major recharge is through two sinkholes located in the north-west section of the swamp (Somaratne et al., 2013). The area encompassed by the 1000 mgL^{-1} salinity contour comprises approximately 20 km^2 . The unconfined aquifer is Murray Group Limestone and contains brackish water with average TDS (total dissolved solids) $> 1400 \text{ mgL}^{-1}$, with a chloride concentration of $> 500 \text{ mgL}^{-1}$ (MacKenzie, 2013). Saturated thickness of the limestone unconfined aquifer is approximately 50–60 m.

Additional data collected by Somaratne et al. (2013) include water samples from surface runoff from Tatiara creek, Poocher Swamp, and from storm runoff to drainage wells from Mount Gambier for chloride analysis. In addition, groundwater samples were collected from diffuse and point recharge dominant zones of the three study basins to supplement existing data and fill data gaps. Selected sampling wells are away from brackish water upward leakage sites and the swamp in Uley South basin and historically known contaminated sites in Mount Gambier.

Three approaches were used for quantifying surface runoff to point recharge sources. For the Uley South basin, runoff estimation is based on the study of Ward et al. (2009). Ward et al. (2009) used LEACHM (Hutson 2003), a variably saturated model of the soil profile that uses the curve number approach described by Williams (1991) to estimate surface runoff. The model considers four surface cover scenarios, catchment slope, vegetation cover, potential evapotranspiration, and four soil and sub-soil profiles as key variables. According to Ward et al. (2009) timing and magnitude of recharge from steeply sloped sites is relatively independent of vegetation type and cover. The LEACHM modelling result is critically dependent on the assumption that all runoff becomes recharge via sinkholes. The simulations were not sensitive to the four soil and sub-soil profiles. The only profile that showed a slight difference was the

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profile with 300 mm of soil on calcrete. This site showed a tendency for slightly less runoff, than the other soil profiles. In general, however, the four profiles considered did not lead to significant differences in recharge hence, the recharge value obtained is the average over the four soil profiles.

For the Mount Gambier urban catchment, Nguyen (2013) used the urban storm water model MUSIC (2009) to quantify storm water runoff using daily rainfall and pan evaporation. This model considers rainfall threshold for the impervious area, soil moisture storage and field capacity, and infiltration coefficient for the pervious area to calculate runoff. A sensitivity analysis indicates field capacity of the soil had the greatest effect on runoff from the pervious area (Nguyen, 2013). In the Nguyen study, MUSIC modelled rainfall and runoff processes for the period 2007–2012 using a daily time step with daily rainfall and evaporation data. For sub-catchments with drainage wells, average percentage of impervious (51 %) and pervious (49 %) areas was determined using digital maps of the city using Geographic Information System (GIS) tools. A rainfall threshold of 1 mm was used for impervious areas. Uniform soil storage capacity and field capacity values of 120 and 80 mm were used for the pervious areas. The initial soil storage capacity was set to 30 % (Nguyen, 2013). The average annual runoff volume from both pervious and impervious areas were calculated as point recharge to drainage wells and three sinkholes.

In Poocher Swamp, surface water is present intermittently, usually from midwinter to early spring (July–September). Swamps tend to form in shallow, low-lying depressions where there is clay topsoil. Water in the swamps does not become saline, therefore it must be partially open to leakage to groundwater (Herczeg et al., 1997). The major outflow is drainage through two sinkholes located in the north-west section of the swamp. Chloride concentrations of groundwater near the Poocher Swamp sinkhole (about $20\text{--}40 \text{ mgL}^{-1}$) are much lower than those in regional groundwaters ($300\text{--}500 \text{ mgL}^{-1}$). This indicates input of low chloride swamp water (about 28 mgL^{-1}) directly into the groundwater system adjacent to the swamp (Herczeg et al., 1997; Somaratne, 2011a) creating a freshwater plume.

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To illustrate the above concept, Somaratne (2011b) investigated this phenomenon in small sub-catchments (0.03–0.12 km²) discharging storm water to drainage wells in the city of Mount Gambier. Thus, drainage wells are used to measure chloride concentration in the aquifer at discrete recharge points, due to their location in a karstic limestone aquifer similar to the Uley South basin. As one might expect in karstic systems, the hydraulic parameters obtained from pumping tests indicate a high degree of heterogeneity. In Uley South, transmissivity ranged from 3000 to 13 000 m² day⁻¹ and with specific yields from 0.03 to 0.72 (Evans, 1997) with a basin average of 0.25–0.3 (Davis, personal communication, 2012; Davis et al., 2012) for the Quaternary Limestone aquifer. In the Blue Lake capture zone, transmissivity is in the range of 450–24 000 m² day⁻¹ and specific yield is 0.1–0.4 for the Gambier Limestone aquifer (Mustafa and Lawson, 2002). These small pockets of fresher water around point recharge sources, which result from surface water entering the aquifer bypassing the soil matrix, affect the CMB, and hence estimated recharge. It is not possible to measure chloride in fresh water pockets around solution features such as sinkholes in Uley South because the depth, extent and distribution of flow paths in the karst is poorly understood, and is likely to remain unknown.

