

(Remarks to Author – in black; Responses in blue):

Anonymous Referee #1

The paper is generally well structured and written, the conclusions are supported by the analysis of the data presented. The paper could be accepted for publication after considering my comments below:

We first would like to thank the reviewer for carefully reviewing the manuscript and for his/her positive feedback on our study. We also appreciate his/her suggestions on how to improve the clarity of the manuscript with inclusion of tables and recommended changes to figures.

- 1) Since the authors analyze data sets from two sites with similar characteristics, a summary table would help the reader to have a clearer view. As it stands now is rather confusing.
- 2) Page 12992 (sub-subsection 2.1.2): Authors mention here that the some characteristics of the IAB site (density of trees and diversity index) are similar to those found at PDG site and refer to Fidelis and Godoy 2003. It is better to include that information in the previous section, when describing the PDG site (carrying the reference as well).

We summarize the main characteristics of each site in a new table. (Table 1, P24 L692)

3) Page 12994, lines 6-8: The software/tool used in the analysis is irrelevant if the authors are confident in the method. If so, the information can be omitted.

Ok. We deleted the used software.

4) Page 12977, lines 22-23: It is not clear why authors present the results of Juarez et al. (2008) without making any other comment or comparison. Would their results (in Amazon rainforest) be comparable with this work? If so, a further discussion should be included. If not, the phrase can be removed.

We deleted the sentence with the results from Juarez et al. (2008).

5) Page 12977, lines 24-27: How would you evaluate Ruhoff results with this work? Do you have a possible explanation on why the present study results are better for R2 and

RMSE values? Would the issue be in their analysis, flux tower data quality or MODIS products?

Ruhoff et al. (2013) compared MOD16 ET product with eddy covariance data from the PDG site during the year of 2001. The authors found values of $R^2 = 0.61$ and RMSE = 0.46 mm d^{-1} , which were not as good as for the present study results. A better result from our model than from the MOD16 ET product was expected because MOD16 runs across the world while our empirical model has been calibrated locally with specific conditions. In other words, a calibrated local model tends to provide better estimates than a global model.

6) Since there is flux data available from PDG since 2000 (used in Ruhoff et al.), why that particular year was not included in this study?

In fact, Ruhoff et al. (2013) compared MOD16 ET product with eddy covariance data from the PDG site only for 2001. Data from the PDG site are available for 2001 through 2003 (see da Rocha et al., 2009). We corrected it in the text. (P10 L274-275)

7) Page 12998, lines 13-14: Authors mention that they observed a significant number of rainfall events in the dry season. Would you label this particular dry season as atypical? If so, would that compromise your analysis?

We found a total of 333 mm in the dry season of 2013 (which is similar to the historical mean in this season of 307 mm) and 92 mm between the months April through July of 2014 (Fig. 3a). (P10 L291-293) Thus, we cannot classify this dry season as atypical.

8) A final table with the found ET results for IAB and PDG sites would make reader's life easier.

We inserted a final table with ET results for IAB and PDG sites. (Table 5, P28 L703)

9) Figure 1: Background is too busy and it is almost impossible to see the map lines and text. Using a plain light colour would be more appropriate.

Thanks for your suggestion. We improved Figure 1. (P29 L707)

10) Figure 2: This figure is very busy and confusing. It is not possible to clearly see the instrumentation, and it does not add anything to the description given in the text. It can be totally removed without prejudice.

Ok. We removed Figure 2.

Other minor observations and suggestions:

a) Figure 4: I would suggest the use of two different colours instead of two shades (black and grey).

Changes made according to reviewer's suggestion. (P31 L715)

b) Figure 5: I suggest the use of colorblind-friendly colours instead of the chosen ones (red+purple+blue).

Changes made according to reviewer's suggestion. (P32 L722)

c) Figure 6: Also here I suggest the use of different colours to be colorblind-friendly. In addition to that, although it may be obvious for a reader that already went through the whole text, the figure caption should bring the description of the used variables. For example "(...) where P is the precipitation, ET is the evapotranspiration, and dS is the water storage change (...)".

Changes made according to reviewer's suggestion. (P33 L727-729)

- d) Page 12990, line 18: "(...) flux tower measurements and vegetation (...)" (no comma). Text corrected. Thanks.
- e) Page 13001, line 10: " evapotranspiration".

Text corrected. Thanks.

Thank you very much for helping us to improve the quality of our manuscript!!

Anonymous Referee #2

In the introduction the deforestation advance appears as the main justification for an article that analyzes only the water balance of a natural savannah ecosystem. This is reflected in the absence of a well-defined goal. The work creates an expectation for the comparison of the water balance in the soil components in a natural savanna site and a near deforested area, which does not happen in the development of the text. The work points the evapotranspiration estimate (by empirical model) as a major focus, when in fact what is presented is the evaluation of the soil water balance components.

We thank the reviewer for all the comments. In the introduction section, we have highlighted the deforestation problem in the Cerrado and the lack of basic information about the hydrological processes in the undisturbed Cerrado. Therefore, to understand pre-deforestation conditions, we investigated several components of the water balance that are still poorly understood for undisturbed tropical woodland classified as "cerrado sensu stricto denso". We believe that our results provide benchmark values of water balance dynamics in the undisturbed Cerrado that will be useful to evaluate past and future land uses in different sceneries of water scarcity and climate change for this region. The importance of such new information/understanding has been recognized by Reviewer #3 for the Brazilian Cerrado

The paper indicates that the sites represent pre-clearing conditions. But in fact do not represent natural ecosystem?

We believe that our sites are representative of natural ecosystems as both studied locations are characterized by undisturbed tropical woodland classified as "cerrado sensu stricto denso".

The article has a confusing structure in the presentation of the methodology and analysis of results, which need to be better organized. Two sites are indicated as study area PDG and IAB, but only the second has analyzed the water balance effectively.

In modeling evapotranspiration item. In experimental IAB site several micrometeorological data was measured in high temporal frequency. Why not used the penman-monteith parameterization for estimating evapotranspiration, or even a land surface model (e.g. SiB2, CLM3, etc.) that would also examine soil moisture? Recent

studies have incorporated physical considerations in estimating the evapotranspiration o in combination with remote sensing data to study area, such as:

da Silva B. B., Wilcox B. P., da Silva V. d. P. R., Montenegro S. M. G. L. and de Oliveira L. M. M.(2014) Changes to the energy budget and evapotranspiration following conversion of tropical savannas to agricultural lands in Sao Paulo State, Brazil, Ecohydrology, DOI: 10.1002/eco.1580.

