

1 **Residence times and mixing of water in river banks:**
2 **implications for recharge and groundwater – surface water**
3 **exchange**

4
5 **Nicolaas P. Unland** ^{1,2}, **Ian Cartwright** ^{1,2}, **Dioni I. Cendón** ^{3,4}, **Robert Chisari** ³

6
7 ¹School of Earth, Atmosphere and Environment, Monash University, Clayton, Vic 3800,
8 Australia

9 ²National Centre for Groundwater Research and Training, School of Environment, Flinders
10 University, Adelaide, SA 5001, Australia

11 ³Australian Nuclear Science and Technology Organisation, Kirrawee DC, NSW 2232,
12 Australia

13 ⁴School of Biology, Earth and Environmental Sciences, University of New South Wales,
14 NSW 2052, Australia

15 Correspondence to: N. P. Unland (nicolaas.unland@gmail.com)

16
17 **Abstract**

18 Bank exchange processes within 50 m of the Tambo River, South East Australia, have been
19 investigated through the combined use of ³H and ¹⁴C. Groundwater residence times increase
20 towards the Tambo River, which suggests the absence of significant bank storage close to the
21 Tambo River. Major ion concentrations and $\delta^2\text{H}$ and $\delta^{18}\text{O}$ values of bank water also indicate
22 that bank infiltration does not significantly impact groundwater chemistry under baseflow and
23 post-flood conditions, suggesting that the gaining nature of the river may be driving the return
24 of bank storage water back into the Tambo River within days of peak flood conditions. The
25 covariance between ³H and ¹⁴C indicates the leakage and mixing between old (~17,200 years)
26 groundwater from a semi-confined aquifer and younger groundwater (<100 years) near the
27 river where confining layers are less prevalent. It is likely that the upward infiltration of
28 deeper groundwater into the semi-confined aquifer during flooding limits bank infiltration.
29 Furthermore, the more saline deeper groundwater likely controls the geochemistry of water in

1 the river bank, minimising the chemical impact that bank infiltration has in this setting. These
2 processes, coupled with the strongly gaining nature of the Tambo River are likely to be the
3 factors reducing the chemical impact of bank storage in this setting. This study illustrates the
4 complex nature of river groundwater interactions and the potential downfall in assuming
5 simple or idealised conditions when conducting hydrogeological studies.

6 **1 Introduction**

7 Documenting water balances in river systems is vitally important to understanding
8 hydrological processes and protecting and managing water resources. While surface runoff
9 and regional groundwater inflows are the two main components of river flow, river banks
10 may act as sites of transient water storage. Bank storage represents water that infiltrates into
11 alluvial aquifers at high river stage and subsequently returns to the river as the river stage
12 declines (e.g., Chen and Chen, 2003; McCallum et al., 2010; Singh, 1968; Winter, 1998).
13 Bank storage is an important hydrological process that may considerably reduce peak river
14 discharge during floods and maintain river discharge during periods of decreased rainfall
15 (Pinder and Sauer, 1971). In addition, bank waters may represent a source of nutrients or
16 contaminants derived from the river that are gradually released following diminishing of the
17 flood peak. The volume and duration of bank storage for a given river stretch will depend on
18 the flood peak height and the flood duration (Cooper and Rorabaugh, 1963), as well as the
19 hydraulic conductivity of the alluvial aquifer and the hydraulic gradient between the aquifer
20 and river (Cartwright et al., 2014; Chen et al., 2006; McCallum et al., 2010). Whiting and
21 Pomeranets (1997) showed that deeply-incised narrow rivers with wider floodplains and
22 coarse alluvial material have greater bank storage potential. The potential for significant
23 storage beneath the streambed was identified by Chen and Chen (2003), while Chen et al.
24 (2006) showed that bank storage will return more rapidly in gaining river sections. Bank slope
25 has also been shown to impact bank storage, with shallower bank slope providing a greater
26 potential for bank storage (Doble et al., 2012).

27 The concentrations of solutes in river water are commonly lower than those in groundwater
28 and mixing between infiltrating river water and groundwater may produce zones of lower
29 solute concentrations in river banks. McCallum et al. (2010) showed that solute
30 concentrations of bank water may take months to return to their original concentration.
31 During that time period, the water that flows from the banks into the river is a mixture of
32 regional groundwater and bank storage waters. This suggests that there may be a component

1 of bank storage waters in river banks with a residence time of months to years. Recognising
2 that bank storage waters may contribute to rivers over long timeframes is important for
3 estimating groundwater discharge by chemical mass balance. If the bank storage waters are
4 chemically similar to surface water rather than regional groundwater, using the composition
5 of regional groundwater as an end member will result in underestimation of the groundwater
6 flux (Cartwright et al., 2014; McCallum et al., 2010; Unland et al., 2013).

7 While the concept of bank storage is well understood, accurately quantifying the volume of
8 water that infiltrates the banks and the duration of bank return flows remains difficult. Many
9 studies have focused on using analytical and numerical solutions to understand bank storage
10 from variations in the river hydrograph and groundwater heads. Most of these studies have
11 concluded that bank storage periods will significantly exceed the duration of flood events.
12 Typically bank storage return to the river is proposed to decrease exponentially after flood
13 events, and in the case of sandy river banks with wide floodplains, residence times can be on
14 the order of years (Doble et al., 2012; McCallum, et al., 2010; Whiting and Pomeranets,
15 1997). These studies commonly assume ideal or generalised conditions such as aquifer
16 homogeneity, vertical river banks and saturated conditions (Doble et al., 2012), making them
17 difficult to apply to many natural settings. Therefore, there is a need to document the extent
18 and timescales of bank storage in specific catchments.

19 Geochemical processes occurring within river banks, such as the bacterial degradation of
20 organic matter or the weathering of minerals can influence the concentrations of DOC, O₂,
21 NO₃, Na, K and other major ions in near-river groundwater (Bourg and Bertin, 1993). Fukada
22 et al. (2003) identified the continuing denitrification of river water as it infiltrated an alluvial
23 aquifer and demonstrated that the chemistry of infiltrating water is likely to vary according to
24 its residence time within the alluvial aquifer. Understanding the source and load of nutrients
25 in rivers is fundamental in understanding their ecology (Boulton, 1993, 2005), while
26 determining the different sources of water in the riparian zone is crucial to effective
27 vegetation management (Cey et al., 1999; Lambs, 2004; Lamontagne et al., 2005; Woessner,
28 2000). Similarly, the impact of infiltrating river water on water quality in alluvial aquifers is
29 important when developing groundwater extraction systems for water supply (Hiscock and
30 Grischek, 2002).

31 It is important to carry out studies of bank storage in a range of environments as variations in
32 climate and river form translate into variations in river regime (e.g., frequency and duration of

1 floods) that may cause differences in bank infiltration. Field studies focussed on bank storage
2 and the dating of bank water in Australian catchments has been limited. Lamontagne et al.
3 (2011) and Cendón et al. (2010) indicated the presence of relatively young (<50 years)
4 groundwater in river banks, and Cartwright et al., (2010) showed that preferential floodplain
5 recharge is likely to occur near rivers during flooding. In contrast, groundwater in upland
6 catchments in Australia has been shown to have relatively long residence times (Atkinson et
7 al., 2013). This study investigates bank storage processes in the Tambo River catchment,
8 Victoria, Australia. The objectives of the study are to use the geochemistry of groundwater in
9 the banks of the Tambo River at different discharges in order to: (1) define the major
10 processes controlling the chemistry of water stored in river banks; (2) determine the age and
11 likely sources water stored in river banks; and (3) identify the factors controlling bank storage
12 and the distance over which bank storage is occurring. While this study uses data from
13 specific field area, the Tambo River is similar to many others globally and the results will
14 help in understanding bank storage processes in general.

15 **1.1 Study area**

16 Investigations took place on the middle catchment of the Tambo River in southeast Australia.
17 The Tambo River is perennial and flows through forest and woodland with cattle grazing on
18 the river floodplains (Department of Agriculture, Fisheries and Forestry, 2006). It discharges
19 into the saline Lake King and the lower ~15 km of the river is estuarine. Average annual
20 precipitation in the catchment increases from 655 mm in the upper reaches to 777 mm in the
21 middle and lower reaches (Bureau of Meteorology, 2013). During the majority of the study
22 period, the discharge of the Tambo River ranged from 10^{10} to 10^{11} m³/sec (Victorian Water
23 Resources Data Warehouse, 2013); however significant rainfall during August 2011 and
24 March 2012 resulted in discharge events that peaked at greater than 5×10^{12} m³/sec (Fig. 2).

25 The upper catchment of the Tambo River drains indurated Ordovician and Devonian
26 turbidites and granites of the Eastern Victorian Uplands, while the lower and middle
27 catchment is in the Gippsland Basin (Birch, 2003). The near-river sediments in the lower and
28 middle catchment comprise coarse Quaternary alluvial gravels and sands. These recent
29 alluvial sediments overly the Plio-Pleistocene Haunted Hill Gravels which represents the
30 shallowest regional-scale aquifer in the Gippsland Basin. Clay layers throughout the
31 Quaternary alluvium and Haunted Hill Gravels act as aquitards, separating a number of
32 aquifer horizons that range from unconfined to fully confined (Hocking, 1976). Deeper

1 aquifer systems in the lower Tambo catchment include the Late Miocene to Early Pliocene
2 Boisdale Formation and the Oligocene to Pliocene Jemmys Point, Tambo River and Lake
3 Wellington Formations (Leonard, 1992; Birch, 2003; Hofmann and Cartwright, 2013). These
4 deeper aquifer systems are separated from the near-surface sediments by clay layers that
5 locally form aquitards. Regional groundwater flow in all aquifers is from the margins of the
6 Gippsland Basin towards Lake King (Southern Rural Water, 2013). Overall, the lower and
7 middle reaches of the Tambo River are gaining especially at low river discharge and hydraulic
8 gradients at these times in the shallow alluvial sediments are towards the river (Unland et al.,
9 2013). The regional groundwater from the deeper aquifers is artesian and head gradients
10 around Lake King (Fig. 1) are upwards (Water Resources Data Warehouse, 2013; Southern
11 Rural Water, 2013). The regional groundwater also has a high salinity (total dissolved solids
12 up to 20,000 mg/L), is generally anoxic, and has a significantly higher temperature (up to 40
13 °C) than that of the shallow groundwater. Springs fed by this deeper groundwater are recorded
14 around Lake King; however, these are rare elsewhere (although as discussed below, the
15 regional groundwater may mix with the shallow groundwater and river water in the banks at
16 Tambo Upper, Fig. 1). The majority of groundwater that discharges into the middle reaches of
17 the Tambo River is most likely derived from the shallower aquifers (Southern Rural Water,
18 2013) and the flow is part of a local rather than regional system (c.f., Toth, 1963). This is
19 consistent with the presence of clays in the Gippsland Basin sediments that produce a
20 compartmentalised aquifer system.

21 Transects of groundwater monitoring bores were set up at three locations on the river banks of
22 the middle Tambo River (Fig. 1, Tables 1, 2 and 3). The transect at Bruthen is 28.5 km
23 upstream of Lake King and consists of 3 bores installed at 5.5, 17.6 and 18.3 m distance from
24 the river and 8.0, 5.4 and 7.1 m depth below ground surface, respectively (Fig. 1). The
25 transect at Tambo Upper, 20.2 km upstream of Lake King, consists of 5 bores installed at 8.8,
26 15.0, 22.3, 23.8 and 37.9 m distance from the Tambo River and 6.7, 6.2, 23.1, 6.7 and 9.8 m
27 depth below ground surface, respectively. The final transect at Kelly Creek, 13.8 km upstream
28 of Lake King, consists of 4 bores installed at 7.0, 17.9, 24.9 and 26.8 m from the Tambo
29 River at depths of 8.1, 7.8, 28 and 7.9 m depth, respectively. Bores at Tambo Upper have 1.5
30 m screens starting 1 m from the borehole bottom while all other installations have a 3 m
31 screened section set at the bottom of the borehole. Sediment samples taken during installation
32 of the shallower (<15 m deep) bores installation via auger drilling indicate that the alluvial
33 aquifer at all transects is dominated by coarse sands with 10 to 20 cm thick clay layers

1 dispersed throughout. Deeper bores were constructed via mud rotary drilling which tends to
2 preferentially return coarser sediment fractions. While clay layers are harder to identify, the
3 groundwater at >20m depth is artesian (see below) which suggests the presence of a clay-rich
4 confining layer. All bores are screened in the alluvial sands/gravels and (except for the
5 deeper bore at Tambo Upper) probably sample the local shallow groundwater rather than the
6 deeper regional groundwater.

7 **2 Methods**

8 Bore and river elevations were determined to ± 1 cm relative to the Australian Height Datum
9 (AHD) using a Trimble digital global positioning system (DGPS). Bores were sampled using
10 an impeller pump set at the screened section and at least 3 bores volumes were pumped before
11 sample collection. Five sets of groundwater samples and 4 sets of river samples were
12 collected between February 2011 and March 2012 at each transect. Sampling during February
13 2011, April 2011 and November 2011 took place at conditions close to baseflow, while
14 sampling during August 2011 and March 2012 took place ~1 week after significant flooding
15 in the catchment (Fig. 2). Rising head slug tests were conducted by pumping bores for ~10
16 minutes with an impeller pump at a rate of 4 L/minute and then allowing groundwater heads
17 to recover. Changes to groundwater levels were recorded using a Rugged TROLL 200 logger
18 logging at 1 second intervals. Hydraulic conductivity was calculated using the Hvorslev
19 method outlined by Fetter (1994). The anisotropy ratio of the sediments was not taken into
20 account as it is not explicitly known and the aim was to provide a general characterisation of
21 the hydraulic properties of the sediments. Commonly, the recovery time was <10 seconds and
22 the estimated hydraulic conductivities are likely to be minimum values.