Water samples were taken from drainage wells and aquifer monitoring wells at three depth intervals in the Gambier Limestone unconfined aquifer. Ten surface runoff samples, 21 samples from drainage wells and 21 samples from monitoring wells were collected. Measured mean chloride concentrations with standard deviation at stormwater entry points to drainage wells, within drainage wells, and at water level monitoring wells, are given in Fig. 4. This example illustrates that the chloride concentration in aquifer monitoring wells differs from the chloride concentration in fresh water pockets at or near drainage wells.

2.2 Derivation of a general form of the CMB equation

If it is possible to account for the mass of chloride in the system, the mode of transport of chloride from the land surface to groundwater becomes irrelevant. In karst aquifers,

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this is an impossible task (Somaratne et al., 2013). The chloride concentration at the discrete recharge points remains close to the surface water concentration, and elsewhere remains close to the diffuse recharge groundwater chloride concentration.

The conventional CMB equation derived by Nyagawambo (2006) is modified in this work to include point recharge. Following the conceptual model developed for the Uley South basin (Somaratne et al., 2013), duality of recharge is considered in the development of a generalized CMB. The method integrates chloride mass balance from the ground surface to the saturated zone in a control volume as shown in Fig. 5.

In this model, sinkholes connected to the watertable bypass the soil zone, directly recharging the aquifer as point recharge. The unconnected sinkholes add runoff deeper into the unsaturated zone, which then rapidly drains into the watertable by mechanisms described by Gunn (1983). A system approach is adopted, in which the hydrologic function of each primary component: ground surface, vadoze zone and saturated zone is considered separately as an integral part of a whole system (Taylor and Greene, 2001).

Following Nyagawambo (2006), the chloride mass balance at the soil surface can be written:

$$\frac{\Delta(s_s c_s)}{\Delta t} = P c_{p+D} - (Q_p + Q_o + F) c_s \quad (4)$$

Note that Q_p and Q_o are expressed as depths of the catchment. Following Nyagawambo (2006), assumptions are made to simplify Eq. (4). The storage fluctuation term may be assumed to be negligible relative to inflows and outflows, if the time of integration is sufficiently long to cover several hydrological years (Nyagawambo, 2006). Water is assumed to evaporate in its pure form and therefore no chloride is lost through evaporative fluxes. It is also assumed that the chloride concentration in surface flows at the point of runoff generation and the point of infiltration remains the same as in surface water. For a negligible change in chloride concentration at the soil surface, Eq. (4) reduces to:

$$F c_s = P c_{p+D} - (Q_p + Q_o) c_s \quad (5)$$

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(Werner, 2010) is based on MODFLOW (McDonald and Harbough, 1988) and PEST (Watermark Numerical Computing, 2004); and 157 mm long-term average recharge value of the Eyre Region Water Resources Planning Committee (2000) is based on Darcy flow and water balance calculations of Evans (1997). However, generalized CMB
 5 estimated recharge is significantly higher than conventional CMB estimated average annual recharge of 53–60 mm (Ordens et al., 2012), 52–71 mm (SA Water unpublished data) and 71 mm (Evans, 1997) for the basin. This highlights the fact that application of
 10 the conventional CMB method to estimate total recharge in groundwater basins characterised by point and diffuse recharge is unsuccessful, either using unsaturated or saturated zone chloride as inputs. Thus generalized CMB provides an alternative reliable method for recharge estimation in karstic settings.

4 Conclusions

Sinkholes in depressions are an important source of point recharge to limestone aquifers. Application of the conventional CMB method to point recharge dominant
 15 groundwater basins substantially under estimates long-term average recharge by not accounting for the effect of localized surface water inputs. For application of the generalized CMB, obtaining diffuse recharge groundwater chloride is critical to accuracy. Groundwater chloride measurements should be obtained away from fresh water pockets or plumes created by point recharge or soil-water chloride should be extracted from
 20 the unsaturated zone above the watertable in order to obtain diffuse zone chloride. Quantification of runoff at the catchment or sub-catchment levels is required for accurate assessment of point recharge and to apply the generalized CMB equation for total recharge estimation. For this purpose, many available watershed models provide runoff estimates as part of an overall water balance. A comparison of the modified CMB
 25 results with the long-term recharge estimates obtained using the WTF method, groundwater flow modelling and Darcy flow calculations, shows slightly less but comparable results. Hence the generalized CMB method augments the conventional CMB method

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by accounting for point recharge that by-passes the unsaturated zone and thus provides an alternative, reliable, long-term recharge estimation method for groundwater basins with point and diffuse recharge mechanisms.