We did not have a flux tower installed at the IAB site but fortunately there were eddy covariance ET data available at the nearby PDG site, which had very similar vegetation and hydrometeorological characteristics (as presented in Table 1 in the new version of the manuscript). It was very important to have a well-calibrated model to estimate this important component of the water balance. We chose to use and calibrate the MODIS EVI model as it has been successfully applied to determine ET in several natural ecosystems (Glenn et al., 2010 and 2011) and capably reproduced the measurements at the PDG site. It is important to note that was the first evaluation of this approach in the Brazilian Cerrado. This approach includes both an available energy component ("quantified by ETo") and an estimate of green biomass (EVI) which is an advantage over a pure "available energy" type model like Priestly-Taylor. Likewise, calibrating a more complicated model like Penman-Monteith, or even a land surface model (LSM), would have been unnecessary and much more difficult.

LSMs like SiB2 and CLM3 are much more complex and have to represent several aspects of land surface interactions, including the hydrology but also biogeochemistry, vegetation dynamics. Calibrating such models can be quite difficult and due to interactions between individual parameterizations, "competing" parameter values can pose undesired difficulties to model simulations (please refer to Rosolem et al., 2012 and 2013).

With respect to use of LSMs to infer soil moisture quantities. It is not entirely accepted by the LSM community that simulated soil moisture by such models represent actual soil water content conditions or, instead, a "index of wetness" as reported, for instance, by Reichle and Koster (2004). We reproduce the text for clarity "..., simulated soil moisture contents reflect the many necessary simplifications imposed in the land surface model and should arguably be considered model- specific "indices of wetness" rather than quantities that can be measured in the field ..."

The presented evapotranspiration estimation method does not allow the conclusion "...Our findings indicate that the fitted equation may be used to compute ET at daily, monthly and annual scales. ", only by comparing with literature data, longer time series analyzes need to be considered. This results in a fragile analysis.

We calibrated and validated the model for one scale (16 days) using measured ET from eddy covariance at the PDG site. From this empirical model is possible to compute ET at 16 days and these results may be interpolated and/or summed to estimate daily, monthly or annual values. We changed the text to make it clear to the readers. (P10 L283-285 and P15 L446-449)

In pg 12992, line 20: forest no fores

Thanks. We fixed it.

Pg. 12994: Equation 2 is not correct to describe the function that minimizes the mean squared error.

Equation 2 is correct for an objective function of the sum of squared differences.

Pg. 12996: In equation 4, S is soil water storage.

Done. (P9 L243)

Pg. 13016: Figures 4(a,b,c) are very small and illegible.

We thank the reviewer for making this point. We improved the quality of the figure and the size of the numbers. Furthermore, we also changed the colors in the Figures following the suggestion of Reviewer #1. (P31 L715)

In figure 5 soil moisture data are presented but not cited in methodology. The values of the volumetric water content in the soil (y-axis) are in different scale of variation of which is presented in the text (pg 12999).

We corrected it keeping the same units that were used in the text (m³ m⁻³). (P32 L722)

The results show the effect of greatly reduced rainfall over the region in the years 2013 and 2014, which limit their comparison with other times and nearby areas. The great influence of climate variability prevents the comparison of the components of the water balance of the soil, with measurements in deforested/grown areas in different periods.

The annual rainfall during the period of study (1248 and 1139 mm for 2012 and 2013, respectively) were approximately 20% less than the historical mean of the 1500 mm; however, this did not compromise our conclusions because our findings of ET, canopy interception, throughfall, and stemflow were similar to those from previous studies in different periods of study (Tables 4 and 5, P27-28). We did not find other runoff studies in the undisturbed cerrado to make a direct comparison. In addition, periods of hydrological extremes (drought and flood) are often more important than mean values in the water resources studies. These extremes can help to know how a specific hydrological component can explain various changes of the system.

The conclusions reproduce the results without discussions.

A thorough discussion about the results is provided in Section 3 (Results and Discussions). We have limited the Conclusions section to include only the main findings of this study.

Anonymous Referee #3

GENERAL COMMENT:

The manuscript presents new information about water balance over the brazilian Cerrado based on observations. The study analyses a wide range of hydrological fluxes as precipitation, evapotranspiration, interception, surface runoff, infiltration and soil moisture. The authors conclude that 4-20 % of precipitation is intercepted in the canopy, a small fraction runs off and most of the water infiltrates. It was not clear how water flows out the soil (evaporation, subsurface flow or groundwater recharge). Also, observations and previous studies show that removing Cerrado vegetation may generally increase runoff. The contribution of the paper is to bring new information about hydrological processes over an important region (Cerrado) that is still not fully studied. The questions addressed in the paper are important as Cerrado is an important region of Brazil/South America that may experience important transformations, which can cause important impacts over hydrology of major/important river basins. The paper is generally well written, most of the methods are appropriated and conclusions are supported by analyses. I would be pleased to see this work published at HESS. Meanwhile, I have some important comments/suggestions that hopefully will help the authors to improve this manuscript.

We thank the anonymous reviewer for his/her kind words in support of our article and for suggesting some improvements to the manuscript.

SPECIFIC COMMENTS:

-Findings from two sites vs Cerrado:

How these findings (typical values of hydrological fluxes) from 2 sites can be generalized to the Cerrado region? Can it be generalized over a typical catchment of Cerrado? For example, should we expect similar runoff rates at different parts of a catchment (close to a stream or upland)? Or should we expect most of surface runoff generated close to streams at saturated areas, following Dunnian concept of flow generation processes?

This is an interesting point made by the reviewerand we thank he/she for that. We believe that some of hydrological fluxes in the Cerrado may vary in a typical catchment influenced mainly by the water content into soil (e.g. close to a stream or upland). Further, Villalobos-Vega et al. (2014) concluded that water table depth has a strong

influence on variations in tree density and diversity, i.e. regions with deep water tables such as the IAB site (35 m) tend to exhibit greater tree abundance and diversity than sites with a shallow water table. Therefore, if there is variation in the vegetation characteristics we can expect changes in the hydrological fluxes on different parts of a catchment. However, this is not an exclusive characteristic of the Cerrado. In generally, land around a stream (riparian areas) exhibits different hydrological fluxes than in the upland. This is the expected scenario in many catchments.