23 Electrical conductivity (EC) was measured in field to $\pm 1\%$ using a calibrated TPS pH/EC
24 meter and groundwater levels were measured using an electronic water level tape. Water
25 samples were preserved by refrigeration in air-tight polyethylene bottles. HCO_3 and dissolved
26 CO_2 were measured within 48 hours of sample collection by titration using a Hach digital
27 titrator and reagents with a precision of $\pm 5\%$. Anion concentrations were determined on
28 filtered (0.45μ cellulose nitrate filters) samples that using a Metrohm ion chromatograph at
29 Monash University, Clayton, with a precision of $\pm 2\%$ estimated by replicate analysis. Cation
30 concentrations were determined on samples that were filtered and acidified to pH <2 using
31 twice distilled 16 M nitric acid by the Varian Vista ICP-AES at the Australian National
32 University or the Thermo Finnigan X series II, quadrupole ICP-MS at Monash University.

1 Drift during ICP-MS analysis was corrected using internal Sc, Y, In, Bi standards, with
2 replicate analysis returning a precision of $\pm 5\%$. Stable isotope ratios were measured at
3 Monash University using ThermoFinnigan MAT 252 and DeltaPlus Advantage mass
4 spectrometers. $\delta^{18}\text{O}$ values of water were measured via equilibration with He- CO_2 at 32°C for
5 24-48 hours in a ThermoFinnigan Gas Bench. $\delta^2\text{H}$ values of water were measured via reaction
6 with Cr at 850°C using a Finnigan MAT H/Device. $\delta^{18}\text{O}$ and $\delta^2\text{H}$ values were measured
7 relative to internal standards that were calibrated using IAEA SMOW, GISP, and SLAP
8 standards. Data were normalised following (Coplen, 1988) and are expressed relative to V-
9 SMOW where $\delta^{18}\text{O}$ and $\delta^2\text{H}$ values of SLAP are -55.5% and -428% , respectively. The
10 precision (1σ) of the analyses based on replicate analyses is $\delta^{18}\text{O} = \pm 0.2\%$, $\delta^2\text{H} = \pm 1\%$.

11 Samples for ^{14}C and ^3H analysis were collected during the April 2011 sampling period (Fig.
12 2). ^3H water samples were distilled and electrolytically enriched prior to analysis by liquid
13 scintillation (Morgenstern and Taylor, 2009). The ^3H concentrations were expressed in tritium
14 units (TU) with uncertainties ranging from $\sim 25\%$ at the quantification limit (0.13 TU) to $< 6\%$
15 for ^3H concentrations above 1.5 TU. For ^{14}C analysis, the total DIC was converted to CO_2 by
16 acidifying the samples with H_3PO_4 and extracting the liberated CO_2 gas using a custom built
17 extraction line. The CO_2 sample was then heated in a sealed glass tube containing baked CuO
18 and Ag and Cu wire at 600°C for 2 h to remove any sulfur compounds that may have been
19 liberated and subsequently graphitised. The graphite targets were analysed using the STAR
20 AMS at ANSTO following Fink et al. (2004). The activity of ^{14}C is expressed as percent
21 modern carbon (pMC) following Stuiver and Polach (1977). The average error associated
22 with radiocarbon measurements is 0.3%.

23 **3 Results**

24 **3.1 Groundwater elevations and hydraulic conductivities**

25 Groundwater elevation at Bruthen varied between 7.45 m AHD in April 2011 and 8.89 m
26 AHD in August 2011. There was less than 6 cm difference in elevations across the transect
27 during any given sampling period. Groundwater elevations in B1 and B2 were within 3 cm of
28 each other during all sampling periods, while elevations in B3 were 2 to 6 cm higher than in
29 B1 and B2 (Fig. 3). Groundwater elevations at Bruthen were 3 to 4 cm higher than river
30 elevations during all sampling periods except during March 2012 when groundwater levels

1 were approximately 90 cm higher than the river elevation. Rising head slug tests at this
2 transect indicate a hydraulic conductivity of $\sim 8.5 \times 10^{-3}$ m/s.

3 Groundwater elevation in the shallow bores at Tambo Upper ranged from 3.30 m AHD in
4 April 2011 to 4.80 m AHD in August 2011. Groundwater elevations in TU5, TU2 and TU1 in
5 individual campaigns were within 3 to 5 cm of each other. Groundwater elevations in TU4
6 were the lowest in the transect, averaging 3.92 m AHD over the study, approximately 9 cm
7 lower than the average levels in TU1, TU2, and TU5 (Fig. 3). The deeper bore (TU3D) was
8 artesian during all sampling periods; this bore samples a deeper, semi-confined aquifer within
9 the alluvial gravel that has higher heads than the surficial aquifer. During February and April
10 2011, groundwater elevations in this bore were 4.85 m and 4.69 m AHD, respectively, while
11 in all other sampling periods the elevation exceeded that of the casing (5.04 m AHD).
12 Groundwater elevations at Tambo Upper were higher than the river elevation during all
13 periods except April 2011. Groundwater elevation closest to the river (TU1) was 32 cm
14 higher than river elevation in February 2011, 5 cm lower than river water during April 2011, 3
15 cm higher than river elevation during August 2011 and 99 cm greater than river elevation
16 during March 2012. Slug tests at this transect yielded hydraulic conductivities ranging from
17 5.1×10^{-4} to 8.6×10^{-5} m/s in the surficial aquifer, and 1.9×10^{-5} m/s in the semi-confined
18 aquifer.

19 At Kelly Creek, groundwater levels in the shallower bores ranged from 3.07 m AHD in April
20 2011 to 3.68 m AHD in August 2011 (Fig. 3). Groundwater levels in these bores generally
21 decreased with proximity to the river during all sample periods except April 2011.
22 Groundwater levels in the deeper bore at Kelly Creek (KC3D) were higher than the shallow
23 bores, ranging from 3.82 m AHD in February 2011 to 4.33 m AHD in November 2011. Slug
24 tests at this transect indicate hydraulic conductivities ranging from 2.4 to 3.4×10^{-5} m/s.

25 **3.2 Electrical conductivity**

26 Groundwater EC values at Bruthen ranged from 136 to 607 $\mu\text{S}/\text{cm}$. Groundwater at B3 was
27 generally the most saline, ranging from 261 to 607 $\mu\text{S}/\text{cm}$, while that from B1 ranged from
28 136 to 293 $\mu\text{S}/\text{cm}$. Shallow groundwater at Tambo Upper was more saline than that from
29 Bruthen, ranging from 717 $\mu\text{S}/\text{cm}$ to 2,682 $\mu\text{S}/\text{cm}$. Shallow groundwater at Tambo Upper was
30 also generally more saline closer to the river than further from the river, averaging 2,110
31 $\mu\text{S}/\text{cm}$ at TU1 and TU2 over the study period, compared to 980 $\mu\text{S}/\text{cm}$ at TU4 and TU5.

1 Deeper groundwater at Tambo Upper was consistently the most saline in the transect, ranging
2 from 2,490 $\mu\text{S}/\text{cm}$ in April 2011 to 3,250 $\mu\text{S}/\text{cm}$ in August 2011. Groundwater at Kelly Creek
3 was generally more saline than Tambo Upper, with EC values ranging from 2,000 to 2,777
4 $\mu\text{S}/\text{cm}$ over the study period. Groundwater EC values were less variable at Kelly Creek and
5 did not generally increase or decrease with proximity to the Tambo River.

6 **3.3 Stable isotopes**

7 $\delta^{18}\text{O}$ and $\delta^2\text{H}$ values of groundwater and river water generally plot close to the both local
8 (LMWL) and global meteoric water lines (GMWL); however river water at Kelly Creek from
9 February 2011 plots to the right of the GMWL (Fig. 4). Groundwater and surface water
10 samples from February 2011 define an array with a slightly shallower trend than the LMWL,
11 while those from March 2012 define an array with a steeper trend than the LMWL. $\delta^{18}\text{O}$
12 values at Bruthen ranged from -4.3 to -7.5 ‰ and were generally higher closer to the river at
13 B1 (average = -4.8 ± 0.4 ‰) than further from the river at B2 and B3 (average = -5.3 ± 2.2 ‰).
14 Stable isotope values were less variable at Tambo Upper with $\delta^{18}\text{O}$ values ranging from -5.3
15 to -6.3 ‰. Groundwater at TU3D, TU1 and TU2 has lower $\delta^{18}\text{O}$ values (average = -6.0 ± 0.2
16 ‰) than at TU4 and TU5 (average = -5.6 ± 0.2 ‰). Shallow groundwater at Kelly Creek
17 showed little variability, with $\delta^{18}\text{O}$ values ranging from -5.3 to -5.8 ‰ during the study. The
18 deeper groundwater at Kelly Creek had slightly lower $\delta^{18}\text{O}$ values (average = -5.9 ± 0.5 ‰)
19 than the shallow groundwater. River water had lower $\delta^{18}\text{O}$ values than groundwater during all
20 sampling periods except February 2011. During this period, $\delta^{18}\text{O}$ values of river water
21 increased from -5.7 ‰ at Bruthen to -3.4 ‰ at Kelly Creek. There was less variation in the
22 river water at other times during the study, with $\delta^{18}\text{O}$ values ranging from -7.9 to -7.5 ‰.

23 **3.4 ^3H and ^{14}C**

24 ^3H and ^{14}C activities in April 2011 were the highest in groundwater from Bruthen, ranging
25 from 2.7 to 2.8 tritium units and 98.0 to 99.3 pMC, respectively. ^3H activities were higher in
26 groundwater further from the river at Tambo Upper at TU4 and TU5 (^3H activities 1.6 and 1.2
27 tritium units, respectively) compared to groundwater closer to the river at TU1 and TU2 (^3H
28 activities 0.40 and 0.36 tritium units, respectively). ^3H activities in the deeper groundwater at
29 TU3D were below detection. ^{14}C activities show a similar variation, with higher activities at
30 TU4 and TU5 (94.5 and 79.2 pMC) compared to groundwater at TU1 and TU2 (35.4 and 38.0
31 pMC). Deeper groundwater at TU3D had lower ^{14}C activities (10.6 pMC). ^3H activities in

1 groundwater at Kelly Creek decreased from 0.51 tritium units at KC4 to 0.40 and 0.36 tritium
2 units at KC1 and KC2, respectively. ^{14}C activities follow a similar trend, decreasing from 84.2
3 pMC at KC4 to 80.4 pMC at KC1.

4 **3.5 Major ions**

5 Despite sampling groundwater from similar aquifers, there are considerable differences in the
6 geochemistry of groundwater from the three locations.

7 Groundwater from Bruthen is $\text{HCO}_3\text{-Ca-Na}$ type (Fig. 5). NO_3 , Br and Cl are the most
8 variable anions over time, with relative standard deviations of 120, 91 and 82% of the mean
9 concentrations, respectively. Na is the most variable cation at the site with relative standard
10 deviations of 53% of the mean concentrations. The concentration of most major cations at
11 Bruthen decrease with increasing Cl concentrations, however K has a weak positive
12 correlation with Cl (Fig. 6). Molar Na:Cl ratios at Bruthen generally range from 2 to 4 during
13 periods of lower rainfall in the catchment (February 2011, April 2011 and November 2011),
14 and are generally below 1 during periods of increased rainfall (August 2011 and March 2012).
15 Molar Cl:Br ratios at Bruthen increase from 140 to over 1,000 with increasing Cl
16 concentrations (Fig. 6).

17 Groundwater from Tambo Upper is Cl-Na-Ca type (Fig. 5). NO_3 is the most temporally
18 variable anion at Tambo Upper, with a relative standard deviation of 97% of the mean. In
19 contrast, temporal variations in cations are relatively low at Tambo Upper, with relative
20 standard deviations of between 15 and 21% of the mean values. At Tambo Upper Na and K
21 concentrations increase and Ca and Mg concentrations decrease with increasing Cl
22 concentrations (Fig. 8). Groundwater further from the river at Tambo Upper (TU4 and TU5)
23 has Cl concentrations below 10 mmol/L, K concentrations below 0.2 mmol/L and Na
24 concentrations below 7 mmol/L (Table 1). Deeper groundwater from Tambo Upper (TU3D)
25 has Cl concentrations greater than 15 mmol/L, K concentrations greater than 0.8 mmol/L and
26 Na concentrations greater than 16 mmol/L. Groundwater closer to the river at Tambo Upper
27 (TU1 and TU2) has concentrations of Na, K, Mg and Ca that are intermediate between those
28 of groundwater at TU3D and groundwater at TU4 and TU5.

29 Shallow groundwater at Kelly Creek is Cl-Ca-Na type. SO_4 is the most variable anion at Kelly
30 Creek, with relative standard deviations of between 108 and 206% of mean values at each
31 bore. Na is the most variable cation at Kelly Creek, with relative standard deviations of

1 between 22 and 43% of mean values. At Kelly Creek, shallow groundwater has Cl
2 concentrations that range from 11.6 to 20.1 mmol/L and Ca concentrations that range from
3 3.1 to 8.5 mmol/L (Fig. 7). Ca, Na, K, and Mg concentrations generally increase with Cl
4 concentrations. Deeper groundwater from Kelly Creek shows similar trends in major ion
5 concentrations to shallower groundwater; however, the relative proportion of Na and Mg is
6 higher and the relative proportion of Ca is lower. Molar Cl:Br ratios in groundwater at Kelly
7 Creek increase from ~650 to ~1,000 while Na:Cl ratios decrease from 1.4 to 0.4 as Cl
8 concentrations increase.

