- 1 Journal: Hydrology and Earth System Sciences
- 2 Manuscript type: Research Article
- 3
- 4 Impacts of high inter-annual variability of rainfall on a century of
- 5 extreme hydrological regime of northwest Australia
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1 Abstract

2 Long-term hydrological records provide crucial reference baselines of natural variability 3 that can be used to evaluate potential changes in hydrological regimes and their impacts. 4 However, there is a dearth of studies of the hydrological regimes for tropical drylands 5 where intraseasonal and interannual variability in magnitude and frequency of precipitation 6 are extreme. Here, we sought to identify the main hydroclimatic determinants of the 7 strongly episodic flood regime of a large catchment in the semi-arid, subtropical northwest 8 of Australia and to establish the background of hydrologic variability for the region over the 9 last century. We used a monthly sequence of satellite images to quantify surface water expression on the Fortescue Marsh, the largest water feature of inland northwest Australia, 10 from 1988 to 2012. We used this sequence together with instrumental rainfall data to build 11 12 a statistical model with multiple linear regression and reconstruct monthly history of floods 13 and droughts since 1912. We found that severe and intense regional rainfall events, as 14 well as the sequence of recharge events both within and between years, determine surface water expression on the floodplain (i.e., total rainfall, number of rain days and 15 carried-over inundated area; $R_{adi}^2 = 0.79$; p value < 0.001, $E_{BMSP} = 56$ km²). The most 16 severe reconstructed inundation over the last century was in March 2000 (1000 km²), 17 which is less than the 1300 km² area required to overflow to the adjacent catchment. The 18 19 Fortescue Marsh was completely dry for 32% of all years, for periods of up to four consecutive years. Extremely wet years (seven of the 100 years) caused the Marsh to 20 21 remain inundated for up to 12 months; only 25% of years (9% of all months) had floods of greater than 300 km². The prolonged, severe and consecutive yearly inundations between 22 23 1999 and 2006 were unprecedented compared to the last century. While there is high inter-annual variability in the system, if the frequency and intensity of extreme rainfall 24 25 events for the region were to increase (or be similar to 1999–2006), surface water on the 26 Marsh will become more persistent, in turn impacting its structure and functioning as a 27 wetland.

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1 1 Introduction

2 Quantifying the hydrological response to changes in rainfall patterns remains challenging 3 in arid environments, especially for remote tropical and minimally-gauged drylands such as 4 the Pilbara region of northwest Australia. Tropical drylands are often characterised by 5 extreme hydroclimatic conditions, where rainfall is highly heterogeneous in its distribution 6 and the majority of streams and rivers are ephemeral but highly responsive to intense 7 rainfall events. For example, peak surface flow rates generated from ephemeral rivers and 8 creeks in the Pilbara can reach thousands of cubic metres per second after such events 9 (WA Department of Water, 2014). These factors contribute to high spatial and temporal heterogeneity of recharge-discharge mechanisms across any one catchment, which in turn 10 presents considerable challenges for predicting of resultant impacts of hydroclimate 11 12 change on catchment hydrology. Several lines of evidence suggest the Pilbara has been particularly wet during the late 20th century (e.g., Cullen and Grierson, 2007; Shi et al., 13 14 2008: Taschetto and England. 2009: Fierro and Leslie. 2013) and that the frequency of extreme precipitation events may be increasing (e.g., Gallant and Karoly, 2010). However, 15 there is no consensus on whether the observed higher summer rainfall can be attributed to 16 an overall 'wettening trend' or whether the recent 'wet' period may be a feature within the 17 18 range of natural 'extreme' variability characteristic of this region. The consequences of 19 intensification and shifts in frequency of the hydrological cycle as well as greater variability of precipitation patterns have already been documented in other parts of the world, 20 21 including alterations in the seasonality and extent of floods or drought (Harms et al., 2010; 22 Feng et al., 2013).

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24 Ecological disturbances such as flood and drought cycles are usually described by their 25 extent, spatial distribution, frequency (or return interval), predictability and magnitude 26 (i.e., severity, intensity and duration) (White and Pickett, 1985). Determining how altered 27 hydrologic regimes (floods and droughts) may in turn impact vulnerable ecosystems, 28 including wetlands, requires detailed understanding of the links between the distribution of 29 precipitation and flows across multiple spatial and temporal scales (e.g., Kiem et al., 2003; 30 Kiem et al., 2004; Verdon-Kidd and Kiem, 2009; Ishak et al., 2013). The Pilbara region of 31 northwest Australia, in common with other hot arid regions of the world including the Indian 32 Thar, Namib-Kalahari and Somali deserts, is characterised by some of the most variable 33 annual and inter-annual rainfall patterns on the planet (van Etten, 2009). In the Pilbara,

1 tropical cyclones and other low-pressure systems forming off the west Australian coast in 2 the tropical Indian Ocean often result in severe flooding events (WA Department of Water, 3 2014). These events punctuate years of prolonged drought, which together define the 4 "boom-bust" nature of productivity in highly variable desert ecosystems (McGrath et al., 5 2012). Surface water availability or persistence of water features, physical disturbances 6 and hydrological connectivity resulting from this highly dynamic regime in turn play a 7 central role in shaping aguatic and terrestrial ecosystem processes, species life history 8 strategies and interactions and population dynamics (Box et al., 2008; Leigh et al., 2010; 9 Pinder et al., 2010; Sponseller et al., 2013). Changes in hydroclimatic patterns and 10 extremes that might alter the natural disturbance regime would thus have profound 11 consequences for the structure and functioning of often highly specialised and adapted 12 arid ecosystems (Newman, 2006; Leigh et al., 2010).