-Discharge from stream gauges:

The author did a good job in the analyses most of the hydrological fluxes. However, analyses concerning the sinks of soil water are not conclusive. It would be interesting to look at discharge data from stream gauges and convert it to runoff (mm/year) to compare it with the water balance terms obtained in this study. This way, it would be possible to infer about the sink of soil water (evaporation or subsurface and groundwater flow). For example, how overall runoff coefficient compares with runoff ratio obtained using precipitation and discharge from stream gauges? The conclusions concerning water storage can change depending on the results from such analyses.

This is a good point raised by the reviewer. However, we have not monitored streamflow at the IAB site, but surely will be a great point to take into account in a future investigation.

-ET model:

Eq. 4 is a nonlinear function between EVI and ET. But the authors mention that the fitted equation (5) can be used for daily, monthly and annual scales. But as it is not linear, I'm not sure if the equation fitted for one scale (16 days) could be used in other scales (daily of annual).

Our findings indicate that from this fitted equation is possible to compute ET at 16 days and these results may be interpolated and/or summed to estimate daily, monthly or annual values. We have changed the text slightly to refer that. (P10 L283-285 and P15 L446-449)

-Define DBH

DBH is Diameter at breast height, which is the tree diameter measured at 1.30 m above the ground. (P7 L202)

-Figure 4: Please improve quality.

Figure 4 has been improved as suggested by reviewer.

Thank you very much for helping us to improve the quality of our manuscript!!

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1 The water balance components of undisturbed tropical

woodlands in the Brazilian Cerrado

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Abstract: Deforestation of the Brazilian Cerrado region has caused major changes in hydrological processes. These changes in water balance components are still poorly understood, but are important for making land management decisions in this region. To better understand pre-deforestation conditions, we determined the main components of the water balance for an undisturbed tropical woodland classified as "cerrado sensu stricto denso". We developed an empirical model to estimate actual evapotranspiration (ET) by using flux tower measurements and, vegetation conditions inferred from the enhanced vegetation index and reference evapotranspiration. Canopy interception, throughfall, stemflow, surface runoff, and water table level were assessed from ground measurements. We used data from two Cerrado sites, "Pé de Gigante" - PDG and "Instituto Arruda Botelho" - IAB. Flux tower data from the PDG site collected from 2001 through 2003 were used to develop the empirical model to estimate ET. The other hydrological processes were measured at the field scale between 2011 and 2014 at the IAB site. The empirical model showed significant agreement (R^2 =0.73) with observed ET at the daily time scale. The average values of estimated ET at the IAB site ranged from 1.91 to 2.60 mm d⁻¹ for the dry and wet season, respectively. Canopy interception ranged from 4 to 20% and stemflow values were approximately 1% of gross

precipitation. The average runoff coefficient was less than 1%, while Cerrado deforestation has the potential to increase that amount up to 20 fold. As relatively little excess water runs off (either by surface water or groundwater) the water storage may be estimated by the difference between precipitation and evapotranspiration. Our results provide benchmark values of water balance dynamics in the undisturbed Cerrado that will be useful to evaluate past and future land cover and land use changes for this region.

Keywords: evapotranspiration, throughfall, stemflow, runoff, savanna, deforestation, water balance, canopy interception.

1 Introduction

As global demand for agricultural products such as food, fiber, and fuel grows to unprecedented levels, the supply of available land continues to decrease, which is acting as a major driver of cropland and pasture expansion across much of the developing world (Gibbs et al., 2010; Macedo et al., 2012). Vast areas of forest and savannas in Brazil have been converted into farmland, and there is little evidence that agricultural expansion will decrease, mainly because Brazil holds a great potential for further agricultural expansion in the twenty-first century (Lapola et al., 2014).

The Amazon rainforest and Brazilian savanna (Cerrado) are the most threatened biomes in Brazil (Marris, 2005). However, the high suitability of the Cerrado topography and soils for mechanized agriculture, the small number and total extent of protected areas, the lack of a deforestation monitoring program, and the pressure resulting from decreasing deforestation in Amazonia indicates that the Cerrado will continue to be the main region of farmland expansion in Brazil (Lapola et al., 2014). In fact, Soares-Filho et al. (2014) reported that the Cerrado is the most coveted biome for agribusiness expansion in Brazil, given its 40 ± 3 Mha of land that could be legally deforested.

The Brazilian Cerrado, one of the richest ecoregions in the world in terms of the biodiversity (Myers et al., 2000), covers an area of 2 million km² (~22% of the total area of Brazil), however, areas of remaining native vegetation represent only 51% of this total (IBAMA/MMA/UNDP, 2011). In addition to being an important ecological and agricultural region for Brazil, the Cerrado is crucial to water resource dynamics of the country, and includes portions of 10 of Brazil's 12 hydrographic regions (Oliveira et al., 2014). Further, the

largest hydroelectric plants (comprising 80% of the Brazilian energy) are on rivers in the Cerrado. As savannas and forests have been associated with shifts in the location, intensity and timing of rainfall events, lengthening of the dry season and changed streamflow (Davidson et al., 2012; Spracklen et al., 2012; Wohl et al., 2012), it is clear that land cover and land use change promoted by the cropland and pasture expansion in this region have the potential to affect the ecosystems services and several important economic sectors of Brazil, such as agriculture, energy production and water supply.

Although all indications are that farmland expansion will continue in the Cerrado and that the land cover and land use will promote changes in water balance dynamics, few studies have been undertaken to investigate the hydrological processes at the field scale (plots or hillslope). In general, the studies on the Cerrado hydroclimatic variability have been done on large areas (Loraie et al., 2011; Davidson et al., 2012; Oliveira et al., 2014). Evapotranspiration (ET) has been the most intensively studied component of the water balance at the field scale, usually based on eddy covariance methods (Vourlitis et al., 2002; Santos et al., 2003; da Rocha et al., 2009; Giambelluca et al., 2009) or by the water balance in the soil (Oliveira et al., 2005; Garcia-Montiel et al., 2008). However, other water balance components such as rainfall interception, canopy throughfall, stemflow, surface runoff, infiltration, percolation, subsurface flow and groundwater recharge are poorly understood in the Cerrado due to lack of available observations.

To understand pre-deforestation conditions, the objective of this study was to determine the main components of the water balance for an undisturbed tropical woodland classified as "cerrado sensu stricto denso". We developed an empirical model to estimate actual evapotranspiration (ET) by using flux tower measurements and vegetation conditions inferred from the enhanced vegetation index (EVI) and reference crop evapotranspiration (ETo). Canopy interception, throughfall, stemflow, and surface runoff were assessed from ground measurements. We used data from two cerrado sites, "Pé de Gigante" - PDG and "Instituto Arruda Botelho" - IAB. Flux tower data from the PDG site collected from 2001 to 2003 was used to develop the empirical model to estimate ET. The other hydrological processes were measured at the field scale between 2011 and 2014 at the IAB site. A more comprehensive accounting of individual water balance components in the Brazilian Cerrado ecosystem is of paramount importance for understanding hydrological cycle shifts in the future due to possible land-use/land-cover changes.