9 **4 Discussion**

10 The following section focusses on identifying the source of water stored in the banks of the
11 Tambo River. Through groundwater dating, the prevalence of bank storage is evaluated and
12 patterns in groundwater recharge and flow are identified. These evaluations are further
13 coupled with major ion and stable isotope analysis under changing hydrological conditions, in
14 order to identify processes controlling the chemistry of bank water and the potential impacts
15 to river and groundwater quality.

16 **4.1 Hydrogeochemical processes**

17 Na:Cl ratios were higher in the groundwater at Bruthen and Kelly Creek during periods of
18 lower rainfall compared to periods of higher rainfall (Figs 6 and 7). This suggests that the
19 groundwater present in the banks during periods of low rainfall has longer residence times
20 that facilitate water-rock interaction, specifically the dissolution of Na-bearing minerals such
21 as plagioclase (Edmunds, 2009; Herczeg et al., 2001). The same trend is also apparent in the
22 deeper groundwater at Tambo Upper (Fig. 8). Cl:Br ratios at Bruthen and Kelly Creek
23 increase with increasing Cl concentrations (Figs 6 and 7). As evapotranspiration, which is the
24 dominant process that controls groundwater salinity in southeast Australia (Herczeg et al.,
25 2001; Cartwright et al., 2007, 2010), does not impact Cl:Br ratios, this implies the addition of
26 Cl from an external source. Halite dissolution is a potential Cl source; however, there are no
27 obvious stores of halite in the catchment. An alternative source of Cl is KCl fertilizers that are
28 used locally (Department of Environment and Primary Industries, 2013). K:Cl ratios decrease
29 with increasing Cl concentrations at Bruthen (Fig. 6), which would not be expected for a
30 KCl source; however, K may be removed from waters recharging through the soils by
31 vegetation (e.g., Schachtman and Schroder, 1994) or sorption onto clay minerals such as illite

1 (Griffioen, 2001). In any case, the observation that increased Cl concentrations coincide with
2 increased rainfall suggest that infiltration facilitates the transport of Cl from the land surface
3 and/or the soil profile into shallow groundwater (c.f., Panno et al., 2006).

4 Mixing between deeper groundwater (TU3D) and shallow groundwater (TU4) appears to
5 dominate the chemical variability of groundwater throughout the rest of the transect at Tambo
6 Upper (Fig. 8). The deeper groundwater has elevated Na:Cl and K:Cl ratios, likely to be
7 attributed to greater residence times and Na and K mineral dissolution. Deeper groundwater is
8 also relatively saline (17.11 to 27.03 mmol/L) compared to shallower groundwater at TU4
9 (3.94 to 6.24 mmol/L). Groundwater throughout the rest of the transect contain intermediate
10 Cl concentrations and cation to chloride ratios, as is consistent with mixing between the two
11 end members. It is also apparent that the relative proportion of the deeper groundwater end
12 member in the bank waters at Tambo Upper increases the during wetter periods in August
13 2011 and March 2012, suggesting that hydraulic loading of the deeper, semi-confined aquifer
14 is driving increased flow of deeper groundwater into the overlying alluvial aquifer at these
15 times. As indicated in section 3.5, high variations in NO₃ concentrations were observed both
16 temporally and spatially across the transects. These variations were not systematic with
17 changing hydrological conditions or other major ions, suggesting that perhaps changing redox
18 conditions have impacted the observed NO₃ concentrations. However, as redox conditions
19 were not recorded and multi species analysis of N were not undertaken, these processes
20 remain unresolved.

21 **4.2 Aquifer interactions**

22 The ¹⁴C and ³H activities in groundwater may be predicted from their atmospheric
23 concentrations and groundwater residence times. The activities of these isotopes in the
24 atmosphere were elevated due to nuclear tests that occurred mainly in the 1960s (the so-called
25 “bomb pulse”). For this study, present-day ³H activities are taken to be 3.2 tritium units which
26 is the weighted average rainfall ³H activity for July 2005 to June 2011 in the Melbourne area
27 (Tadros et al., 2014), and we assume that pre-bomb pulse tritium activities were similar to
28 these (Allison and Hughes (1977). For intervening years, the mean weighted average of ³H
29 concentration of precipitation in Melbourne was taken as that of local precipitation with the
30 record extrapolated for years with no data (Cartwright et al., 2013). Unlike ³H, ¹⁴C activities
31 of atmospheric CO₂ were similar in the northern and southern hemispheres (Fontes, 1983).
32 The data of Hau et al. (2013) were used for ¹⁴C activities of precipitation from 1950 to 2011.

1 Pre 1950, ^{14}C activities are assumed to have decreased from 100 pMC in 1905 to 97.5 pMC in
2 1950 due to fossil fuel burning (Suess, 1971).

3 Lumped-parameter models are commonly used to describe groundwater flow in shallow
4 unconfined and semi-confined aquifers (Małozzewski and Zuber, 1991; Małozzewski and
5 Zuber, 1982; Morgenstern, 2010; Zuber et al., 2005). Piston flow models assume that no
6 mixing takes place between recharge and water in the aquifer, and is suitable for settings
7 where dispersion is low. Conversely, exponential flow models assume a vertical stratification
8 of groundwater ages in an aquifer and are suitable for the sampling of fully penetrating wells
9 or surface water bodies fed by aquifers receiving homogeneous recharge. This study uses the
10 exponential piston flow model (EPFM) which combines a portion of piston flow followed by
11 a portion of exponential flow and is appropriate for bores in unconfined to semi-confined
12 aquifers screened below the water table that do not sample the shallowest groundwater that
13 has very short residence times (Morgenstern, 2010; Cartwright and Morgenstern, 2012).

14 For the EPFM the activity of ^3H or ^{14}C at time t (C_t) is given by:

15

$$16 \quad C_t = \int_0^\infty C_i(t - \tau)g(\tau)e^{-\lambda\tau}d\tau \quad (1)$$

17 where C_i is the initial ^3H or ^{14}C activity, λ is the decay constant ($5.63 \times 10^{-3} \text{ yr}^{-1}$ for ^3H , 1.21
18 $\times 10^{-4} \text{ yr}^{-1}$ for ^{14}C), τ is the transit time, and $g(\tau)$ is the system response function. The system
19 response function is given by:

20

$$21 \quad g(\tau) = 0 \quad \text{for } \tau < T(1-f) \quad \text{and} \quad (2a)$$

$$22 \quad g(\tau) = (fT)^{-1} e^{(-\tau/(fT)+1/f-1)} \quad \text{for } \tau > T(1-f) \quad (2b)$$

23

24 where T is the mean residence time and f is the ratio of exponential flow to piston flow for the
25 total flow volume (Zuber et al., 2005). f was estimated at 0.8 for shallow bores neighbouring
26 the Tambo River on the basis of bore depth, screen length and aquifer lithology (c.f.,
27 Cartwright and Morgenstern, 2012; Cartwright et al., 2013).

28 While there are some differences in the estimated groundwater residence times between
29 different types of flow models, the predicted variation in ^{14}C and ^3H activities are similar in

1 all flow models that involve attenuation of the bomb-pulse peak of ^3H and ^{14}C during flow
2 (e.g., as discussed by Cartwright et al., 2013). A similar covariance of ^{14}C and ^3H activities
3 would be obtained using a dispersion model (Zuber et al., 2005) or the renewal rate model of
4 Le Gal La Salle et al. (2001; Fig. 9). The ^{14}C and ^3H activities also constrain mixing within
5 the groundwater system (Le Gal La Salle et al., 2001; Cartwright et al., 2007, 2010, 2013).
6 Mixing between recently-recharged groundwater and older groundwater with low ^{14}C and
7 negligible ^3H activities will displace water compositions to the left of the predicted a ^{14}C vs.
8 ^3H trends. Closed-system calcite dissolution that lowers a ^{14}C but which does not impact ^3H
9 activities produces a similar displacement.

10 The co-variance between ^3H and ^{14}C for groundwater samples is shown in Fig. 9.
11 Groundwater from Bruthen and Kelly Creek has ^3H and ^{14}C activities lie close to the
12 predicted co-variance curves. Groundwater in aquifers dominated by siliclastic sediments
13 typically undergo up to 20% closed system calcite dissolution (Vogel, 1970; Clark and Fritz,
14 1997) and if that is the case for this groundwater, it would explain the slightly lower than
15 expected ^{14}C activities. Regardless, there can be very limited mixing between older and
16 younger groundwater at these localities. By contrast, groundwater from TU1, TU2 and TU5 at
17 Tambo Upper follow a trend consistent with the mixing between younger groundwater in the
18 shallow aquifer (TU4) and older groundwater in the deeper semi-confined aquifer (TU3D)
19 (Fig. 9). The trend indicates increased leakage from the deeper aquifer into the surface aquifer
20 closer to the river at TU1 and TU2. This is consistent with higher groundwater levels and
21 electrical conductivities at TU1 and TU2 (Fig. 3) that would result from increased
22 connectivity with artesian groundwater in the deeper, semi-confined aquifer. This connection
23 may have resulted from erosion of the clay layers closer to the Tambo River during periodic
24 flooding.

25 **4.3 Groundwater residence times and mixing**

26 Groundwater residence times were calculated using the ^3H activities and the EPFM with $f =$
27 0.8. Groundwater from Bruthen has relatively short residence times of 2 to 4 years.
28 Groundwater from Kelly Creek has longer residence times (96 to 100 years), which is
29 consistent with the higher degrees of mineral dissolution at Kelly Creek discussed previously.
30 Groundwater from TU4 at Tambo Upper has an intermediate residence time of 27 years. To
31 assess the sensitivity of these results, f values in this study were varied between 0.6 and 1.0.
32 This results in variations of <0.1 years at Bruthen and <15 years at Kelly Creek. Uncertainties

1 in groundwater age based on the uncertainty of ^3H activities were <1 year at Bruthen (based
2 on an uncertainty of 0.14 tritium units) and <1.5 years at Kelly Creek (based on an uncertainty
3 of 0.04 tritium units). As deeper groundwater from Tambo Upper site is ^3H free, residence
4 times were calculated from ^{14}C activities. Making the assumption of 15% calcite dissolution,
5 age estimates based on Clarke and Fritz (1997, their Eq. (2) pg. 206) are ~17,200 years.

6 The relatively young groundwater residence times from the shallow aquifers implies that
7 groundwater recharge in the area is dominantly local, probably within a few hundreds of
8 meters of the Tambo River. Mean groundwater residence times from the Bruthen bores are
9 similar and within analytical uncertainty, preventing calculation of horizontal flow velocities.
10 Mean groundwater residence times at Kelly Creek increase from 96 years at KC4 to 100 years
11 at KC2. The age of groundwater at KC1 is 99 years and within the analytical uncertainty of
12 groundwater at KC2. Based on these data, groundwater at Kelly Creek has a horizontal flow
13 velocity of between 1.3 and 6.5 m/year towards the river.

14 The ^3H and ^{14}C activities predicted by the mixing between groundwater from TU4 and deeper
15 groundwater are shown in Fig. 10. While it is possible that groundwater from TU4 has
16 already undergone some mixing with deeper groundwater (and C inputs from the aquifer are
17 less than 10% opposed to the 10-20% indicated), this remains difficult to define. As such,
18 mixing estimates at Tambo Upper will be conservative with respect to the input of deeper
19 groundwater. Groundwater from TU1, TU2 and TU5 plot below the mixing trend in Fig. 10.
20 While there are uncertainties in these calculations it is possible that ^3H activities are lower
21 than expected due to the decay of ^3H in shallow groundwater. Exponential piston flow
22 modelling of water at TU1 and TU2 indicates that a residence time of ~20 years would be
23 required to cause the observed deviation in ^3H activities from the mixing trend shown in Fig.
24 10. This suggests a horizontal flow rate of 1.8 ± 0.6 m/year towards the Tambo River at the
25 Tambo Upper transect. This is consistent with shallow groundwater recharge on the
26 floodplains of the Tambo River and groundwater flow towards the river, which is expected
27 given the gaining nature of this section of the river (Unland et al., 2013).

28 **4.4 Implications for groundwater – surface water interaction**

29 The distribution of groundwater residence times does not support increased bank storage in
30 the area immediately (within 10's of meters) neighbouring the Tambo River. If this was so,
31 groundwater closer to the Tambo River would contain a higher proportion of younger water

1 than groundwater further from the river and groundwater ages would decline towards the
2 river. Instead, groundwater ages increase towards the Tambo River at Kelly Creek and Tambo
3 Upper, while groundwater at Bruthen was approximately the same age at 18 m and 6 m
4 distance from the Tambo River.

5 As the ^3H and ^{14}C activities were analysed for groundwater sampled in April 2011, these data
6 can only be used to evaluate bank storage for the hydrological conditions at and immediately
7 prior to sampling. This included a discharge event that increased river height by 0.5 m
8 approximately 2 weeks prior to sampling. As such, these data indicate that an increase in river
9 height of 0.5 m is not large enough to produce bank storage 5 to 10 m distance from of the
10 river for a period greater than 2 weeks. Major ions and stable isotopes were analysed at
11 several times, including after flood events which increased river height by ~5 m. Again there
12 is little evidence of river water infiltrating into the river banks following these events. The
13 curves expected for the mixing between shallow groundwater furthest from the river, deeper
14 groundwater, and river water at each transect with respect to Cl:Br, Na:Cl, and K:Cl ratios are
15 shown in Fig. 11. Data are shown for February 2011 and August 2011 to represent baseflow
16 conditions when bank infiltration is likely to have the least impact on groundwater chemistry
17 and post flood conditions, when bank infiltration is most likely to impact groundwater
18 chemistry.