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Remote sensing has proven to be the most suitable and often only tool for investigating 14 15 spatial and temporal variability of arid zone remote wetlands in the arid zone (e.g., McCarthy et al., 2003; Bai et al., 2011; Thomas et al., 2011) and improved understanding 16 of ecohydrological processes at the regional scale particularly (Gardelle et al., 2010; Haas 17 18 et al., 2011; McGrath et al., 2012). High temporal resolution is also needed to accurately 19 characterise the seasonal cycles and mechanisms that generate the complex spatial and 20 temporal patterning of floods at basin and regional scales and to effectively address the 21 consequences of changes in disturbance regimes for different ecosystems. For example, 22 satellite imagery has recently been successfully combined with hydrological modelling to 23 extend wetland flood regime records from tropical Australia (e.g., Karim et al., 2012) and to 24 investigate mechanisms such as connectivity among floodplains (e.g., Trigg et al., 2013). 25 Similar approaches have also been used to understand the evolution of daily flood and dynamics of floodplain vegetation on the east coast of Australia (Powell et al., 2008). 26 27 Remote sensing techniques have also been utilised to calibrate hydraulic models of 28 dynamic flow processes during floods, albeit over relatively short time periods (e.g., Bates, 29 2012; Neal et al., 2012; Wen et al., 2013). However, flood regime analyses based solely 30 on remotely-sensed data do not adequately capture the lengthy temporal scales of flood 31 and drought cycles in many arid and semi-arid regions, which require calibration periods 32 that encompass variability at interannual, decadal and multidecadal scales, especially to 33 elucidate relationships with climatic drivers and geomorphological processes (Roshier et

1 al., 2001; Mori, 2011; Ishak et al., 2013; Kiem and Verdon-Kidd, 2013).

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3 Here, we sought to identify the main hydroclimatic determinants of flooding regimes at the 4 catchment scale and to establish the background of variability of surface water expression 5 over the last century in the semi-arid northwest of Australia. First, we identified the main 6 rainfall variables that influence surface water expression on the Fortescue Marsh, the 7 largest internally draining wetland in the Pilbara region (Fig. 1), by combining monthly 8 satellite imagery from the Landsat archive with instrumental data from 1988–2012 via 9 statistical multivariate linear modelling. Second, we used the model to extend the flooding 10 regime record of the Marsh to the 1912–2012 period based on instrumental records of 11 rainfall. The development of this high-resolution temporal series allowed us to explore and 12 better understand the factors governing surface water expression in a semi-arid landscape 13 at multiple temporal scales, and particularly the significance of extreme events. These 14 larger temporal windows are needed to better understand long-term functioning of arid 15 zone wetlands such as the Marsh but more broadly to establish improved context for more informed water management strategies in these sensitive regions. 16

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19 **2 Methods**

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21 2.1 Study site – the Fortescue Marsh

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23 The Fortescue Marsh (hereon referred to as the Marsh; Fig. 1) is an ephemeral wetland of 24 some 1300 km², which is comprised of a complex network of riverine floodplains and 25 freshwater and floodplain lakes. The Marsh is the largest wetland of inland northwest Australia and formally recognised as nationally significant for its ecological and hydrologic 26 27 values (Environment Australia, 2001; McKenzie et al., 2009; Pinder et al., 2010). 28 Vegetation across the Marsh is dominated by salt-tolerant chenopod (*Tecticornia*) 29 shrublands, with eucalypt and Acacia woodlands growing adjacent to the most permanent 30 water features (Beard, 1975). As the largest freshwater feature for hundreds of kilometres, 31 the Marsh (local aboriginal name *Martuvitha*) is also of considerable heritage significance 32 including as a key focus for aboriginal communities for more than 40 000 years and since 33 the late 1800s for early European pastoralists (Slack et al., 2009; Law et al., 2010; Barber

1 and Jackson, 2011).

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The Marsh acts as an internally draining basin for the 31 000 km² upper Fortescue River 3 4 catchment (21–23°S; 119–121°E; Fig. 1), which is physiographically separated from the 5 Lower Fortescue River catchment by the Goodiadarrie Hills (> 410 m a.s.l.; 6 www.water.wa.gov.au). The upper Fortescue River is the main drainage of the catchment, 7 flowing north to northwest into the wetland system. However, numerous ephemeral creeks 8 on the southern and northern flanks of the Fortescue Valley (Fig. 1) discharge to the 9 marsh directly (www.water.wa.gov.au; Table A1). Flow in the Fortescue River is 10 characterised as "variable, summer-dominated and extremely intermittent" (Kennard et al., 2010), and only very large rainfall events generate continuous flow, which contrasts with 11 12 the normally dry stream beds of the dry season (WA Department of Water, 2014). Only 13 one official daily stream gauging station is currently operational on the river (>100 km 14 upstream of the Marsh). Other stations were only installed along the main creeks in two of the 13 sub-catchments of the Upper Fortescue River catchment (Fig. 1), and records did 15 not overlap consistently in time (Table A1). Recently, sub-daily gauging stations were 16 installed along Coondiner Creek and sections of Weeli Wolli Creek with pluviographs and 17 18 used to implement stable isotope water balance models for these sub-catchments over 19 relatively short (i.e., < 6 years) time periods (Dogramaci et al., 2015). The Ophthalmia Dam, constructed on the Fortescue River at Newman in 1981 to provide the town with 20 21 drinking water, has a 32 GL capacity and receives from a relatively small and low lying 22 fraction of the catchment (14.5%) with minimal observed impact on the riverine ecosystem 23 at the mouth of the Marsh (Fig. 1; Payne and Mitchell, 1999).