2 Data and Methods

2.1 Study Sites

We developed this study using data from two cerrado sites, "Pé de Gigante" - PDG and "Instituto Arruda Botelho" - IAB, referenced throughout the text as PDG and IAB, respectively. Both sites are located in the State of São Paulo and are separated from each other by approximately 60 km (Fig. 1). The physiognomy of PDG and IAB sites was classified as "cerrado sensu stricto denso", which is also known as cerrado woodland, and has a characteristic arborous cover of 50% to 70% and trees with heights of 5 to 8 m (Furley 1999). Similar soil characteristics, hydroclimatology and phenology were found between these sites (Table 1).

105 Insert Figure 1

106 Insert Table 1

'Pé de Gigante' site (PDG)

We used field measurements collected at the PDG flux tower located on a contiguous 1060 ha undisturbed woodland in the municipality of Santa Rita do Passo Quatro, São Paulo State (latitude 21°37' S, longitude 47°39' W, elevation:~ 700 m). According to the Köppen climate classification system, the climate in this area is Cwa humid subtropical, with a dry winter (April to September) and hot and rainy summer (October to March). The soil is classified in the Brazilian Soil Classification System (SiBCS) as Ortic Quartzarenic Neosol (RQo) with less than 15% clay. Net radiation (Rn), latent heat (LE), sensible heat (H) fluxes and ancillary meteorological data were measured at a height of 21 m and recorded every half-hour from January 2001 through December 2003. Details about the equipment and measurement procedures used are provided by da Rocha et al. (2002, 2009).

'Instituto Arruda Botelho' site (IAB)

The IAB site is a 300 ha, undisturbed woodland located in the municipality of Itirapina, São Paulo State (latitude 22°10' S, longitude 47°52' W, elevation: 780 m). The soil is also classified as Ortic Quartzarenic Neosol with sandy texture in the entire profile (85.7% sand, 1.7% silt, and 12.6% clay), and soil bulk density of 1.7 g cm⁻³. We installed an 11 m instrumental platform to measure basic above-canopy meteorological and soil variables (Table 2). A datalogger (Campbell CR1000, Logan UT, USA) sampled the weather station and soil data every 15 s and recorded averages on a 10 min basis.

130 Insert Table 2

2.2 Modeling Evapotranspiration

In Brazil, there are a few flux tower sites in native cerrado vegetation. These sites were located in the States of São Paulo (da Rocha et al., 2002 and 2009), Brasilia (Giambelluca et al., 2009; Miranda et al., 1997), and Mato Grosso (Vourlitis et al., 2002). There is a lack of information about ET in other Cerrado regions. To fill this gap, some authors have combined vegetation indices (VI) from the remote sensing data with ground measures of ET (usually flux tower) to spatially extrapolate ET measurements over nearby regions with few or no ground data. This process consists in the use of ground measurements of ET from flux towers set in natural ecosystems to develop a best-fit equation between ET, satellite-derived VIs, ancillary remote sensing data, and ground meteorological data (Glenn et al., 2010, 2011). Such an approach has been successfully applied to determine ET in natural ecosystems such as: riparian zones (Scott et al., 2008), shrublands (Nagler et al., 2007), rangeland and native prairie (Wang et al., 2007) temperate grassland, boreal forest, tundra (Mu et al., 2009) and Amazon rainforest (Juárez et al., 2008).

VIs are a ratio derived from the red and near-infrared spectral reflectance, and are strongly correlated with physiological processes that depend on photosynthetically active radiation absorbed by a canopy, such as transpiration and photosynthesis (Glenn et al., 2010). Normalized Difference Vegetation Index (NDVI) and the Enhanced Vegetation Index (EVI) from the Moderate Resolution Imaging Spectrometer (MODIS) on the NASA Terra satellite are VIs widely used in environmental studies. However, previous studies have shown that EVI can better capture canopy structural variation, seasonal vegetation variation, land cover

variation, and biophysical variation for high biomass vegetation (Huete et al. 2002; Juárez et al., 2008). In addition, EVI has been a better predictor of ET than NDVI (Nagler et al., 2005a, b; Glenn et al., 2007; Wang et al., 2007).

We developed an empirical relationship between ET from the PDG flux tower, MODIS
Enhanced Vegetation Index (EVI) and reference crop evapotranspiration (ETo) following the
approach used by Nagler et al. (2013):

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$$ET = ETo \left[a \left(1 - e^{(-bEVI)} \right) - c \right]$$
 (1)

where a, b and c are fitting coefficients and $(1 - e^{(-bEVI)})$ is derived from the Beer-Lambert Law modified to predict absorption of light by a canopy. The coefficient c accounts for the fact that EVI is not zero at zero ET since bare soil has a low but positive EVI (Nagler et al., 2004, 2013).

Daily average ET values from the PDG flux tower were computed by first filling the gaps in the 1-hour data that were due to sensor malfunctions or bad measurements. Gaps were filled using 1-hour averages of photosynthetically active radiation (PAR) and a 14-day look-up tables of ET values averaged over 100 micromoles m⁻² s⁻¹ intervals (Falge et al., 2001). Then we computed daily ET averages over every 16 days to be in sync with the 16-day EVI data. We used EVI data provided by the MODIS product MOD13Q1 (http://daac.ornl.gov/MODIS/). These data are provided by National Aeronautics and Space Administration (NASA) as atmospherically and radiometrically corrected 16-day composite images with a 250 m spatial resolution. We obtained the MODIS EVI pixel centered on the flux tower. Daily ETo was computed according to the FAO-56 method (Allen et al., 1998) and then averaged over 16 days.

We used the parameter optimization tool Genetic Algorithm to fit Eq. 1, incorporating the time series of measured ET, EVI and ETo for 2001 through 2003. This process consisted of minimizing the sum of squared differences between the ET observed from eddy covariance and estimated by Eq. 1:

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$$function = \sum_{i=1}^{n} [ET(i)obs - ET(i)sim]^2$$
 (2)

where ET(i)obs is the observed ET and ET(i)sim is modeled ET at time (i).