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Table 2. Comparison of generalized and conventional CMB estimated recharge. Recharge volume in m³ is given in brackets.

Groundwater basin	Recharge using generalized CMB (mmyr ⁻¹)	Recharge using conventional CMB (mmyr ⁻¹)
Uley South	120 (13.6 × 10 ⁶)	56 (6.3 × 10 ⁶)
Blue Lake capture zone	457 (7.6 × 10 ⁶)	95 (1.6 × 10 ⁶)
Poocher Swamp fresh water bubble	127.3 (2.55 × 10 ⁶)	14 (0.28 × 10 ⁶)

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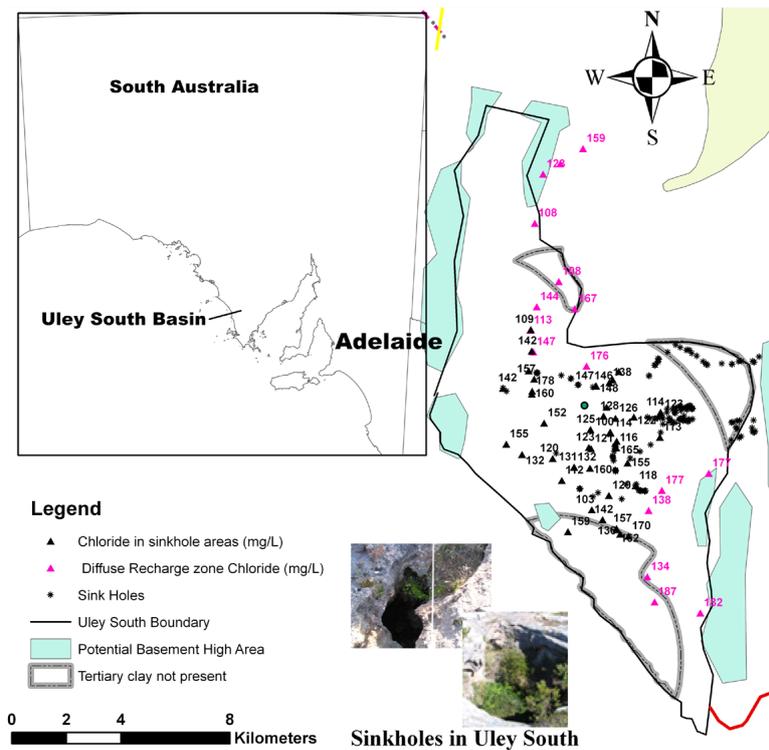


Fig. 1. Sinkholes in Uley South Basin and groundwater chloride (Somaratne et al., 2013).

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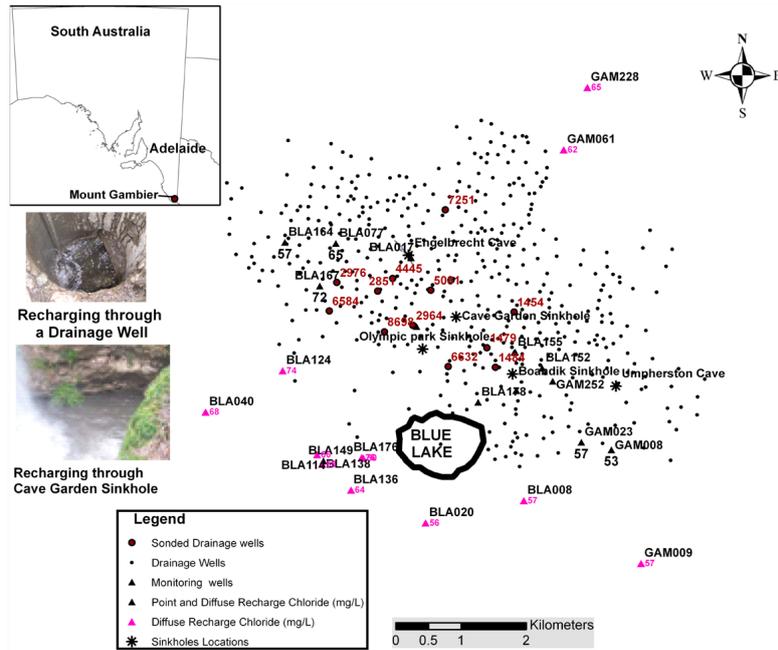


Fig. 2. Drainage wells and chloride in aquifer monitoring wells in the Blue Lake capture zone (Somaratne et al., 2013).