19 The composition of groundwater from the two bores closest to the Tambo River at each
20 transect are not consistent with the trends expected for mixing between river water and deeper
21 or shallow groundwater further from the river. For example, during both February and
22 August at Bruthen, Cl:Br, Na:Cl and K:Cl ratios from groundwater at B1 and B2 plot to the
23 left the curves expected for mixing between river water and groundwater further from the
24 river at B3 (Fig. 11a and b). This is partly due to the higher Cl concentrations in river water
25 and groundwater from B3 (which range from 0.3 to 0.5 mmol L⁻¹) compared to groundwater
26 from B1 and B2 (which range from 0.1 to 0.35 mmol L⁻¹). If mixing between river water and
27 groundwater at B3 was occurring, Cl concentrations at B1 and B2 would be intermediate
28 between those in the river and at B3.

29 The same is true at Tambo Upper where Cl concentrations in groundwater at TU1 and TU2
30 are higher (~ 15 mmol L⁻¹) than in river water or groundwater further from the river at TU4
31 (<10 mmol L⁻¹) during both baseflow and flood conditions (Fig. 11c and d). The same is true
32 for EC values which are lower in river water and at TU 4 (ranging from 120 to 881 $\mu\text{S cm}^{-1}$)

1 over the study) compared to EC values at TU1 and TU2 which ranged from 1,350 to 2,682 μS
2 cm^{-1} over the study. Again, if mixing between river water and water at TU4 had a significant
3 impact on groundwater chemistry at TU1 and TU2, Cl concentrations and EC values would
4 be expected to be intermediate between the river water and groundwater from TU4. Instead,
5 groundwater in TU1 and TU2 has a geochemistry that is similar to that which would be
6 expected for mixing between TU4 and TU3D (Fig. 11c and d). As asserted in section 4.1,
7 such mixing is implied by the ^3H and ^{14}C data. Similarly, $\delta^2\text{H}$ and $\delta^{18}\text{O}$ values of groundwater
8 close to the Tambo River do not decline after significant flooding, as would be expected for
9 the infiltration of river water with the lower $\delta^2\text{H}$ and $\delta^{18}\text{O}$ values observed during flooding.

10 These observations indicate that river water penetrates <5 m into the banks during flooding
11 suggesting limited bank infiltration. The absence of significant bank infiltration is consistent
12 with results of Vekerdy and Meijerink (1998) and Wett et al. (2002) who found bank
13 infiltration to be minimal in confined and semi-confined aquifers, where pressure loading
14 from the flood wave propagated rapidly into the neighbouring aquifers, limiting bank
15 infiltration. While most bores near the Tambo River are screened in the alluvial aquifer which
16 is unconfined, leakage of the underlying semi-confined aquifer into the alluvial aquifer does
17 occur (Fig. 9). This upward leakage occurs close to the river where erosion has removed some
18 of the confining layers (Rinaldi and Darby, 2007). It is possible that bank storage is occurring,
19 but that the gaining nature of the Tambo River near these transects is driving the return of
20 bank water back into the river before sampling has taken place (Fig. 12). If this is the case, the
21 storage period (~1 week after the flood peak) is significantly shorter than the several weeks to
22 months predicted by modelling (e.g. Cooper, 1963; Doble et al., 2012; McCallum et al., 2010;
23 Whiting and Pomeranets, 1997). It is likely that the apparent absence of bank storage near the
24 Tambo River is being driven by a combination of such factors, including: (1) upward flow of
25 groundwater from the deeper aquifer into the river bank due to pressure loading on the
26 floodplain; (2) the return of any bank waters back into the river under strongly gaining
27 conditions; and (3) the high TDS water from the confined system masking the chemical
28 impact of infiltrating river water.

29 **5 Conclusions**

30 This study was able to determine the absence of significant bank storage near the Tambo
31 River by dating near river groundwater and characterising its major ion chemistry before and
32 after flooding. The combined use of ^3H and ^{14}C activities indicate that recharge in the alluvial

1 aquifer is dominantly local (with 100's of meters of the Tambo River). The covariance
2 between ^3H and ^{14}C activities show that mixing between relatively old groundwater from a
3 deeper semi-confined aquifer, and younger groundwater from the unconfined alluvial aquifer
4 is occurring in parts of the Tambo River bank. It is further shown that by coupling ^3H and ^{14}C
5 to define a mixing trend, deviations in the activity of ^3H from the trend can be used to
6 estimate the likely age of groundwater along its flow path. Na:Cl ratios >1 in groundwater
7 sampled during baseflow conditions and in older groundwater from the area indicate the
8 dissolution of Na bearing minerals and is consistent with the weathering of silicic sands in the
9 aquifer. Increasing Cl:Br ratios and increasing Cl concentrations during periods of increased
10 rainfall indicate an input of Cl, as is consistent with the mobilisation of Cl accumulated in the
11 soil profile through the use of fertilizers. Increasing groundwater age with proximity to the
12 Tambo River is consistent with the gaining nature of the Tambo River, but does not suggest
13 that exchange between groundwater and surface water increases with increasing proximity to
14 the river. Major ions, $\delta^2\text{H}$ and $\delta^{18}\text{O}$ values support this and do not show trends consistent with
15 an increased input of river water to the groundwater closer to the river. These results suggest
16 that either the strongly gaining nature of the Tambo River at the study locations is preventing
17 significant lateral infiltration of river water into the bank, or that the rapid propagation of
18 pressure into the underlying semi-confined aquifer, followed by leakage into the above
19 unconfined aquifer is preventing significant bank infiltration.

20 These results are indicative of the highly complex nature of groundwater and surface water
21 processes that may be occurring within river banks and illustrates that while models can
22 significantly help in conceptualising our understanding of groundwater-surface water
23 interactions, field studies can offer complementary information that may otherwise be
24 overlooked. In this setting, the assumption of typical bank storage processes and the use of a
25 regional groundwater end member during mass balance calculations would lead to poor
26 groundwater flux estimates. In this context, the importance of nested piezometers near rivers
27 to more accurately characterise near river groundwater processes has been highlighted. Even
28 the monitoring of relatively inexpensive parameters such as groundwater levels, EC and Cl
29 can provide significant information to researchers and groundwater managers when
30 conducting such studies.

31

32

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7

1 **References**

- 2 Allison, G.B. and Hughes, M.W.: The history of tritium fallout in southern Australia inferred
3 from rainfall and wine samples. *Earth and Planetary Science Letters* 36, 334-340, 1977.
- 4 Atkinson, A.P., Cartwright, I., Gilfedder, B.S., Hofmann, H., Unland, N.P., Cendón, D.I. and
5 Chisari, R.: A multi-tracer approach to quantifying groundwater inflows to an upland
6 river; assessing the influence of variable groundwater chemistry. *Hydrological
7 Processes*. DOI: 10.1002/hyp.10122, 2013.
- 8 Birch, W.D.: *The Geology of Victoria*. Geological Society of Australia (Victorian Division),
9 Sydney, 2003.
- 10 Boulton, A.J.: Stream ecology and surface-hyporheic hydrologic exchange: Implications,
11 techniques and limitations. *Marine and Freshwater Research*, 44, 553-564, 1993.
- 12 Boulton, A.J.: Chances and challenges in the conservation of groundwaters and their
13 dependent ecosystems. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 15,
14 319-323, 2005.
- 15 Bourg, A.C.M. and Bertin, C.: Biogeochemical processes during the infiltration of river water
16 into an alluvial aquifer. *Environmental Science & Technology*, 27, 661-666, 1993.
- 17 Bureau of Meteorology: Commonwealth of Australia, Bureau of Meteorology.
18 <http://www.bom.gov.au> (last access: 28 August 2013), 2013.
- 19 Cartwright, I., Fifield, L.K., and Morgenstern, U.: Using ^3H and ^{14}C to constrain the degree of
20 closed-system dissolution of calcite in groundwater. *Applied Geochemistry*, 32, 118-
21 128, 2013.
- 22 Cartwright, I., Gilfedder, B. and Hofmann, H.: Contrasts between estimates of baseflow help
23 discern multiple sources of water contributing to rivers. *Hydrology and Earth System
24 Sciences*, 18, 15-30, 2014.
- 25 Cartwright, I. and Morgenstern, U.: Constraining groundwater recharge and the rate of
26 geochemical processes using tritium and major ion geochemistry: Ovens catchment,
27 southeast Australia. *Journal of Hydrology*, 475, 137-149, 2012.
- 28 Cartwright, I., Weaver, T.R., Simmons, C.T., Fifield, L. K., Lawrence, C. R. and Chisari, R.:
29 Physical hydrogeology and environmental isotopes to constrain the age, origins, and

1 stability of a low-salinity groundwater lens formed by periodic river recharge: Murray
2 Basin, Australia. *Journal of Hydrology*, 380, 203-221, 2010.

3 Cartwright, I., Weaver, T.R., Stone, D. and Reid, M.: Constraining modern and historical
4 recharge from bore hydrographs, ^3H , ^{14}C , and chloride concentrations: Applications to
5 dual-porosity aquifers in dryland salinity areas, Murray Basin, Australia. *Journal of*
6 *Hydrology*, 332, 69-92, 2007.

7 Cendón D.I., Larsen J.R., Jones B.G., Nanson G.C., Rickleman D., Hankin S.I., Pueyo J.J.
8 and Maroulis J.: Freshwater recharge into a shallow saline groundwater system, Cooper
9 Creek floodplain, Queensland, Australia. *Journal of Hydrology* 392, 150-163, 2010.

10 Cey, E.E., Rudolph, D.L., Aravena, R. and Parkin, G.: Role of the riparian zone in controlling
11 the distribution and fate of agricultural nitrogen near a small stream in southern Ontario.
12 *Journal of Contaminant Hydrology*, 37, 45-67, 1999.

13 Chen, X., Chen, D.Y. and Chen, X.-H.: Simulation of baseflow accounting for the effect of
14 bank storage and its implication in baseflow separation. *Journal of Hydrology*, 327,
15 539-549, 2006.

16 Chen, X. and Chen, X.: Stream water infiltration, bank storage, and storage zone changes due
17 to stream-stage fluctuations. *Journal of Hydrology*, 280, 246-264, 2003.

18 Coplen, T.B.: Normalisation of oxygen and hydrogen isotope data. *Chemical Geology:*
19 *Isotope Geoscience section*, 72, 293-297, 1988.

20 Cooper, H.H. and Rorabaugh, M.I.: Ground-water movements and bank storage due to flood
21 stages in surface streams, United States Department of the Interior, Geological Survey
22 Water-Supply Paper, Washington, 1963.

23 Clark, I.D. and Fritz, P.: *Environmental Isotopes in Hydrogeology*. Lewis, New York, USA,
24 206 pp., 1997.

25 Doble, R., Brunner, P., McCallum, J. and Cook, P.G.: An analysis of river bank slope and
26 unsaturated flow effects on bank storage. *Ground Water*, 50, 77-86, 2012.

27 Department of Environment and Primary Industries: Victorian Resources Online,
28 Groundwater. [http://vro.dpi.vic.gov.au/dpi/vro/vrosite.nsf/pages/water-gw-quality-](http://vro.dpi.vic.gov.au/dpi/vro/vrosite.nsf/pages/water-gw-quality-quantity)
29 [quantity](http://vro.dpi.vic.gov.au/dpi/vro/vrosite.nsf/pages/water-gw-quality-quantity) (last access: 5 July 2013), 2013.

- 1 Department of Agriculture, Fisheries and Forestry: Gippsland Catchment.
2 <http://adl.brs.gov.au/water2010/pdf/catchment2230summary.pdf> (last access: 19
3 November 2012), 2006.
- 4 Edmunds, W.M.: Geochemistry's vital contribution to solving water resource problems.
5 *Applied Geochemistry*, 24, 1058-14073, 2009.
- 6 Fetter, C. W. 1994. *Applied hydrogeology*. Merrill, New Jersey.
- 7 Fink, D., Hotchkis, M., Hua, Q., Jacobsen, G., Smith, A.M., Zoppi, U., Child, D., Mifsud, C.,
8 Vender Gaast, H., Williams, A. and Williams, M.: The Antares AMS facility at
9 ANSTO. *Nuclear Instruments and Methods in Physics Research*, 223, 109-115, 2004.
- 10 Fontes, J.-Ch.: Dating of Groundwater. *Guidebook on Nuclear Techniques in Hydrology*.
11 *Technical Reports Series No. 91*, IAEA, Vienna. pp. 285–317, 1983.
- 12 Fukada, T., Hiscock, K.M., Dennis, P.F. and Grischek, T.: A dual isotope approach to identify
13 denitrification in groundwater at a river-bank infiltration site. *Water Research*, 37,
14 3070-3078, 2003.
- 15 Griffioen J.: Potassium adsorption ratios as an indicator for the fate of agricultural potassium
16 in groundwater. *Journal of Hydrology*, 254, 244–254, 2001.
- 17 Herczeg, A.L., Dogramaci, S.S. and Leany, F.W.: Origin of dissolved salts in a large, semi-
18 arid groundwater system: Murray Basin, Australia. *Marine and Freshwater Research*,
19 52, 41-52, 2001.
- 20 Hiscock, K.M. and Grischek, T.: Attenuation of groundwater pollution by bank filtration.
21 *Journal of Hydrology*, 266, 139-144, 2002.
- 22 Hocking, J.B.: Definition and revision of Tertiary stratigraphic units, onshore Gippsland
23 Basin. *Geological Survey of Victoria*, 1976.
- 24 Hofmann, H. and Cartwright, I.: Using hydrogeochemistry to understand inter-aquifer mixing
25 in the on-shore part of the Gippsland Basin, southeast Australia. *Applied Geochemistry*,
26 33, 84-103, 2013.
- 27 Hughes, C. E. and Crawford, J.: A new precipitation weighted method for determining the
28 meteoric water line for hydrological applications demonstrated using Australian and
29 global GNIP data. *Journal of Hydrology*, 464–465, 344-351, 2012.