For model validation, we calibrated the model using 2001 and 2002 data and then predicted ET for 2003. After this validation process we fit Eq. 1 again, but this time considering the full time series that was available. The coefficient of determination (R^2) ,

standard deviation of differences between observed and estimated ET (*SD*), root mean square (*RMSE*) and the Student's t-test with a 95% confidence level were used to evaluate the significance of the linear relationship between the observed and estimated ET.

2.3 Hydrological processes measured at the IAB site

Canopy interception

Canopy interception (CI) was computed as the difference between the gross precipitation (P_g) and the net precipitation (P_n), where P_g is the total precipitation that fell at the top of the canopy and P_n was computed as the sum of two components: throughfall (TF) and stemflow (SF):

194
$$CI = P_g - P_n = P_g - (TF + SF)$$
 (3)

We measured the P_g from an automated tipping bucket rain gauge (model TB4) located above the canopy at 11 m height (Table 2). TF was obtained from 15 automated tipping bucket rain gauges (Davis Instruments, Hayward, CA) distributed below the cerrado canopy and randomly relocated every month during the wet season. Each rain gauge was installed considering an influence area of 10 x 10 m. SF was measured on 12 trees using a plastic hose wrapped around the trees trunks, sealed with neutral silicone sealant, and a covered bucket to store the water. Selected trees to be monitored were divided into two groups considering the diameter at breast height (DBH), which is the tree diameter measured at 1.30 m above the ground. Therefore, we monitored 7 trees with 5 cm < DBH < 20 cm and 5 trees with DBH > 20 cm. The volume of water in each SF collector was measured after each rainfall event that generated stemflow, totaling 42 SF measurements during the study period. The volume of water measured from each sample tree was expressed as an equivalent volume per m² of basal area, and then this value was multiplied by the site basal area (27.75 m² ha⁻¹) to compute stemflow in mm (Dezzeo and Chacón, 2006 and MacJannet et al., 2007). We measured P_g , TF and SF from September 2012 through July 2014.

Surface runoff

Surface runoff was measured from 100 m² experimental plots of 5 m width and 20 m length from January 2012 through July 2014. To evaluate the cover influence on the surface

runoff, experimental plots were installed under native vegetation and bare soil with steepness of approximately 0.09 m m⁻¹. Each treatment had three replications and plots on bare soil were located about 1 km from the plots under undisturbed cerrado. The boundaries of the plots were made using galvanized sheet placed 30 cm above the soil and into the soil to a depth of 30 cm. Surface runoff was collected in storage tanks at the end of each plot. Plots under bare soil were built with three storage tanks with 310 liters capacity each and two splitters of one seventh, i.e. one seventh were collected in the second tank and one forty ninth in the third tank. In the plots under cerrado vegetation only one storage tank with a capacity of 310 liters for each plot was used to collect runoff and soil loss because of the expected lower runoff amounts from those plots.

Surface runoff was measured for each erosive rain event under the undisturbed cerrado and bare soil. Periods of rainfall were considered to be isolated events when they were separated by periods of precipitation between 0 (no rain) and 1.0 mm for at least 6 h, and were classified as erosive events when 6.0 mm of rain fell within 15 min or 10.0 mm of rain fell over a longer time period (Oliveira et al., 2013). We used this approach because in general only erosive rainfall has promoted surface runoff in the study area. A total of 65 erosive rainfall events were evaluated during the study period.

Groundwater recharge

The water table level was monitored from December 2011 through July 2014 from a well with 42 m in depth installed in the undisturbed cerrado. Water-table fluctuation data were measured daily from a pressure sensor (Mini-Diver model DI501, Schlumberger Limited, Houston, USA).

2.4 Water balance at the IAB site

We evaluated the water balance components in the IAB site at the daily, monthly and annual time scales from January 2012 to March 2014 (Eq. 4). We used measured data of precipitation, surface runoff, and direct recharge. Evapotranspiration was estimated using the fitted equation from the EVI and reference evapotranspiration data.

$$\frac{dS}{dt} = P - ET - Q - R \tag{4}$$

243 where S is the soil water storage change with time, P is precipitation, ET is evapotranspiration, O is runoff, and R groundwater recharge.

3 Results and Discussion

3.1 Modeling ET

The daily average (\pm standard deviation) reference evapotranspiration (ETo), measured evapotranspiration (ET), and EVI at the PDG site were 4.56 ± 0.73 mm d⁻¹, 2.31 ± 0.87 mm d⁻¹, and 0.41 ± 0.09 , respectively. We found a significant correlation between observed ET and EVI with a correlation coefficient of 0.75 (p < 0.0001). EVI showed similar seasonality that was observed for the ET and ETo during wet and dry seasons (Fig. 2). The average ET and EVI values for the wet season were 2.81 ± 0.57 mm d⁻¹ and 0.48 ± 0.05 , and for the dry season 1.70 ± 0.70 mm d⁻¹ and 0.33 ± 0.05 , respectively.

257 Insert Figure 2

The fitted equation considering the periods of calibration, validation and full time series at 16-day averages showed good results in the ET estimates, with a coefficient of determination (R^2) greater than 0.70 and standard deviation of differences between observed and estimated ET (SD) and root mean square (RMSE) less than 0.50 mm d⁻¹ and 21%, respectively (Table 3). The final form of the fitted equation was:

264
$$ET = ETo [10.36 (1 - e^{(-12.31EVI)}) - 9.74]$$
 (5)

266 Insert Table 3

The modeled values of ET estimated for the full period, wet and dry seasons $(2.30 \pm 0.76 \text{ mm d}^{-1}, 2.81 \pm 0.31 \text{ mm d}^{-1}, \text{ and } 1.69 \pm 0.60 \text{ mm d}^{-1}, \text{ respectively})$ were not significantly different (p = 0.05) from the observed values of ET during the same period. Furthermore, we found better values of R², SD, and RMSE of 0.78, 0.16 mm month⁻¹, and 17.07% at the

monthly scale. The annual average ET observed and estimated for the three years studied (2001-2003) were 822 mm yr⁻¹ and 820 mm yr⁻¹, respectively, with an RMSE of 6.12%. Observed ET during 2001 from the PDG site was compared previously by Ruhoff et al. (2013) with the ET estimated from the product MOD16 (Mu et al., 2011). The authors found values of $R^2 = 0.61$ and RMSE = 0.46 mm d⁻¹, which were not as good as for the present study results. In a review paper about ET estimation in natural ecosystems using vegetation index methods, Glenn et al. (2010) reported values for different temporal scales ranging from 0.45 to 0.95 for the R^2 and of 10 to 30% for the RMSE. They concluded that the uncertainty associated with remote sensing estimates of ET is constrained by the accuracy of the ground measurements, which for the flux tower data are on the order of 10 to 30%. Hence, the values of SD and RMSE reported in the present study are within the error bounds of the likely ground measurement errors. Our findings indicate that from this fitted equation is possible to compute ET at 16 days and these results may be interpolated and/or summed to estimate daily, monthly or annual values.