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25 The Fortescue River Valley paleodrainage, eroded from the Hamersley Basin sedimentary rocks, lies between the Hamersley Range in the south and the Chichester Range in the 26 27 north, constituting the main topographical features of the eastern Pilbara (Dogramaci et al., 28 2012). The Fortescue Marsh consists of colluvial and alluvial sedimentary deposits up to 29 ~50m developed on the top of the Oakover Formation, a sequence of younger Tertiary 30 lacustrine carbonate, silcrete and mudstone rocks deposited in the Fortescue River Valley 31 (Clout, 2011). The Oakover Formation is underlain by fractured dolomite and shale of the 32 Wittenoom Formation (Clout, 2011). The recent sediments consist mainly of detrital clays, 33 iron oxides and gypsum. The alluvial and colluvial aquifers of the Fortescue Marsh are

1 frequently confined by impermeable consolidated massive clays and calcrete and silcrete 2 layers. Surface runoff is high via the steep gradients of creeks and gorges; recent tracer 3 studies from the Weeli Wolli Creek and Coondiner Creek (Fig. 1) showed that residence 4 time of water in the upper sections of the catchment was short (days to weeks) 5 (Dogramaci et al., 2015). The groundwater under the Marsh is highly saline and likely 6 developed by evaporation of floodwater and consequent recharge to underlying aguifers 7 (Skrzypek et al., 2013). The most reported permanent water feature on the Marsh is 14 8 Mile Pool, located at the mouth of the upper Fortescue River; this pool does not retain 9 water significantly diluted nor flushed by groundwater, which contrasts to other small 10 through-flow pools in upper parts of the secondary tributaries of the catchment (Fellman et al., 2011; Skrzypek et al., 2013). 11

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13 **2.2 Climate and rainfall patterns**

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15 Rainfall in the Pilbara comes from troughs, monsoonal depressions, and onshore circulations (Leroy and Wheeler 2008; Risbey et al., 2009). Over the 1912–2012 historical 16 period, the upper Fortescue River catchment received on average 290 mm yr⁻¹, of which 17 18 75% fell during the monsoonal summer (November–April) (Fig. 2a; Australian Bureau of 19 Meteorology, www.bom.gov.au/cgi-bin/silo/cli_var/area_timeseries.pl). "Meteorologically dry" years received less than 200 mm rainfall, while "wet" years received over 300 mm 20 21 (Fig. 2a), as defined by the left-skewed mode of the yearly rainfall frequency distribution 22 (35% of all years). Scattered, small-scale storms cause daily rainfall to be highly variable 23 among the 17 weather stations (Fig. 1a, Appendix A, Table 1) of the upper Fortescue River 24 catchment (www.bom.gov.au/climate/data/). Evaporation is highest during the summer and 25 generally exceeds rainfall (Skrzypek et al., 2013); average temperatures in summer range 26 between 30-40 °C, and in winter months between 24-35 °C

27 (www.bom.gov.au/climate/data/).

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Heavy summer storms and tropical cyclones often generate large floods in the major river
systems of the Pilbara, particularly on the coast, while winter rainfall is typically not
sufficient to generate surface flows (Fig. 2; WA Department of Water, 2014). Tropical
cyclones and other closed lows accounted for most of the extreme rainfall events in the
northwest of Australia over the 1989–2009 period (Lavender and Abbs, 2013). Numerous

1 historical tracks of cyclones have been recorded in the upper Fortescue River catchment 2 during the last century (www.bom.gov.au/cyclone/history/). When TC tracks were recorded 3 within a 500 km radius of the Marsh, total monthly rainfall in the catchment was 4 significantly greater (p value < 0.01) than the 1912–2012 monthly averages for no-TC 5 months (Fig. 2b). Rain intensity during TC months was also higher (17-22 mm of rain per 6 rain day) than in no-TC months (8–10 mm of rain per rain day). Not surprisingly, extremes in the rainfall record (defined here as exceeding the 95th and 99th percentile of all monthly 7 total rainfall occurrences, or $E_{x_{95}}$ and $E_{x_{99}}$, respectively) are linked to the occurrence of 8 9 tropical cyclones. In fact, half of the months falling in the Ex95 (i.e., > 104 mm rainfall · month⁻¹) recorded at least one TC (30 out of 60 months). Further, at least one TC 10 occurrence was recorded for nine out of 12 months falling in the Exag, i.e., months 11 12 recording 190-258 mm of rainfall.