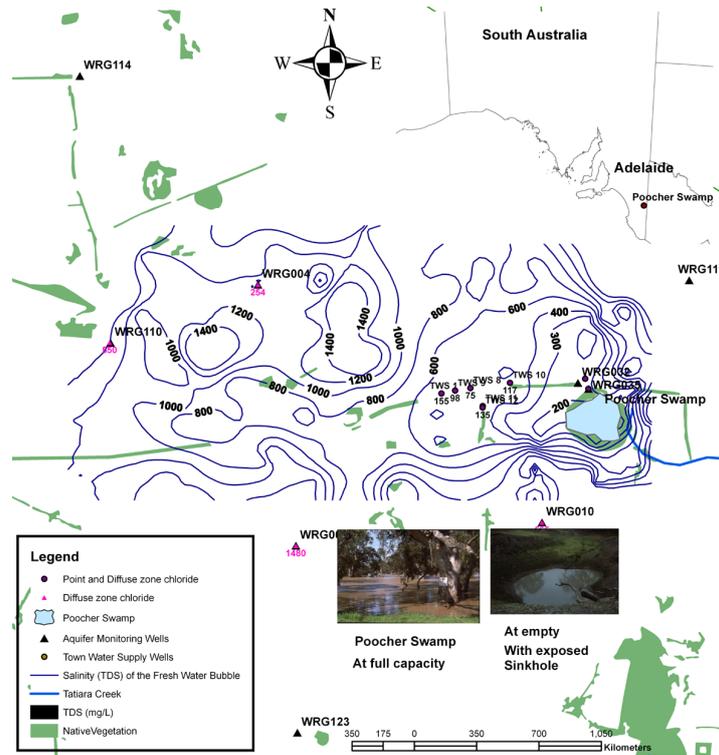


Fig. 3. Poacher Swamp fresh water bubble (Somaratne et al., 2013).

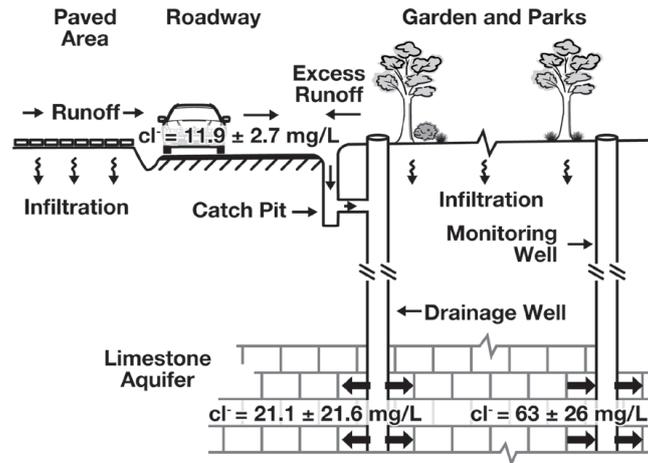
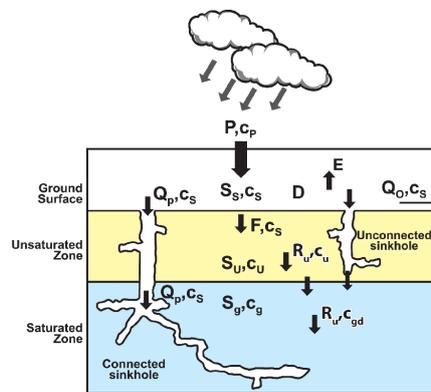


Fig. 4. Schematic diagram showing an example of differing chloride concentrations in drainage and aquifer monitoring wells.

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- S_s = Surface Storage (L)
- c_s = Chloride concentration of the surface water (ML^{-3})
- P = Average annual rainfall (LT^{-1})
- c_p = Chloride concentration of the rainfall (ML^{-3})
- D = Dry deposition ($MT^{-1}L^{-2}$)
- E = Evapotranspiration (LT^{-1})
- Q_p = Runoff to sinkholes (LT^{-1})
- Q_o = Runoff out from the catchment (LT^{-1})
- F = Infiltration into soil profile (LT^{-1})
- S_u = Unsaturated zone storage (L)
- c_u = Chloride concentration in recharging water (ML^{-3})
- R_u = Diffuse recharge through the soil profile (LT^{-1})
- S_g = Groundwater storage (L)
- c_g = Groundwater chloride concentration (ML^{-3})
- c_{gd} = Chloride concentration of groundwater in diffuse recharge zone (ML^{-3})

Fig. 5. Schematic diagram for applying chloride mass balance to a control volume.

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