- 1 Hua, Q., Barbetti, M. and Rakowski, A.Z.: Atmospheric Radiocarbon for the period 1950–
2 2010. *Radiocarbon*, 55, 2059-2072, 2013.
- 3 Lambs, L.: Interactions between groundwater and surface water at river banks and the
4 confluence of rivers. *Journal of Hydrology*, 288, 312-326, 2004.
- 5 Lamontagne, S., Leaney, F.W. and Herczeg, A.L.: Groundwater-surface water interactions in
6 a large semi-arid floodplain: implications for salinity management. *Hydrological*
7 *Processes*, 19, 3063-3080, 2005.
- 8 Lamontagne, S., Taylor, A.R., Cook, P.G. and Barrett, C.: Interconnection of Groundwater
9 Systems – River Losses from Losing/Disconnected Streams. Gwydir River Site Report.
10 CSIRO: Water for a Healthy Country National Research Flagship.
11 [http://www.clw.csiro.au/publications/waterforahealthycountry/2011/wfhc-](http://www.clw.csiro.au/publications/waterforahealthycountry/2011/wfhc-interconnection-Gwydir-River.pdf)
12 [interconnection-Gwydir-River.pdf](http://www.clw.csiro.au/publications/waterforahealthycountry/2011/wfhc-interconnection-Gwydir-River.pdf) 2011.
- 13 Le Gal La Salle, C., Marlin, C., Leduc, C., Taupin J.D., Massault, M. and Favreau, G.:
14 Renewal rate estimation of groundwater based on radioactive tracers (^3H , ^{14}C) in an
15 unconfined aquifer in a semi-arid area, Iullemeden Basin, Niger. *Journal of Hydrology*,
16 254, 145-156, 2001.
- 17 Leonard, J.G.: An overview of Victoria's groundwater resources. Background report S3, short
18 term planning guidelines. Drought management plan for Victoria's water resources.
19 Department of Water Resources, Melbourne, Victoria, 1992.
- 20 Małoszewski, P. and Zuber, A.: Determining the turnover time of groundwater systems with
21 the aid of environmental tracers: 1. Models and their applicability. *Journal of*
22 *Hydrology*, 57, 207-231, 1982.
- 23 Maloszewski, P. and Zuber, A.: Influence of Matrix Diffusion and Exchange Reactions on
24 Radiocarbon Ages in Fissured Carbonate Aquifers. *Water Resources Research*, 27,
25 1937-1945, 1991.
- 26 McCallum, J.L., Cook, P.G., Brunner, P. and Berhane, D.: Solute dynamics during bank
27 storage flows and implications for chemical base flow separation. *Water Resources*
28 *Research*, 46, DOI: 10.1029/2009WR008539, 2010.

- 1 Morgenstern, U., Stewart, M.K. and Stenger, R.: Dating of streamwater using tritium in a
2 post-bomb world: continuous variation of mean transit time with streamflow.
3 Hydrology and Earth System Sciences, 14, 2289-2301, 2010.
- 4 Morgenstern, U. and Taylor, C.B.: Ultra low-level tritium measurement using electrolytic
5 enrichment and LSC. *Isotopes in Environment and Health*, 45, 96–117, 2009.
- 6 Panno, S.V., Hackley, K.C., Hwang, H.H., Greenberg, S.E., Krapac, I.G., Landsberger, S. and
7 O’Kelly, D.J.: Source identification of sodium and chloride in natural waters:
8 Preliminary results. *Ground Water*, 44, 176-187, 2006.
- 9 Pinder, G.F. and Sauer, S.P.: Numerical simulation of flood wave modification due to bank
10 storage effects. *Water Resources Research*, 7, 63-70, 1971.
- 11 Rinaldi, M. and Darby, S.E.: 9 Modelling river-bank-erosion processes and mass failure
12 mechanisms: progress towards fully coupled simulations, 11, 213-239, 2007.
- 13 Schachtman, D.P. and Schroeder, J.I.: Structure and transport mechanism of a high-affinity
14 potassium uptake transporter from higher plants. *Nature*, 370, 655-658, 1994.
- 15 Singh, K.P.: Some Factors Affecting Baseflow. *Water Resources Research*, 4, 985-999, 1968.
- 16 Southern Rural Water. Gippsland Groundwater Atlas. <http://www.srw.com.au/gippsland> (last
17 access: 25 June 2014), 2013
- 18 Stuiver, M. and Polach, H.A.: Reporting of ¹⁴C data. *Radiocarbon*, 19, 355-363, 1977.
- 19 Suess, H.: Climatic changes and atmospheric radiocarbon. *Palaeogeography
20 Palaeoclimatology Palaeoecology* 10, 199–202, 1971.
- 21 Tadros, C.V., Hughes, C.E., Crawford, J.C., Hollins, S.E. and Chisari, R.: Tritium in
22 Australian precipitation: a 50 year record. *Journal of Hydrology*, 513, 262-273, 2014.
- 23 Tóth, J. A theoretical analysis of groundwater flow in small drainage basins. *Journal of
24 Geophysical Research*, 68, 4795-4812, 1963.
- 25 Unland, N.P., Cartwright, I., Andersen, M.S, Rau, G.C., Reed, J., Gilfedder, B.S., Atkinson,
26 A.P. and Hofmann, H.: Investigating the spatio-temporal variability in groundwater and
27 surface water interactions: a multi-technique approach. *Hydrology and Earth System
28 Sciences*, 17, 3437-3453, 2013.

- 1 Vekerdy, Z. and Meijerink, A.M.J.: Statistical and analytical study of the propagation of
2 flood-induced groundwater rise in an alluvial aquifer. *Journal of Hydrology*, 205, 112-
3 125, 1998.
- 4 Vogel, J.C.: Groningen radiocarbon dates IX. *Radiocarbon* 12, 444–471, 1970.
- 5 Water Resources Data Warehouse:
6 [http: //www.vicwaterdata.net/vicwaterdata/home.aspx](http://www.vicwaterdata.net/vicwaterdata/home.aspx) (last access: 28 August 2013),
7 2013.
- 8 Wett, B., Jarosch, H. and Ingerle, K.: Flood induced infiltration affecting a bank filtrate well
9 at the River Enns, Austria. *Journal of Hydrology*, 266, 222-234, 2002.
- 10 Whiting, P.J. and Pomeranets, M.: A numerical study of bank storage and its contribution to
11 streamflow. *Journal of Hydrology*, 202, 121-136, 1997.
- 12 Winter, T.C., Harvey, J.W., Franke, O.L. and Alley, W.M.: Ground water and surface water a
13 single resource, U.S. Geological Survey Circular, 1998.
- 14 Woessner, W.W.: Stream and fluvial plain ground water interactions: Rescaling
15 hydrogeologic thought. *Ground Water*, 38, 423-429, 2000.
- 16 Zuber, A., Witczak, S., Rozaski, K., Iiwka, I., Opoka, M., Mochalski, P., Kuc, T.,
17 Karlikowska, J., Kania, J., Jackowicz-Korczynski, and M., Duliski, M.: Groundwater
18 dating with ^3H and SF_6 in relation to mixing patterns, transport modelling and
19 hydrochemistry. *Hydrological Processes*, 19, 2247-2275, 2005.

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Site	Date	EC μS/cm	Level m	Dist. m	F mg/L	Cl mg/L	Br mg/L	NO ₃ mg/L	SO ₄ mg/L	HCO ₃ mg/L	Na mg/L	K mg/L	Ca mg/L	Mg mg/L	O δ ¹⁸ O	H δ ² H	³ H TU	¹⁴ C pMC	δ ¹³ C ‰	MGRT years
R	Feb-11	121	7.84	0.00	0.08	9.63	0.03	0.24	1.94	<i>46.4</i>	7.65	1.38	7.77	3.95	-5.69	-39.1				
1	Feb-11	146	7.87	5.56	0.04	9.60	0.05	0.04	6.05	<i>50.9</i>	12.1	1.65	8.33	3.05	-4.97	-33.6				
2	Feb-11	191	7.88	17.6	0.06	5.61	0.02	0.05	23.5	<i>41.3</i>	7.16	3.00	10.4	5.14	-7.53	-47.9				
3	Feb-11	261	7.90	18.3	0.12	13.3	0.07	0.02	11.5	<i>85.0</i>	19.2	2.52	10.3	7.33	-5.42	-36.5				
R	Apr-11	109	7.61	0.00	0.05	3.63	0.01	0.11	0.90											
1	Apr-11	200	7.66	5.56	0.04	6.75	0.10	0.37	5.83	<i>67.1</i>	16.9	2.03	12.2	4.09	-5.37	-34.7	2.65	98.04	-13.9	3.4
2	Apr-11			17.6	0.15	10.3	0.17	0.30	11.4	<i>183</i>	27.3	2.68	8.47	7.31	-5.93	-37.4	2.84	99.33	-15.4	2.2
R	Aug-11	145	8.81	0.00	0.04	17.6	0.02	2.17	5.02	<i>38.7</i>	11.7	1.68	6.44	4.85	-7.61	-46.6				
1	Aug-11	179	8.84	5.56	0.06	5.66	0.03	0.20	5.36	<i>52.5</i>	2.52	2.44	11.3	4.91	-4.34	-24.7				
2	Aug-11	173	8.84	17.6	0.12	12.7	0.08	0.03	36.2		BD	2.51	11.5	8.93	-5.38	-29.2				
3	Aug-11	293	8.89	18.3	0.06	18.8	0.04	1.48	30.2		BD	5.28	10.4	6.35	-6.66	-36.5				
1	Nov-11	229	7.91	5.56	0.07	11.1	0.05	0.00	6.07	<i>75.1</i>	15.2	2.10	10.4	5.37	-4.81	-31.6				
2	Nov-11	215	7.88	17.6	0.08	14.6	0.05	2.72	10.1	<i>49.5</i>	13.7	3.91	8.20	4.58	-6.10	-34.6				
3	Nov-11	607	7.94	18.3	0.05	78.2	0.21	0.08	12.1	<i>63.1</i>	33.1	3.32	15.8	14.4	-7.40	-44.4				
R	Mar-12	156	7.47	0.00	0.09	17.5	0.03	0.77	4.12	<i>45.9</i>	11.1	1.53	8.17	5.08	-7.56	-46.2				
1	Mar-12	293	8.36	5.56	0.05	17.8	0.06	0.03	16.5	<i>62.1</i>	9.41	1.72	17.8	6.38	-5.13	-28.7				
2	Mar-12	136	8.36	17.6							6.87	2.43	4.91	2.27						
3	Mar-12	287	8.38	18.3	0.16	23.4	0.07	0.00	22.0	<i>44.7</i>	19.1	1.66	8.43	6.87	-4.67	-22.2				