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2.3 Mapping flood history based on the Landsat archive (1988–2012)

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We mapped the flood history (i.e., surface water expression) of the Marsh floodplain area 17 (~1300 km²; Fig. 1) between 1988 and 2012 from high-resolution (i.e., ca. two-week 18 19 intervals) Landsat images that captured patterns of surface water expression (see 20 Appendix A, Sect. A2 for details). The Marsh floodplain area is defined here as elevations 21 below 410ma.s.l. and within the upper Fortescue River catchment (Fig. 1). Surface water 22 features were extracted from Landsat images using an automated thresholding method in 23 ArcGIS v. 9.2 and flood areas (F_A) were calculated using Fragstats v. 4.1 (see Appendix A, 24 Sect. A2 for details). We calculated potential errors associated with using the pixel 25 resolution (30 m) of Landsat images and the thresholding approach to classify surface 26 water features (see Appendix A, Sect. A2 for details). Based on these potential errors, estimated monthly change in flood area (ΔF_A) of less than 6 km² should be considered 27 with caution. However, given the scale of variation in F_A on the Marsh (ca. 0–1000 km², 28 29 Fig. 3) this error is relatively small.

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To provide further confidence in our dataset within the estimated errors we used two 40 cm resolution digital ortho-images produced from aerial photographs taken in July 2010 and April 2012 (Fortescue Metals Group Limited, Perth, Australia), and one 5 m resolution

1 image taken in August 2004 (Landgate, Government of Western Australia), to confirm that 2 our flood areas mapped from Landsat images taken on similar dates (i.e., within one week 3 of the ortho-image dates) were within 1 pixel (30 m) of the flood area visible in the ortho-4 images (Fig. A1a). A groundtruthing expedition in the dry season (November 2012; Fig. A1 5 b, c) that noted boundaries by GPS route tracking while walking along the water edge (~1-6 2 m distance from standing water) of the Moorimoordinia Native Well and a delineation of the inundation plume in the wet season (February 2012; Fig. A1 d) by GPS route tracking 7 8 during low-altitude helicopter survey along the water plume were also conducted and 9 confirm that our thresholding method captured standing water on the Marsh (Appendix 10 A2).

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12 **2.4 Modelling floodplain wetting and drying events**

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14 **2.4.1 Statistical model development and selection**

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16 Of the 493 Landsat images processed, only 208 images (TM & ETM) were used to build a calibration dataset for statistical modelling of hydrologic change between the 1988–2012 17 18 period (Fig. 3). Following selection of the latest observation for each month (or of the first 19 observation of the next month if within the first week; n = 265), only ΔF_A between two 20 consecutive months (n = 232) that were above the estimated errors were included. As a 21 result, 160 ΔF_A values were used in the final calibration dataset. Most (70 %) ΔF_A values 22 were calculated over a ca. month-long interval (i.e., 30±7 d), but this interval ranged from 23 16 to 48 days for the full calibration dataset.

24

25 We used a multiple linear regression (in R v. 2.11.1) to identify the main climatic drivers of $\Delta F_{\rm A}$ on the Marsh and generate a predictive statistical model to reconstruct monthly $\Delta F_{\rm A}$ 26 27 for the last century (1912–2012). Climatic variables tested as predictors in the model 28 included: monthly total rainfall, number of rain days, mean temperature and potential 29 evapo-transpiration calculated from weather station records and monthly gridded datasets 30 (see Appendix B, Table A3 for details). To account for the potential effect of system 31 memory, we included F_A in the previous 1 to 12 months as predictors in the model. Initially, 32 the sensitivity of each predictor was tested and only the hydroclimatic variables that were 33 significant in explaining the variation in F_A were used in the model. The model that

provided the best fit between the predicted and observed values in the calibration set as per the coefficient of variation (R^2_{adj}) adjusted for the number of variables and the smallest root mean square error E_{BMS} was selected.

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5 2.4.2 Validation of model and 1912–2012 reconstruction

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The model's predictive accuracy was tested by both cross-validation and calculation of the E_{RMS} of prediction (E_{RMSP}). A random ten-fold cross-validation (CV) was computed using the CVIm function of the DAAG R package v. 1.16 (Maindonald and Braun, 2013). The E_{RMSP} , which indicates how well the model fits an independent subset of the data, was obtained by removing block subsets representing a third of the calibration occurrences (i.e., 1988–1997; 1998–2004; 2005–2012).

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14 We used the modelled ΔF_A to reconstruct the total area flooded (F_A) from the earliest 15 available instrumental data in the region, i.e., from March 1910 to December 2012. However, the value of F_A in March 1910 being unknown, the observed F_A minimum, 16 average and maximum of the calibration period (1988–2012) were used as starting points 17 18 and long-term statistics for the hydrological regime were calculated from the meeting point 19 of the three time series, i.e., January 1912. Yearly statistics were calculated for the rain 20 year, i.e., November–October. We used comparisons with an aerial photographic survey 21 from 1957 (Edward de Courcy Clarke Earth Science Museum, UWA), early MSS Landsat 22 imagery (1972–1988) and droughts/flood events reported by early surveyors and 23 pastoralists to local newspapers (www.trove.nla.gov.au) to provide historical anchors to 24 our 1912–2012 time series (see references in-text).