3.2 Canopy interception, throughfall, and stemflow

The gross precipitation (P_g) in the IAB site during the 23 month study period was 1929 mm, where 78% of this total occurred from October through March (wet season). We found similar values of 766 mm and 734 mm for the two wet seasons studied, 2012-2013 and 2013-2014. We found a total of 333 mm in the dry season of 2013 (which is similar to the historical mean in this season of 307 mm) and 92 mm between the months April through July of 2014 (Fig. 3a). The sum of throughfall (TF) was 1566 mm, which corresponded to 81.2% of $P_{\rm g}$. Individual wet season TF values were 81.9 and 82.3% of $P_{\rm g}$ while total dry season $P_{\rm g}$ was 74.8%. The coefficient of determination between P_g and TF was 0.99 (p < 0.0001) over the 253 rainfall days (Fig. 3b). Stemflow values (by 42 events) ranged from 0.3 to 2.7% with an average of 1.1% of $P_{\rm g}$. The greatest values of SF were found in the beginning of the wet season (October and November) and the smallest values occurred in the middle of the wet season (January and February). This suggests that there is an influence of condition of trees trunks (dry and wet) and canopy dynamics in the stemflow. Furthermore, we found greater values of SF in the trees with 5 cm < DBH < 20 cm (1.6% of P_g) than the trees with DBH > 20 cm (0.4% of P_g), which is consistent with results reported by Bäse et al. (2012) for the transitional Amazonia-Cerrado forest.

Insert Figure 3

We found only three previous studies about interception process in the Brazilian Cerrado. The values reported in the literature for TF and SF, ranged from 80 to 95% of $P_{\rm g}$ and <1 to 2.4% of $P_{\rm g}$, respectively (Table 4). In the present study the canopy interception (CI) was 17.7% of $P_{\rm g}$. Therefore, considering our findings and previous studies presented in Table 4 we can suggest that CI in the undisturbed cerrado ranges from 4 to 20% of $P_{\rm g}$. However, future studies are necessary to understand the influence of physiognomies of the Cerrado in the CI processes. This region is large and complex and varies from grassland to savanna to forest (Furley, 1999; Ferreira and Huete, 2004). In addition, other characteristics such as conditions trees trunks (crooked and twisted), stand structure, canopy cover, rainfall features, and the litter interception should be better studied in specific studies of rainfall interception processes.

319 Insert Table 4

3.3 Cerrado water balance

The measured annual precipitation at the IAB site was 1248 mm, 1139 mm, 421 mm for 2012, 2013 and January through July of 2014, respectively. We measured 65 rainfall events that generated surface runoff during the study. The runoff coefficient for individual rainfall events (total runoff divided by total rainfall) ranged from 0.003 to 0.860 with an average value and standard deviation of 0.197 ± 0.179 in the bare soil plots. The highest values were found for larger, more intense rainfall events, or in periods with several consecutive rainfall events, which induced high soil moisture contents and consequently greater runoff generation. Moreover, the runoff coefficient found for the bare soil plots (~20%) indicates that the soil in the study area (sandy soil) has a high infiltration capacity. Runoff coefficients ranged from 0.001 to 0.030 with an average of less than 1% (0.005 \pm 0.005) in the plots under undisturbed cerrado. Youlton (2013) studied in two hydrological years (2011-12 and 2012-13) the surface runoff using plots installed in the same experimental area as the present study and found values of 3.6 to 5.1% and 2.0 to 5.0% for the runoff coefficient under pasture and sugarcane,

respectively. Cogo et al. (2003) reported values of runoff coefficient for soybeans and oat ranging from 2.0 to 4.0% depending to the soil tillage and management. Pasture, sugarcane and soybeans are the main cover types that have been used to replace the undisturbed cerrado lands (Loarie et al., 2011; Lapola et al., 2014). Therefore our results indicate that the cerrado deforestation has the potential to increase surface runoff around 5 fold when the cerrado is replaced for pasture and croplands and up to 20 fold for bare soil conditions.

Infiltration was calculated after subtracting interception (without accounting for the litter interception) and surface runoff from the gross precipitation. Thereby we found that 79% of gross rainfall infiltrated into the soil. Fig. 4 shows the amount of infiltration and the volumetric water content (VWC) up to 1.5 m in depth. We found a rapid increase in the VWC as a function of infiltration, indicating that the sandy soil found in the IAB site promoted fast infiltration, mainly in the first meter depth of the soil profile. VWC ranged from 0.08 to 0.23 m³ m⁻³ and 0.08 to 0.17 m³ m⁻³ for 0.1 and 1.5 m soil depths, respectively. However, it is important to note that the root zone for trees in the cerrado is usually deep (more than 10 m in depth) and limited by the water table level (Oliveira et al, 2005; Garcia-Montiel et al., 2008; Villalobos-Vega et al., 2014). Therefore, the 1.5 m soil profile is not representative for evaluating the water use by vegetation, but is useful to evaluate the response for rainfall events and evaporative processes. Oliveira et al. (2005) concluded that the water stored in deep soil layers (1 to 4 m) provides approximately 75% of the total water used for an undisturbed cerrado classified as "cerrado sensu stricto denso", the class that includes the IAB and PDG sites.

357 Insert Figure 4

The amount of water infiltrated into the soil was not enough to elevate the water table level in the well during the study period, from December 2011 to July 2014. This was because the water table in the monitored well was approximately 35 m deep. In other words, there is a large distance from the soil surface to the water tables, and the amount of water that eventually reached the saturated zone was not enough to cause an immediate change in the water table level. One of the first studies of groundwater dynamics in the undisturbed cerrado was conducted by Villalobos-Vega et al., (2014) from 11 monitored wells with water tables ranging from 0.18 to 15.56 m. The authors found little water table change in regions with

deep water table (up to 15.56 m), and in some wells the recharge water took up to 5 months to reach the groundwater table. They also concluded that water table depth has a strong influence on variations in tree density and diversity, i.e. regions with deep water tables such as the IAB site (35 m) tend to exhibit greater tree abundance and diversity than sites with shallow water table. Therefore, the infiltrated water in the present study was likely either extracted and transpired by the vegetation, drained by lateral subsurface flow (not measured in this studied, but probably small due to the flat topography of the site) or stored in the vadose zone.