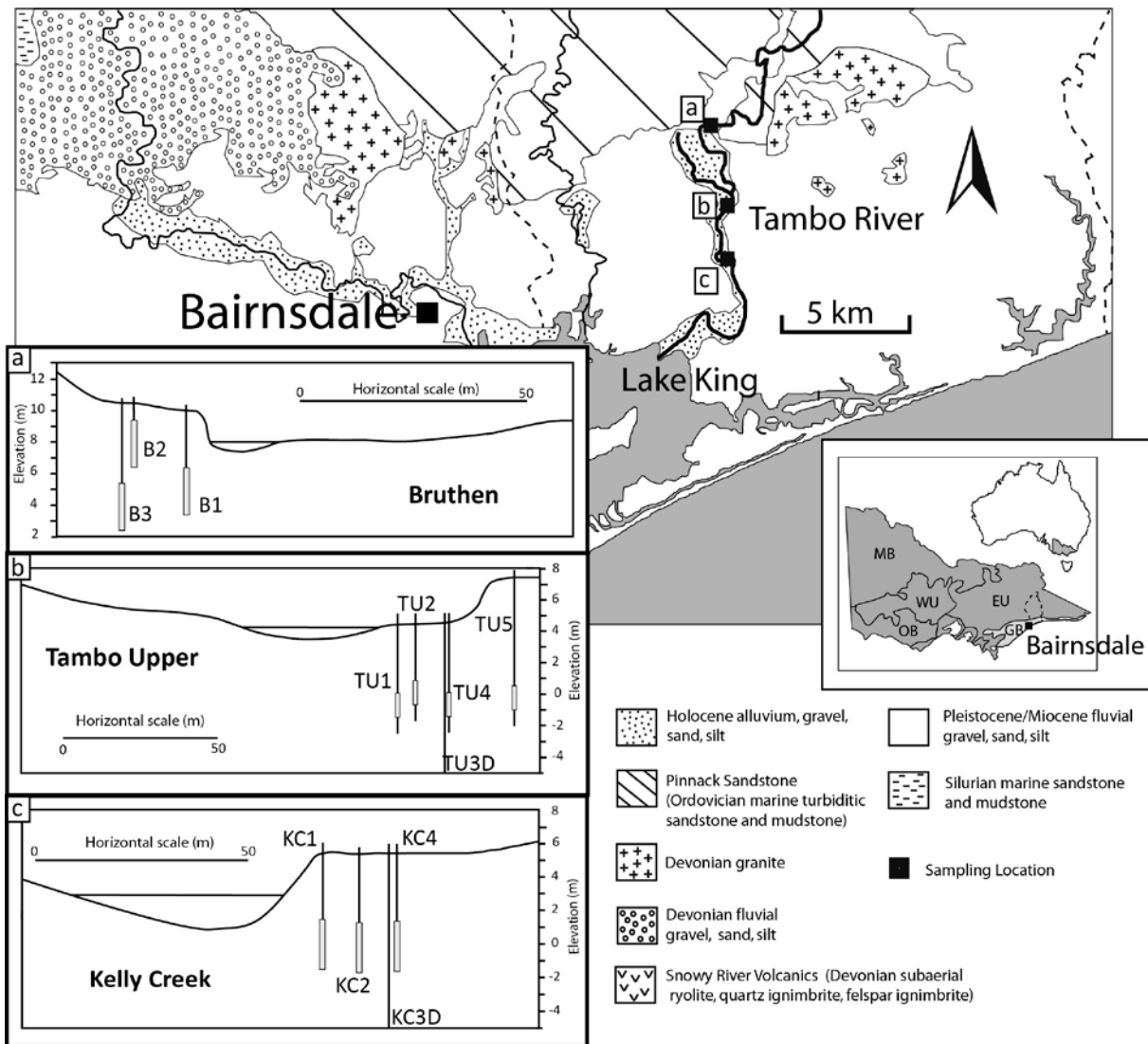
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2 Table 1. Summary data for Bruthen transect. HCO₃ data in italics = calculated via charge balance. MGRT = mean groundwater residence time
3 as calculated via exponential piston flow modelling (see section 4.3).
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Site	Date	EC μS/cm	Level m	Dist. m	F mg/L	Cl mg/L	Br mg/L	NO ₃ mg/L	SO ₄ mg/L	HCO ₃ mg/L	Na mg/L	K mg/L	Ca mg/L	Mg mg/L	O δ ¹⁸ O	H δ ² H	³ H TU	¹⁴ C pMC	δ ¹³ C ‰	MGRT years
R	Feb-11	120	3.58	0.00	0.08	9.78	0.04	0.11	2.04	<i>48.1</i>	7.83	1.74	8.24	3.86	-5.63	-38.3				
1	Feb-11	2,395	3.90	8.82	0.14	572	1.88	0.33	25.2	<i>212</i>	279	23.0	99.6	30.3	-6.06	-38.4				
2	Feb-11	2,155	3.86	15.03	0.12	555	1.77	1.58	14.7	<i>157</i>	280	19.2	62.4	33.9	-5.97	-39.6				
3D	Feb-11	2,656	4.85	22.34	0.36	607	1.98	2.54	22.8	<i>316</i>	374	35.4	64.6	29.8	-6.32	-39.1				
4	Feb-11	764	3.78	23.73	0.12	176	0.52	0.07	21.3	<i>60.7</i>	72.8	4.65	33.9	17.7	-5.43	-37.3				
5	Feb-11	1,023	3.89	37.87											-5.69	-38.8				
R	Apr-11	112	3.43	0.00	0.04		0.01	0.08	0.92		16.0	4.34	10.0	15.1	-7.88	-61.9				
1	Apr-11	2,210	3.38	8.82	0.14	460	1.43	0.22	36.8	<i>291</i>	280	26.9	111	40.7	-5.98	-37.0	0.40	38.0	-8.7	99/78
2	Apr-11	2,180	3.36	15.03	0.11	533	1.76	1.24	25.9	<i>272</i>	333	23.1	86.8	52.8	-5.98	-38.2	0.36	35.4	-7.2	100/80
3D	Apr-11	2,488	4.69	22.34	0.34	534	1.72	0.33	10.9	<i>439</i>	468	33.8	71.8	41.2	-6.18	-38.3	<0.03	10.6	-5.7	17,200
4	Apr-11	717	3.30	23.73	0.11	140	0.42	0.47	15.9	<i>100</i>	98.1	4.42	37.1	21.2	-5.35	-34.9	1.55	94.5	-10	27
5	Apr-11	1,043	3.39	37.87	0.11	203	0.61	0.07	50.3	<i>80.5</i>	122	7.10	64.9	20.7	-5.74	-36.9	1.21	79.2	-9.3	68/52
R	Aug-11	148	4.74	0.00	0.06	18.1	0.02	2.46	4.21	<i>36.5</i>	11.6	1.85	5.87	4.78	-7.66	-45.2				
1	Aug-11	2,682	4.77	8.82	0.11	599	1.63	0.53	12.6	<i>406</i>	339	27.3	105	38.2	-6.01	-32.9				
2	Aug-11	2,207	4.75	15.03	0.19	501	1.42	0.27	22.2	<i>279</i>	280	21.3	78.0	31.7	-5.92	-31.7				
3D	Aug-11	3,250	5.04	22.34	0.26	753	2.09	0.11	34.2	<i>799</i>	600	48.4	80.9	45.2	-6.03	-33.4				
4	Aug-11	774	4.70	23.73	0.08	207	0.41	0.12	5.59	<i>89.8</i>	98.3	3.44	32.6	17.6	-5.38	-27.7				
5	Aug-11	1,039	4.80	37.87	0.14	227	0.43	0.08	11.6	<i>118</i>	115	3.31	36.5	20.9	-5.34	-29.4				
1	Nov-11	2,018	3.74	8.82	0.30	439	1.42	0.56	49.7	<i>383</i>	303	15.5	79.4	26.5	-6.07	-35.0				
2	Nov-11	2,168	3.70	15.03	0.07	531	1.21	0.12	4.46	<i>294</i>	241	14.2	91.6	54.4	-5.93	-37.9				
3D	Nov-11	2,938	>5.04	22.34	0.19	639	2.28	0.21	46.8	<i>760</i>	554	41.2	63.1	38.7	-6.23	-39.1				
4	Nov-11	864	3.61	23.73	0.18	178	0.56	0.21	1.24	<i>170</i>	98.8	4.00	36.0	20.1	-5.37	-35.2				
5	Nov-11	1,337	3.70	37.87	0.08	276	0.83	0.11	48.3	<i>207</i>	158	6.21	68.4	21.7	-5.84	-38.1				
R	Mar-12	165	3.32	0.00	0.09	19.6	0.04	0.54	4.21	<i>48.5</i>	12.0	1.76	8.69	5.47	-7.64	-46.4				
1	Mar-12	1,350	4.30	8.82	0.07	385	0.93	0.13	33.5	<i>27.7</i>	141	19.4	68.5	24.2	-5.65	-28.3				
2	Mar-12	1,763	4.26	15.03	0.07	544	1.11	0.07	29.7	<i>BD</i>	190	11.7	70.4	35.7	-5.88	-26.6				
3D	Mar-12	3,210	>5.04	22.34	0.24	958	2.46	0.05	41.0	<i>BD</i>	484	35.2	58.7	33.3	-6.12	-40.8				
4	Mar-12	881	4.22	23.73	0.10	221	0.53	0.09	1.45	<i>73.0</i>	88.7	4.48	38.0	19.7	n/a	-34.9				
5	Mar-12	1,320	4.32	37.87	0.09	344	0.76	0.30	40.4	<i>57.6</i>	148	6.38	63.3	21.2	-5.84	-26.7				

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2 Table 2. Summary data for Tambo Upper transect HCO₃ data in italics = calculated via charge balance. MGRT = mean groundwater residence
3 time as calculated via exponential piston flow modelling (see section 4.3). MGRT's at TU1, TU2 and TU5 have been given for the ³H
4 activities calculated by mixing and the measured ³H activity, respectively.

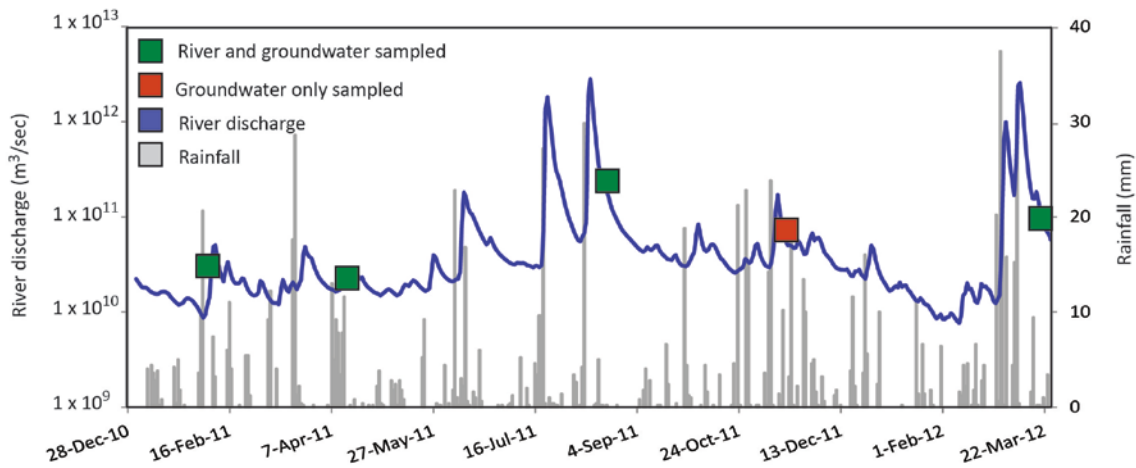
Site	Date	EC μS/cm	Level m	Dist. m	F mg/L	Cl mg/L	Br mg/L	NO ₃ mg/L	SO ₄ mg/L	HCO ₃ mg/L	Na mg/L	K mg/L	Ca mg/L	Mg mg/L	O δ ¹⁸ O	H δ ² H	³ H TU	¹⁴ C pMC	δ ¹³ C ‰	MGRT years
R	Feb-11	18,340	3.05	0	0.38	6,322	22.0	0.30	825	899	3,558	151	182.0	521	-3.38	-24.6				
1	Feb-11	2,004	3.14	7.02	0.3	518	1.54	0.35	0.43	294	178	3.82	178.0	33.6	-5.63	-37.0				
2	Feb-11	2,349	3.17	17.9	0.43	597	1.77	0.77	2.01	374	202	3.93	231.0	32.4	-5.61	-36.9				
3D	Feb-11		3.82	24.9							162	11.4	23.6	9.31	-5.81	-37.8				
4	Feb-11	2,364	3.2	26.8	0.38	637	1.92	0.56	1.14	271	197	2.59	229.0	29.6	-6.68	-41.1				
R	Apr-11	4,210		0	0.21	255	0.93	0.13	29.1	8,690	2683	153	403.7	116	-7.69	-45.5				
1	Apr-11	2,145	3.15	7.02	0.3	474	1.36	0.27	0.28	446	204	3.29	206.9	43.6	-5.58	-35.5	0.40	80.4	-2.8	99
2	Apr-11	2,455	3.07	17.9	0.33	558	1.57	0.43	1.00	488	252	2.83	288.9	42.3	-5.63	-36.2	0.37	83.6	-2.9	100
3D	Apr-11	2,669	3.95	24.9	0.39	413	1.30	0.68	68.9	445	372	17.8	123.3	48.7	-5.58	-36.3				
4	Apr-11	2,099	3.15	26.8	0.57	630	1.83	1.87	0.78	591	313	3.12	340.1	47.8	-6.17	-38.5	0.51	84.2	-3.7	96
R	Aug-11	170		0	0.05	21.6	0.03	2.24	4.82	39.0	13.8	2.38	6.3	5.03	-7.48	-46.0				
1	Aug-11	2,568	3.63	7.02	0.24	590	1.40	0.84	0.38	256	146	4.27	218.9	42.8	-5.27	-29.5				
2	Aug-11	2,777	3.63	17.9	0.26	655	1.44	0.18	1.56	355	135	3.73	286.3	49.8	-5.28	-29.6				
3D	Aug-11	2,438	4.30	24.9	0.23	608	1.30	2.12	61.4	292	218	12.1	185.7	51	-5.55	-29.2				
4	Aug-11	2,717	3.68	26.8	0.26	513	1.26	0.01	0.52	291	73.4	2.28	262.8	35.5	-5.42	-30.3				
1	Nov-11	2,742		7.02	0.45	533	1.70	2.70	0.68	848	286	6.14	249.3	48.7	-5.56	-35.5				
2	Nov-11	2,542	3.36	17.9	0.32	563	1.62	0.13	51.4	711	269	3.26	267.5	42.9	-5.69	-36.9				
3D	Nov-11	2,738	4.33	24.9	0.41	436	1.44	0.04	0.57	998	338	11.0	192.0	50.7	-5.60	-36.3				
4	Nov-11	2,218	3.41	26.8	0.35	532	1.71	2.48	0.54	552	226	2.49	229.5	34.4	-5.48	-36.2				
R	Mar-12	195		0	0.08	33.4	0.09	0.21	7.25	465	82.1	11.9	47.4	30.1	-7.58	-45.5				
1	Mar-12	2,770		7.02	0.3	598	1.48	0.04	2.44	397	218	4.39	215.2	37.9	-5.52	-33.1				
2	Mar-12	2,495		17.9	0.41	605	1.53	0.10	1.11	430	219	3.72	237.8	32.8	-5.45	-32.8				
3D	Mar-12	2,345		24.9	0.37	560	1.45	0.16	0.46	637	292	8.75	199.4	41.6	-5.65	-34.2				
4	Mar-12	2,715		26.8	0.32	713	1.64	0.13	43.4	24.3	189	3.42	217.5	27.6	-5.62	-30.7				

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2 Table 3. Summary data for Kelly Creek transect. HCO₃ data in italics = calculated via charge balance. MGRT = mean groundwater residence
3 time as calculated via exponential piston flow modelling (see section 4.3).