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27 3 Results and discussion

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3.1 Hydroclimatic determinants of floods and droughts

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- Total rainfall in the upper Fortescue River catchment (R), number of rain days (R_d) and
- 32 carried-over inundated area (F_{At-1}) were the strongest hydroclimatic determinants of the
- 33 monthly flooding and drying (ΔF_A) regime at the Fortescue Marsh (*p* value<0.001; Table

1). The high R_{adi}^2 (0.79, p value < 0.001) indicates that the final model included the most 1 2 important contributors to ΔF_A variation (Table A4). R alone tested independently of the 3 other variable explained 64% of the variance (p < 0.001), and including R_d , improved variance explained by only 8% (p < 0.001). Although there is some collinearity between R 4 5 and R_d (Table A4), we considered it important to include both hydroclimatic variables (R 6 and $R_{\rm d}$) from a mechanistic point of view, precisely because of the highly variable nature of 7 our system. For example, in our study system, while it is common that 200 mm may fall 8 over just two days, at other times 200 mm may fall over 28 days (www.bom.com.au). 9 These very contrasting monthly distributions of rainfall demonstrate vastly different intensities and in turn generate quite different run-off; the dynamics of rainfall in such a 10 11 highly heterogeneous climate are thus best captured by inclusion of both variables, where more R_d modulates negatively the impact of R. In addition, the inclusion of R and R_d may 12 13 account to some extent for the recorded changing rainfall intensity over the century (Shi et 14 al., 2008; Taschetto and England, 2009; Gallant and Karoly, 2010; Fierro and Leslie, 2013). The model's predictive accuracy was similar for both tests performed, i.e., the 15 E_{BMSCV} and the best E_{BMSP} = 56 km² (Table A5; Fig. A3). However, the subset model used 16 to calculate E_{BMSP} , which excluded the particularly wet and variable 1998–2004 period 17 18 from the calibration period, performed the worst at reconstructing ΔF_A for the 1998–2004 verification period ($R_{adi}^2 = 0.64$; $E_{RMSP} = 86 \text{ km}^2$), indicating this period constituted an 19 important range for the calibration of the model. Both other calibration models (excluding 20 the 1988–1997 or the 2005–2012 periods) were more accurate (E_{BMSP} = 58 and 56 km², 21 22 respectively), and the overall variance explained improved to 81 and 82% when either of 23 these dry, less variable periods was removed from the model.

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A lack of surface water is returned by the model as areas $\leq 0 \text{ km}^2$. The negative values (\leq 25 26 0 km²) for 'area' can conceptually be explained as the depletion of the groundwater 27 resources and lowering of the water table below the ground level. While our calibration 28 period captures an exceptional range of intraseasonal and interannual variability in this 29 extreme system, changes in the collinearity structure between highly collinear variables 30 may occur over time and thus affect the relative contribution of the predictors and the 31 reliability of the reconstructed estimates (Dormann et al., 2013). However, the relationship 32 between R and R_d variables appears to have remained strongly linear between equivalent 33 time periods over the reconstructed period, with only minor changes in the fit, slope and

1 intercept. Nevertheless, the coefficients of these variables should not be used outside the 2 scope of this study. Mechanistically, we do not expect the mutual influence of R and R_d on 3 surface flow, where for the same volume of rain more water flushes through the river 4 network if it occurs over fewer rain days, to have changed drastically in the semi-arid 5 region over the last 100 years, or at least not beyond the reported error of the model. 6 Hence, this reconstruction should be used to examine long-term patterns of change in 7 hydrological status and meteorological determinants as opposed to fine-grained catchment 8 processes of recharge provided by higher spatio-temporally resolved hydrological models. 9 For further details on the modelling statistics, refer to the Pearson correlation matrix for the 10 modelled variables (Appendix A, Table A6) and the distribution of observed against reconstructed ΔF_A values (Appendix A, Fig. A2). 11

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13 The goodness-of-fit and relatively small errors of the model provide confidence in the 14 reconstruction starting in the early 1900s. While our calibration period captures an exceptional range of intraseasonal and interannual variability in this extreme system. 15 changes in the collinearity structure between highly collinear variables may occur over time 16 17 and thus affect the relative contribution of the predictors and the reliability of the 18 reconstructed estimates (Dormann et al., 2013). However, the relationship between R and 19 R_{d} variables appears to have remained strongly linear between equivalent time periods 20 over the reconstructed period, with only minor changes in the fit, slope and intercept (Fig. 21 A4). Nevertheless, the coefficients of these variables should not be used outside the scope 22 of this study. Mechanistically, we do not expect the mutual influence of R and R_d on 23 surface flow to have changed drastically in the semi-arid region over the last 100 years, 24 where for the same volume of rain more water flushes through the river network if it occurs 25 over fewer rain days, or at least not beyond the reported error of the model. Hence, this 26 reconstruction should be used to examine long-term patterns of change in hydrological 27 status and meteorological determinants as opposed to fine-grained catchment processes 28 of recharge provided by higher spatio-temporally resolved hydrological models. However, 29 the model tended to underestimate ΔF_A following very intense rainfall events (large rainfall 30 over 1–3 days), which might be partly attributed to the monthly resolution (Appendix A, Fig. 31 A2). Reconstructed values of ΔF_A for any given month are calculated for the last day of the month and as such do not account for the timing of intense events during that month. A 32 33 large rainfall event that occurred early in a month would thus result in smaller ΔF_A than a