Groundwater recharge is also affected by land use and land cover change (Scanlon et al., 2005; Dawes et al., 2012). We found that the undisturbed cerrado tends to provide more infiltration than areas covered with pasture and cropland. On the other hand, the cerrado vegetation has significant canopy interception and evapotranspiration that result in little groundwater recharge as compared to pasture and cropland. Using 23 monitoring wells distributed in a watershed located 5 km away from the IAB site, Wendland et al. (2007) showed that the groundwater recharge varies with the land cover. The authors reported values of annual recharge and water table depth, respectively, ranging from 145 to 703 mm yr⁻¹ (5 to 16 m) in pasture, 324–694 mm yr⁻¹ (9 to 22 m) in orange citrus, and 37–48 mm yr⁻¹ (21 m) in eucalyptus forests. Therefore, cerrado deforestation has the potential to change groundwater recharge dynamics.

The average values of actual evapotranspiration (ET) estimated by Eq. 5 for the IAB Cerrado site for the full period, wet and dry seasons were similar to that observed in the PDG site (Table 5). The annual average ET estimated for the two years studied (2012-2013) was 823 mm yr⁻¹, which also is consistent with that found by Giambelluca et al. (2009) of 823 mm yr⁻¹ and the PDG site of 822 mm yr⁻¹. Given that surface runoff was less than 1% of precipitation and groundwater recharge and subsurface lateral flow was likely small, vadose zone water storage is basically the difference between precipitation and evapotranspiration (Fig. 5).

394 Insert Table 5

395 Insert Figure 5

Water deficits in the Cerrado region usually happen from April through September (dry season), however we found an atypical water decrease in the wet season (months of March and November 2012, and January 2014). Indeed, the rainfall amounts in these months were 71%, 56% and 39% less than the historical mean of 1973 to 2013 (156 mm, 147 mm and 270 mm) observed at the climatological station from the Centro de Recursos Hídricos e Ecologia Aplicada at the University of São Paulo, located approximately 3 km from the study area. In addition, we note that the annual rainfall during the period of study (1248 mm and 1139 mm for 2012 and 2013, respectively) were approximately 20% less than the historical mean of the 1500 mm. The decreased rainfall in São Paulo State in recent years has caused problems of water scarcity (Rodrigues et al., 2014).

3.4 Broader implications for hydrological processes in the Cerrado Regions

Values of water fluxes found in this study indicate that deforestation of the undisturbed Cerrado lands has the potential to increase runoff and groundwater recharge, and decrease canopy interception and evapotranspiration, at local or regional scales. However, the interaction of these processes over large areas may be different than that reported here, because hydrological interactions and responses are dependent on the scale studied (Costa et al., 2003; Oliveira et al., 2014). Our results represent one of the first measured values for this undisturbed condition, and therefore may be used as a benchmark for future studies. Future investigations are necessary to better understand the hydrological processes in the undisturbed Cerrado, including the poorly studied water fluxes such as canopy interception, surface runoff, infiltration, percolation, subsurface flow and groundwater recharge. Further, as the Cerrado is a large biome that has different conditions of vegetation, soil types and hydrometeorology, more investigations should be conducted to cover all its conditions.

As land cover and land use of the Cerrado biome have been quickly changed over recent decades with the expansion of pasture and crops (Gibbs et al., 2015), benchmark values of hydrological processes are crucial to understand pre-disturbance conditions. A better understanding of hydrological processes within the Cerrado region can be used to better constrain and consequently improve hydrometeorological models. For instance, flux tower measurements at the PDG site have been successfully used to improve the understanding of the mechanisms associated with energy and carbon partitioning from several land surface models in the LBA Data Model Intercomparison Project (de Goncalves et al. 2013); as well as

for model diagnostics, and parameter identification and calibration by Rosolem et al. (2012, 2013). The present study can potentially expand such analyses by introducing water partitioning components such as change in soil moisture, infiltration, runoff, and canopy interception. In addition, the use of remote sensing data to estimate hydrological processes, such as the approach developed in the present study to estimate ET, is a viable alternative for evaluating the water balance spatially in the Cerrado. The possibility to assess the water balance spatially will be useful to create environmental zoning plans in this region that seek to conserve and preserve native Cerrado vegetation, and also to suggest appropriate and effective land use management practices for farmers (Oliveira et al., 2011). There is still much work to be done in the Cerrado region to understand its unique hydrology. However, in this study we show findings that contribute toward that goal.

4 Conclusions

We developed an empirical model to estimate actual evapotranspiration by using flux tower measurements and, vegetation conditions inferred from the enhanced vegetation index and reference evapotranspiration. We used flux tower data from the PDG site collected during 2001 to 2003. The empirical model developed in the present study showed a significant agreement with observed ET and better results than from the product MOD16 ET. From this empirical model is possible to compute ET at 16 days and these results may be interpolated and/or summed to estimate daily, monthly or annual values for undisturbed cerrado areas with similar characteristics of hydroclimatology and phenology that observed at the PDG site. Furthermore, from this approach it is possible to assess the ET for large areas of the Cerrado with a good spatial and temporal resolution (250 m and 16 days), therefore, it may be useful for monitoring evapotranspiration dynamics in this region.

Canopy interception, throughfall, stemflow, surface runoff, and water table level were assessed from ground-measurements at the field scale between 2011 and 2014 at the IAB site. We conclude that the canopy interception may range from 4 to 20% of gross precipitation in the cerrado and that stemflow values are around 1% of gross precipitation. Our results also indicate that the average runoff coefficient was less than 1% in the plots under undisturbed cerrado and that the deforestation has the potential to increase up to 20 fold the runoff coefficient value. In addition, we did not find evidence of net groundwater table changes, possibly because the water table is at significant depth at the IAB site, the deep rooting depth

of the trees, and the study period with rainfall smaller than the historical mean. As only little excess water runs off (either by surface water or groundwater) the water storage may be estimated by the difference between precipitation and evapotranspiration.

Deforestation of the Brazilian Cerrado has caused major changes in hydrological processes; however these changes are still poorly understood at the field scale. Thus, understanding pre-deforestation conditions including the main components of the water balance is of paramount importance for an undisturbed cerrado. In this study, we provide benchmark values of water balance dynamics in the undisturbed Cerrado that will be useful to evaluate past and future land use in different sceneries of water scarcity and climate change for this region.