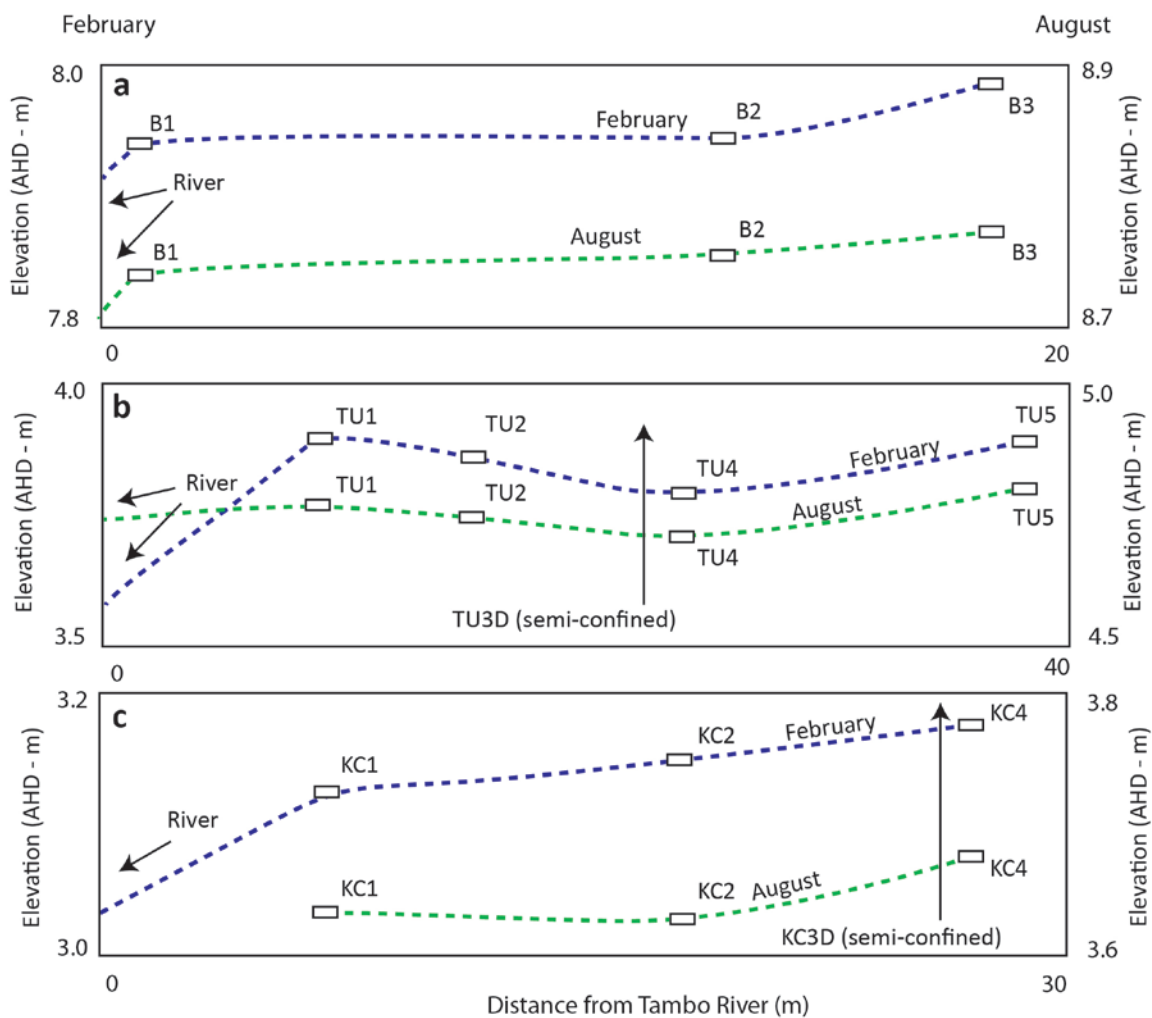
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Figure 1. Location of field area and schematic cross sections of bore transects at Bruthen (a), Tambo Upper (b) and Kelly Creek (c). Screened sections indicated by open boxes. Dashed line = Tambo River basin boundary (transects orientated facing upstream).



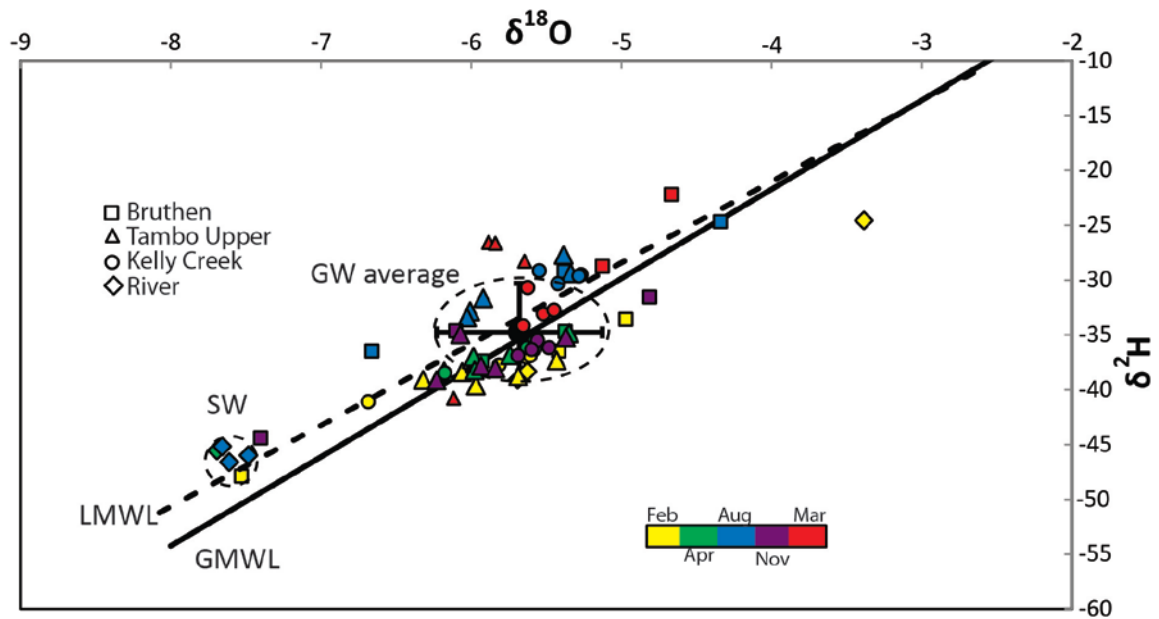
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Figure 2. Surface and groundwater sampling frequency superimposed on Tambo River hydrograph (Battens Landing, station 223209 – Water Resources Data Warehouse, 2013) and rainfall (Bairnsdale Airport, station 85279 – Bureau of Meteorology, 2013).

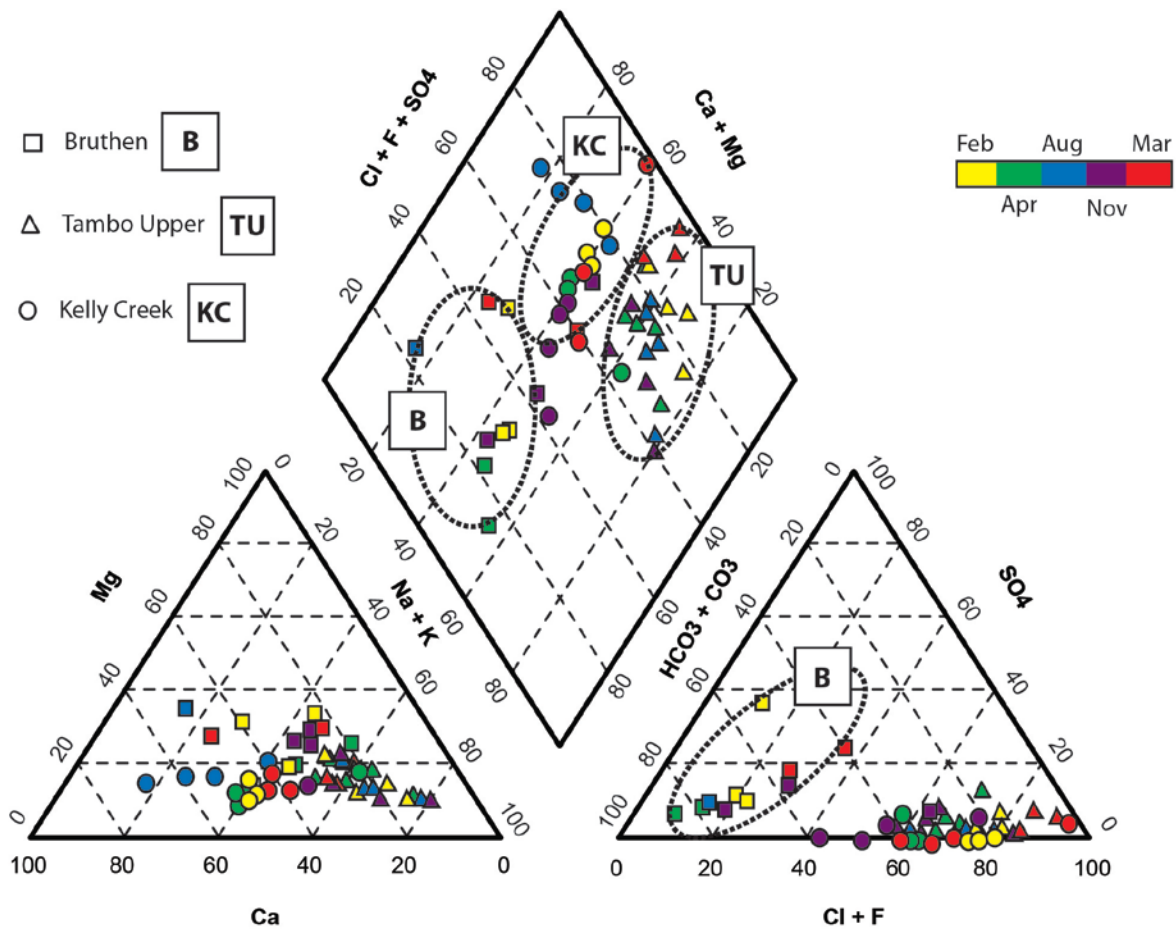


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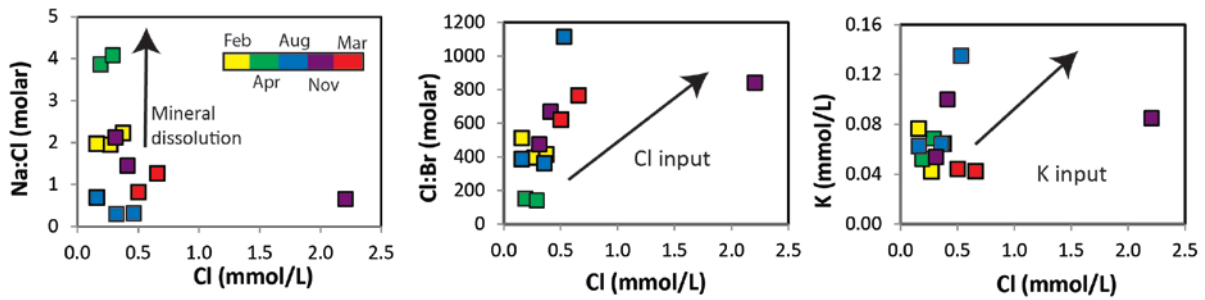
Figure 3. Groundwater elevations during February 2011 and August 2011 at Bruthen (a), Tambo Upper (b) and Kelly Creek (c). White rectangles = measured elevation, dashed lines = interpolated elevations.



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 2 Figure 4. $\delta^{18}\text{O}$ and $\delta^2\text{H}$ values of bank water and river water from the Tambo River. LMWL
 3 defined by Melbourne meteoric water line in Hughes et al., 2012. Dashed lines indicate
 4 typical river water and bank water values.
 5

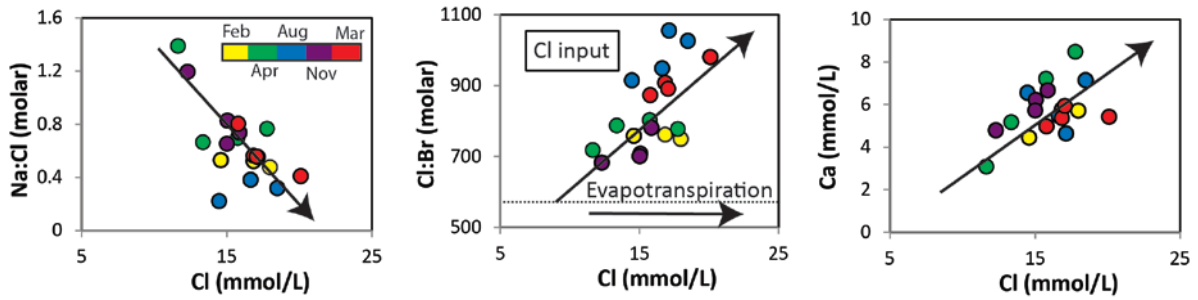


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 7 Figure 5. Piper plot of bank water from the Tambo River. Black markers = HCO_3 measured,
 8 grey and white markers = HCO_3 calculated via charge balance.