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1 large rainfall event that occurred later in a month. The underestimation of ΔF_A values during intense events might also be due to the high spatial heterogeneity of rainfall in the 2 3 catchment, which was readily apparent when events were much larger closer to the Marsh 4 (e.g., Marillana Station, Fig. 1b; www.bom.gov.au/climate/data/). Consequently, our time 5 series mostly reflects regional-scale events rather than more localised events. The use of 6 weighted contributions of the different meteorological stations or sub-catchments within the 7 upper Fortescue River catchment might improve the downscaling of this model. However, 8 the instrumental records in this region are both temporally and spatially patchy, and using 9 higher resolution gridded data would not necessarily truly improve the resolution of the 10 data evenly for the last century (Fig. 1; www.bom.gov.au/climate/data/).

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12 Severe and intense rainfall events (i.e., high R and low R_d) clearly drive the hydrologic regime of this system over the last century. Total rainfall contributed most (R_{β} = 145 km²; p 13 value < 0.001) to monthly flooding of the Marsh (ΔF_A). More than 75mm rain \cdot month⁻¹ in 14 the catchment systematically caused a net wettening (increase in F_{A}) of the Marsh's 15 floodplains while <30 mm rain month⁻¹ was generally insufficient to impact on F_A (Fig. 4). 16 However, more intense rainfall events resulted in much larger flooding episodes. 17 18 Conversely, for the same total rainfall, more rain days in the month strongly dampened the extent of floods ($R_{d\beta} = -63 \text{ km}^2$; p value < 0.001). These "flash floods" drive the current 19 hydrological regime of the Marsh but are also consistent with the hydrochemical evolution 20 21 and modern recharge of shallow groundwater under the Marsh (Skrzypek et al., 2013). By 22 washing down of surface salts deposited on the Marsh during previous evaporation 23 episodes, large floods not only recharge the system, but also deliver freshwater that 24 becomes available at the surface for extended periods of time. This heavy rainfall (as 25 opposed to groundwater) driven system is rather unusual in the arid zone, where many 26 wetlands are groundwater-dominated, playa-like ecosystems (Bourne and Twidale, 2010; 27 Tweed et al., 2011). In arid zone playas, the hypersaline groundwaters from the deep 28 aquifer are connected to surface processes and result in saline waters being exposed 29 (Bourne and Twidale, 2010; Cendon et al., 2010). In contrast, our results support that the 30 Fortescue Marsh is rather a paleosaline lake where vegetation can grow and surface water 31 is largely fresh, but then eventually becomes brackish due to the concentration of solutes 32 with time owing to evaporative losses.

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1 The sequence of events (i.e., F_{At-1}), or the "system memory", was also an important 2 determinant of surface water availability on the Marsh. Water loss $(-\Delta F_A)$ on the Marsh 3 from one month to the next was larger over a months after higher inundation extent (F_{At-1} > 0 km²). For example, after a large flood area increase of 560 km² in August 1942, the 4 water extent decreased by 100 km² over the first month. In contrast, an extent of 200 km² 5 in May 1912 decreased by only 50 km² over the first month, despite a lack of rain in both 6 7 cases. Because of the negative value of the F_{At-1} coefficient, this variable was not only 8 significant in predicting ΔF_A (Table 1), but also enabled the reconstruction of continuous 9 values for F_A over the last 100 years from ΔF_A by accounting for the "maximum drying" 10 capacity" of the system, where F_A became otherwise progressively more negative with time. The F_A values are not just correlated with F_{At-1} , they are unevenly limited (i.e., 11 12 biased) within a certain range when ($F_A < 0$) dependent on F_{At-1} . F_{At-1} hence acts as a 13 weighting variable to account for the size-dependent range of possible values of change in F_{A} . Intervals (Int) between observations (number of days over which the change was 14 observed) did not significantly improve the fit of the model ($Int_{\beta} = -8 \text{ km}^2$; p value= 0.07). 15 16 This variable (Int) was nevertheless included in the model to account for ΔF_A values being 17 calculated over slightly different time intervals (i.e., 30±7 d) in the calibration period and 18 because months of the year include 28 to 31 days. This, Int acted as a constant that 19 contributed to explaining the decrease of surface water every month. Monthly loss of 20 surface water on the Marsh through evaporation and transpiration was reconstructed to be up to 150 km² (i.e., lowest ΔF_A). The most severe water losses occurred during especially 21 22 dry April, May and June (i.e. <3.5 mm rainfall; Fig. 4) following very wet summers. 23 Unsurprisingly, cumulative severe floods resulted in the longest inundation periods 24 recorded on the Marsh, and often contributed to the following year's hydrological status. Over the 1912–2012, 32% of years had up to 400 km² (40% fullness) surface water 25 26 expression carried over to the next year (i.e., winter to summer). In contrast, 68% of years 27 ended with no surface water and depleted aquifers in October (Fig. 5b). 28 29 Our findings indicate that the reconstructed total area flooded at the Marsh represents an 30 integrated ecohydrological catchment response to rainfall, which is expected from such 31 terminal basins (Haas et al., 2011). We observed that the impact of rainfall on inundations

32 and droughts is at least in part modulated by the high local evaporation rate (five to ten-

fold greater than rainfall), which acts as a constant drying force on the surface water even

1 though temperature or potential evapotranspiration (PET) did not significantly improve the 2 fit of the model. In addition, vegetation in drylands typically shows a rapid increase in 3 productivity in the few months following a large rainfall event (e.g., Veenendaal et al., 4 1996; McGrath et al., 2012); thus, runoff from subsequent events might be dampened 5 through enhanced physiological (plant water) use, which is in turn consistent with the 6 negative effect of F_{At-1} on flood area change (Table 1). We suggest that expected seasonal 7 and interannual variation in temperature and/or PET were thus largely accounted for 8 through the use of F_{At-1} and the constant *Interval* variables.