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Table 1. Summary of characteristics of the studied areas.

Description	PDG	IAB
Köppen climate classification system	Cwa humid subtropical	Cwa humid subtropical
Average annual precipitation (mm) and temperature (°C)	1478 and 21.1	1506 and 20.8
Soil texture	sandy texture	sandy texture
	"cerrado sensu stricto	"cerrado sensu stricto
Vegetation physiognomy	denso"	denso"
	15,278 individuals per	13,976 individuals
Absolute density of trees	hectare*	per hectare**

^{*} Reys 2008 and ** Fidelis and Godoy, 2003.

Table 2. Data collected at the IAB site.

Variable description	Sensor	Height or depth (m)	
Temperature and relative humidity	Psychrometer HC2S3	9	
Wind speed and direction anemometer	Anemometer RM Young 05103-5	10	
Net radiation	NR-LITE2	10	
Global solar radiation	LiCor 200X	10	
Precipitation	Texas TB4	10	
Atmospheric pressure	Barometer Vaisala CS106	2	
Soil moisture	EnviroScan SENTEK	0.10, 0.50, 0.70, 1.00, 1.50	

Table 3. Model calibration and validation results reported as the coefficient of determination (R^2), standard deviation of differences (SD), and root mean square errors (RMSE) for 16-day averages

Time series	R^2	SD (mm day ⁻¹)	RMSE (%)
Calibration, 2001-2002	0.71	0.50	20.92
Validation, 2003	0.83	0.33	15.69
Full time series, 2001-2003	0.73	0.45	19.53

Table 4. Previous studies of throughfall (TF) and stemflow (SF) in the Brazilian Cerrado. Percentages denote percent of total rainfall.

percent or total rannian.				
Location	Land cover	TF (%)	SF (%)	Source
Agudos, São Paulo Satate	"cerradão"	72.7	-	Lima and
				Nicolielo, 1983
Uberlândia, São Paulo Satate	"cerrado sensu stricto"	89.0	< 1	Lilienfein and
				Wilcke, 2004
Assis, São Paulo Satate	"cerrado sensu stricto"	95.0	0.7	Honda, 2013
Assis, São Paulo Satate	"cerrado sensu stricto denso"	89.0	1.5	Honda, 2013
Assis, São Paulo Satate	"cerradão"	80.0	2.4	Honda, 2013
Itirapina, São Paulo Satate	"cerrado sensu stricto denso"	81.2	1.1	Present study

Table 5. Average evapotranspiration for PDG and IAB sites.

Evapotranspiration (ET)	PDG	IAB
ET full period (mm d ⁻¹)	2.31 ± 0.87	2.30 ± 0.67
ET wet season (mm d ⁻¹)	2.81 ± 0.57	2.60 ± 0.38
ET dry season (mm d ⁻¹)	1.70 ± 0.70	1.91 ± 0.60
Annual ET (mm yr ⁻¹)	822	823

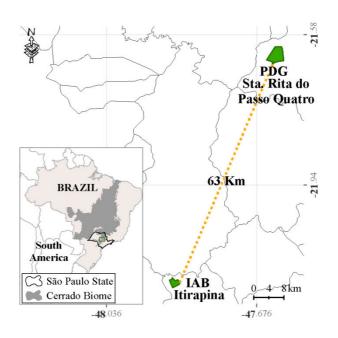


Figure 1. Location of study areas.

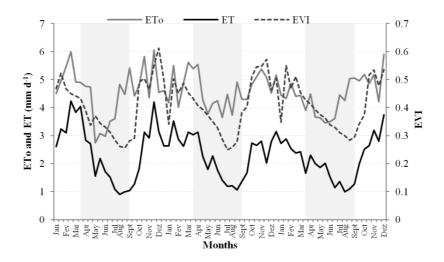


Figure 2. Seasonality of enhanced vegetation index (EVI), reference evapotranspiration (ETo) and observed actual evapotranspiration (ET) data from 2001 through 2003 at the PDG site. The grey shaded bars show the dry seasons.

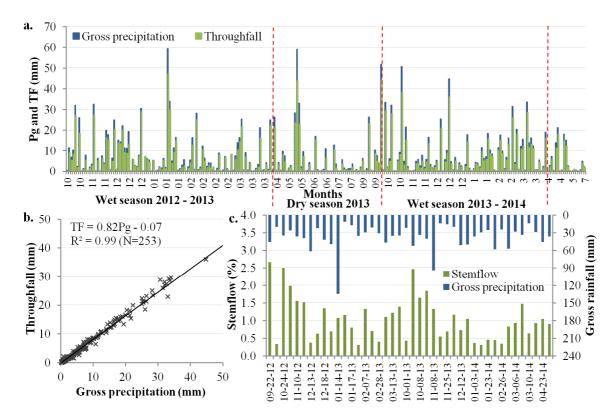


Figure 3. a. Gross precipitation and throughfall for each rain event measured from October, 2012 through July, 2014. Dotted lines in red show the beginning and the end of dry seasons (April through September). **b.** Scatter plot of throughfall against gross precipitation. **c.** Gross precipitation and stemflow measured from September 2012 through May 2014.

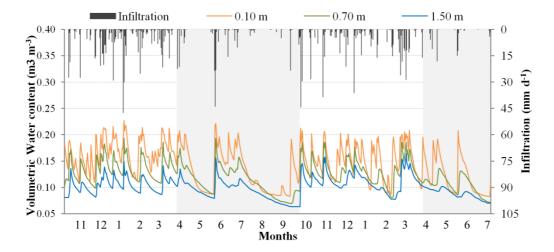


Figure 4. Estimated infiltration and volumetric water content measured at the depth of 0.10 m, 0.70 m, and 1.50 m. Data were collected from October 2012 through July 2014. The grey shaded bars show the dry seasons.

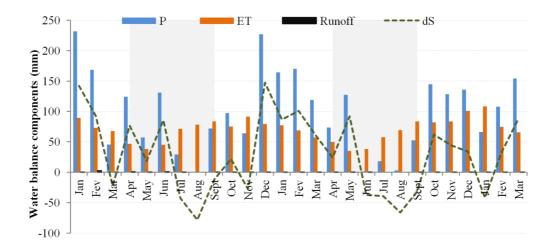


Figure 5. Water balance components at monthly scale from January 2012 through March 2014. The grey shaded bars show the dry seasons. *P* is precipitation, *ET* is evapotranspiration, and dS is soil water storage.