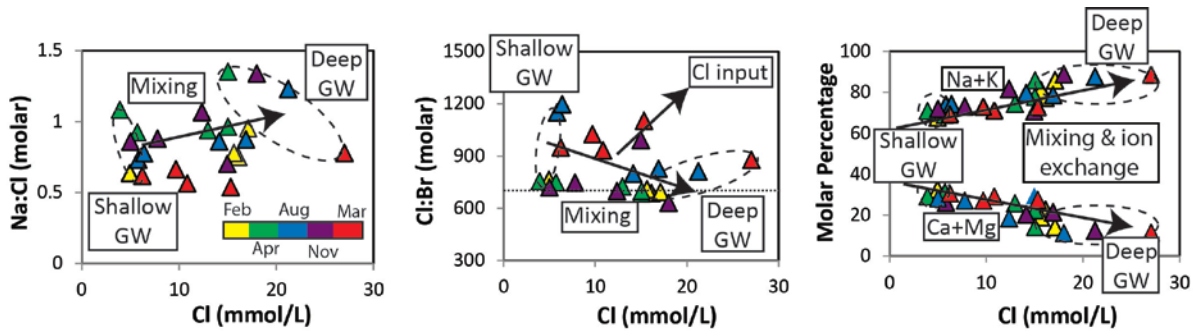
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Figure 6. Trends in major ion chemistry at Bruthen indicating mineral dissolution, the input of Cl into groundwater and the input of K into groundwater.



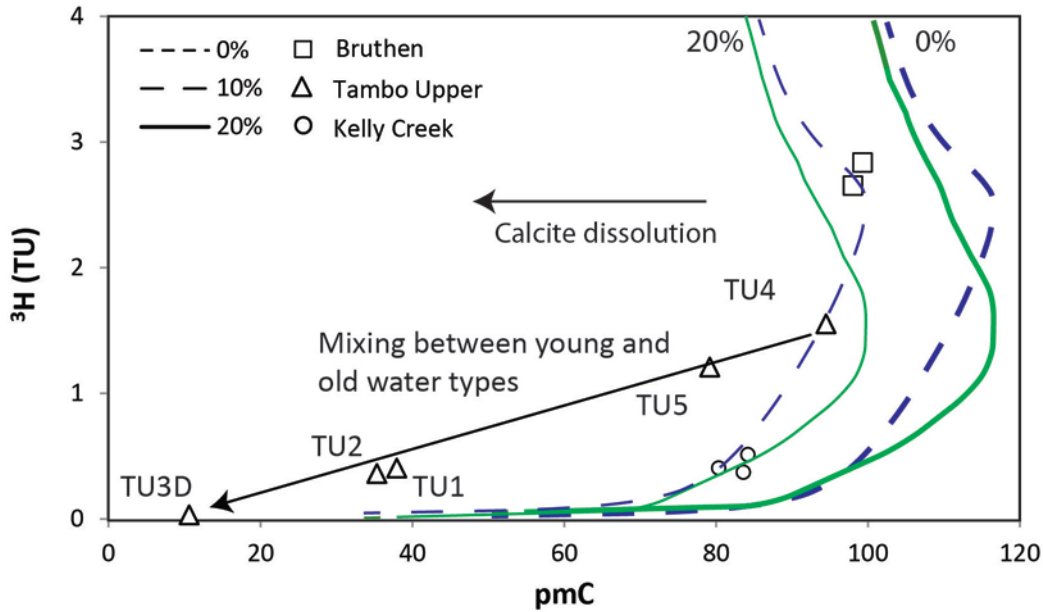
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Figure 7. Trends in major ion chemistry at Kelly Creek indicating Cl inputs during increased rainfall.

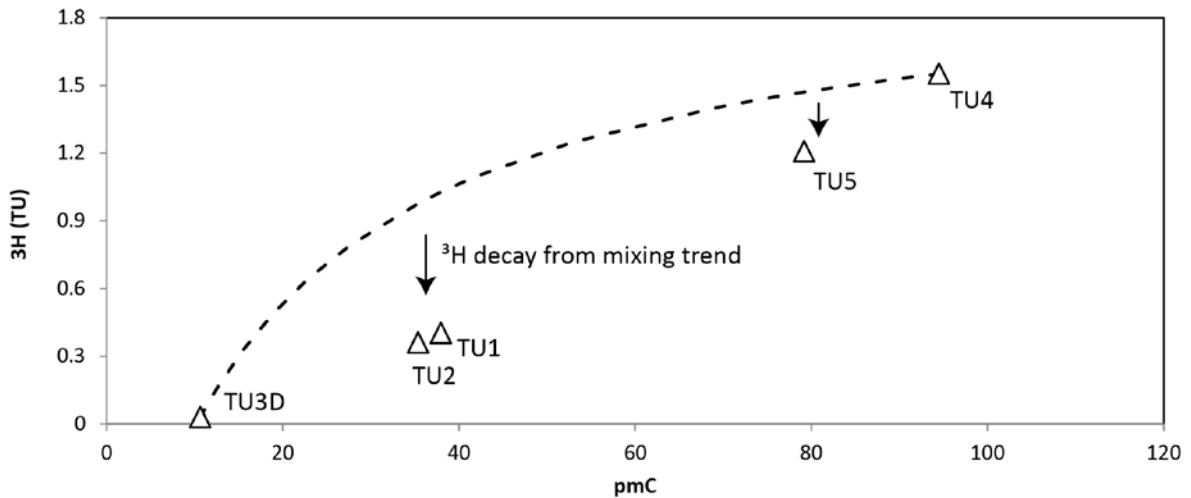


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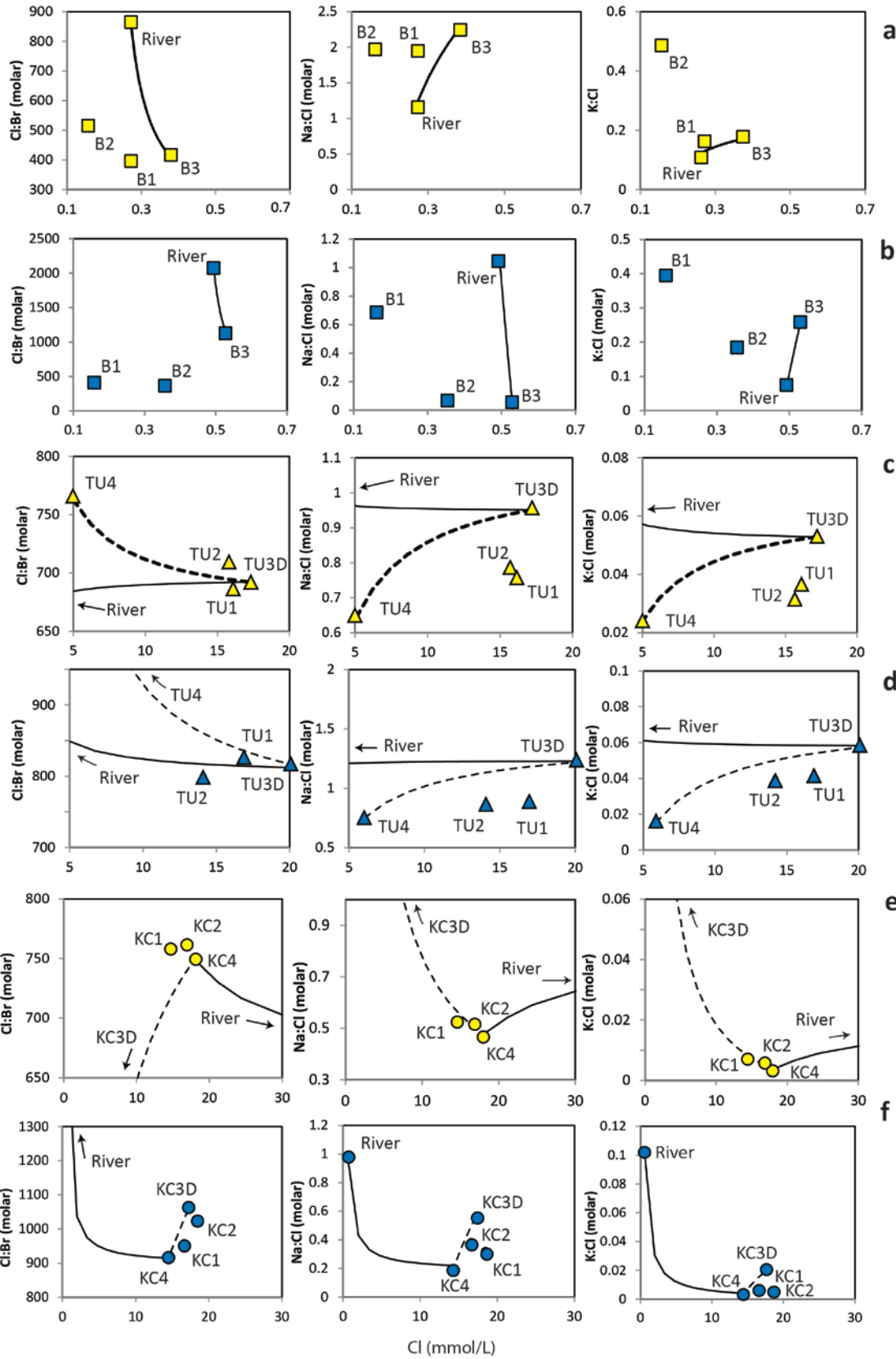
Figure 8. Trends in major ion chemistry at Tambo Upper indicating mixing between groundwater in the shallow, unconfined aquifer and groundwater from the deeper, semi-confined aquifer.



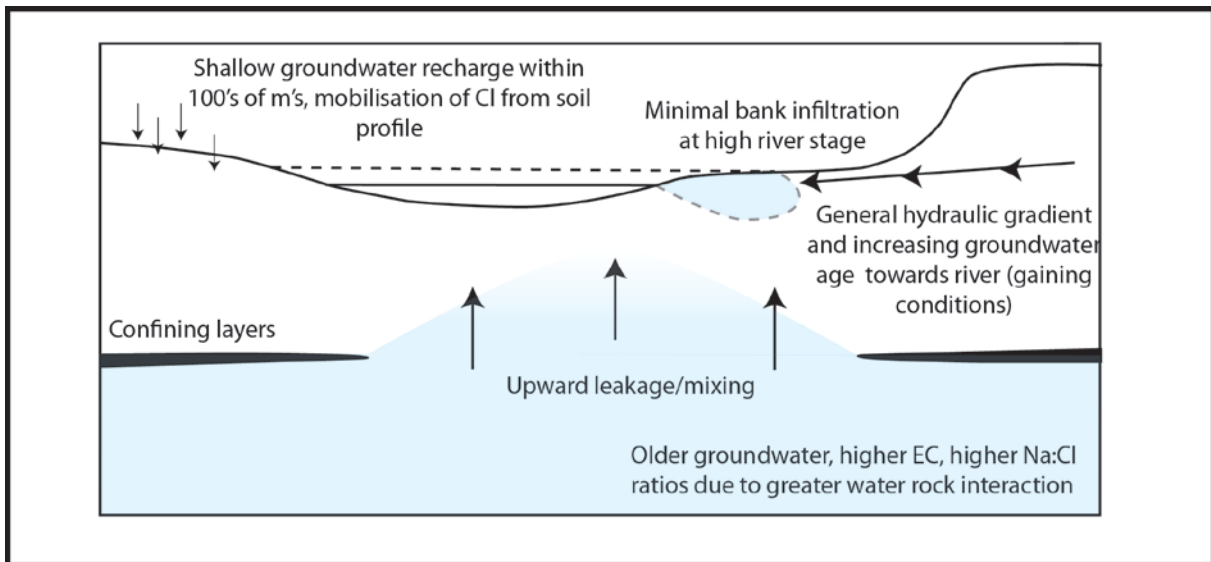
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 2 Figure 9. Co-variance of ^3H and ^{14}C in groundwater and that predicted by Eq. (1) (Solid lines)
 3 for 0% and 20% DIC input from closed system calcite dissolution. Dashed lines are the
 4 predicted of co-variance of ^3H and ^{14}C from the renewal rate model of Le Gal La Salle et al.
 5 (2001).
 6



7
 8
 9 Figure 10. ^3H and ^{14}C activities of groundwater at Tambo Upper and predicted trend for
 10 mixing between deeper groundwater (TU3D) and shallow groundwater (TU4). Curve based
 11 on DIC, ^3H and ^{14}C activities from Table 2.
 12



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3 Figure 11. Predicted mixing curves between river water and groundwater at Bruthen (a,b),
4 Tambo Upper (c,d) and Kelly Creek (e,f) constructed using the end-member compositions
5 from Tables 1 to 3. Yellow data points = February 2011, blue data points = August 2011.



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Figure 12. Schematic representation of the Tambo River and major hydrogeochemical processes at baseflow (solid line) and high flow (dashed line) conditions. The deeper groundwater is from 15 to 20 m below ground surface within the alluvial aquifer and is confined beneath the clay layer.