9

10 **3.2 Spatial and temporal patterns of inundations**

11

12 Our monthly reconstruction reveals that the floodplains of the Fortescue Marsh have had 13 extremely variable interannual severity of total flooded area (F_{Amax}) that in turn determined 14 the duration of inundations for the last century (Fig. 3). Of the last 100 years (1912–2012), almost 25% were large flood years, i.e., years for which the maximum flood area (F_{Amax}) 15 16 was over 300 km² (Fig. 3b). Large inundations typically occurred as a result of one to three-month long flood pulses in the austral summer (February-April). As described 17 18 earlier, these flood pulses were mainly associated with regional hydroclimatic events such 19 as one or more TCs occurring in the austral summer (January-March), and are major 20 drivers of surface water expression at the Marsh for the last century. Following large 21 floods, some level of inundation could be maintained for over 12 months in 7% of years 22 (Figs. 6 and 7). Further, only large flood years generated substantial > 0.5m depth of 23 surface water (Fig. 8a), which would also have the potential to completely submerge the 24 vast chenopod community on the Marsh (Beard, 1975). These large flood years, their 25 consequent supra-seasonal sustained inundations and their connectivity to the western 26 sections (downstream) have been relatively frequent over the last century and reflect the natural variability in the hydroclimatic regime. On the other hand, > 800 km² flood vears 27 28 (only two in the past 100 yr, 1942 and 2000) are considered extreme, infrequent 29 disturbances, which bring exceptional volumes of freshwater to the system (Fig. 8b). The 30 most striking effect of the interannual system memory was observed between 1999 and 31 2006, the period during which inundation extent and duration on the Marsh were above 32 average and unprecedented for the last century. The longest period in the last 100 years in which surface water was consistently present on the Marsh (i.e., $F_A > 0 \text{ km}^2$) was from 33

1 1998 to 2002, including the largest yearly inundation for the entire century in March 2000 $cf \sim 1000 \text{ km}^2$ (Fig. 3c).

3

4 In addition to the large flooding events described above, the majority of years (70–79 %) 5 experienced at least one month of inundation resulting from smaller floods ($F_{Amax} < 40-48$ 6 km²) (Figs. 6 and 7) that in turn also influenced the distribution and connectivity of surface water within the different sections of the Marsh (Fig. 6). During large or severe inundation 7 8 years, the entire floodplain became initially one connected flood area (Fig. 6). Following 9 such an event in 1934, pastoralists experienced the "Marsh becoming a [400 km²] large 10 lake" (Fig. 6; Aitchison, 2006). Going into the winter months, evaporation and a lack of 11 significant input from rainfall events typically resulted in drying and progressive formation 12 of disconnected pools mainly along the northern shore and eastern end of the Marsh (Fig. 13 6). Based on our 25 yr calibration period, similarly severe years resulted in spatially 14 consistent patterns of interannual inundation during both wetting and drving phases (Fig. 15 6). While guite frequent, large flood years do not occur at regular intervals, conferring a poor predictability to surface water in the system. The lowest recurrence was prior to 1960, 16 17 with up to 14 years between two events; post 1960, large events have occurred at 18 intervals of seven years or less, which in turn has resulted in more severe and prolonged 19 inundations e.g., between 1999 and 2006.

20

21 The near yearly recurrence of severe and prolonged inundations over the 1999–2006 22 period in our record is unprecedented relative to the previous 80 or so years and 23 consistent with the heavier summer rainfall events observed in the region over the recent 24 decades (e.g. Shi et al., 2008; Taschetto and England, 2009; Gallant and Karoly, 2010; 25 Fierro and Leslie, 2013). However, rigorous analysis of periodicities would be required for 26 the appraisal of potential multi-decadal trends in the hydrological regime a background of 27 such high variability (e.g., Kiem et al., 2003; Kiem et al., 2004; Verdon-Kidd and Kiem, 28 2010; Ishak et al., 2013). In fact, future investigations and risk analyses in the region 29 should strive to assess the potential influence of known larger-scale climatic drivers and 30 their interactions with each other on intraseasonal and interannual hydroclimate variability 31 in the northwest of Australia (e.g., Kiem and Frank, 2004; Pui et al., 2011; Kiem and 32 Verdon-Kidd, 2013), such as El Niño-Southern Oscillation, the Indian Ocean dipole, the 33 Madden Julian oscillation and the southern annular mode (Risbey et al., 2009; Fierro and

Leslie, 2013). The development and application of high-resolution proxy indicators of past
 hydroclimatic changes for the arid zone could also provide more robust insights on multi decadal trends and ecosystem vulnerability to these changes (e.g., Cullen and Grierson,
 2007).

5

6 **3.3 Significance of predictability and persistence of drought**

7

8 Our reconstruction shows that the Fortescue Marsh floodplains have more often been dry (i.e., where no surface water is evident on the Marsh, or $F_A \le 0 \text{ km}^2$) than wet over the last 9 10 century (Fig. 3c). Hydrological droughts (i.e., series of consecutive months where $F_A \leq 0$ km²) of at least one year were frequent (21 %) between 1912–2012 (Figs. 3c d, and 7). 11 12 The most recent drought that persisted for more than two years occurred between 1990 13 and 1993 (3.2 years). In contrast, particularly extended drought periods were more 14 frequent between the late 1930s and early 1960s, with the longest supraseasonal drought on record lasting 4.3 years (between 1961 and 1965). In such water-restricted and remote 15 16 environments, early pastoralists would have been the first to notice changes in the distribution and availability of freshwater. Reports of "bad drought" on Roy Hill Station in 17 18 early 1939 and winter of 1940, where "no feed" for cattle was available (Aitchison, 2006) 19 corroborate our reconstruction. Dramatic vegetation changes were also documented on 20 the Marsh's floodplain during this dry period (1938–1940), which coincided shortly after 21 with Marillana Station shifting from cattle to sheep farming (Aitchison, 2006). In our time 22 series, this documented drought corresponded to minimal surface water ($F_{Amax} < 150 \text{ km}^2$) 23 at the Marsh due to the occurrence of only minor flood events over these years (Fig. 3c). A 24 20-month period between 1918 and 1919 where F_A at the Marsh was reconstructed as less than 0 km² in our analysis also corresponds to a report by the Roy Hill Pastoral 25 26 Company, one of the main pastoralist in the upper Fortescue River catchment, as a 27 "severe drought" causing the installation of "10 new wells" in 1919 (Dept Lands and 28 Surveys, 1919) (Fig. 3c).

29

Overall, the eastern section of the Marsh experienced the least interannual variability by holding the most reliably inundated freshwater areas (Fig. 6), consistent with the presence of long-lived trees at 14 Mile Pool and Moorimoordinia Native Well (Beard, 1975). The September 1957 aerial photograph also shows these pools partially filled even though

1 there was little summer rain that year, also corroborating our reconstruction of a dry period 2 at that time. These more permanent, shallow water features were restricted to the 3 floodplains at the mouth of the upper Fortescue River and other smaller tributaries draining 4 the steeper slopes of the Chichester Range to the north (Fig. 6). These sections have thus 5 been under a more localised and "high" inundation frequency regime from smaller events 6 (Thomas et al., 2011; Fig. 6). These sequential, smaller events potentially maintain refugia 7 for aquatic populations, which may facilitate recolonisation of other parts of the Marsh 8 following the larger, less frequent flood disturbances that in turn effectively "reset" arid 9 zone ecosystems (Leigh et al., 2010; Stendera et al., 2012). With such spatial variation in 10 flood frequency, we can also expect vegetation communities on the Marsh to form mosaics 11 tightly linked to their different water requirements and tolerances, as has been seen on 12 other floodplains such as those of the Macquarie Marshes in central-eastern Australia 13 (Thomas et al., 2011).

14

15

16 **4 Conclusions**

17

18 We developed a reliable model to predict and characterize the surface water response of a 19 major regional wetland to hydroclimatic variability over the last century. Our approach is 20 readily applicable to extend the temporal record to other ephemeral water bodies. Through 21 greater understanding of system responsiveness to regional rainfall patterns, we also now 22 have improved capacity to assess the long-term ecohydrological functioning of arid 23 floodplains. For example, if current rainfall trends are sustained, increased flooding of the 24 Fortescue Marsh will prolong the inundation period in the year, the connectivity between 25 the different parts of the Marsh and the river network and increase the carry-over for the following year. The resulting enhanced persistence may in turn affect long-term 26 27 hydrochemical and ecological processes of the system, e.g., by an increase in surface 28 water salinity.

- **1** Appendix A: Mapping the flood history
- 2

3 A1 Landsat archive/image selection

4

5 The flood history of the Fortescue Marsh was reconstructed using standard terrain 6 corrected scenes for systematic radiometric and geometric accuracy (Level 1T) from the 7 USGS EarthExplorer Landsat archive (http://earthexplorer.usgs.gov/). The Landsat archive 8 has seasonal to monthly coverage of the Fortescue Marsh from 1972–1988 and fortnightly 9 coverage from 1988–2012. We guantified water coverage, or total flooded area (FA) from 10 a subset of 493 satellite images with the analysis of wavelengths sensitive to water 11 reflectance (Xu, 2006), specifically the short-wave infrared (SWIR) or mid-infrared (MIR) 12 radiation bands 5 (TM, ETM) and 3 (MSS). All image processing was conducted using 13 ArcGIS v.9.2 and ERDAS Imagine 2011. Pixel resolution was 30m x 30m (900 m²) for the 14 observation period (1988-2012).

15

16 A2 Flood area delineation and error

17

18 Water features were relatively straightforward to extract using a simple automated 19 thresholding method (Xu, 2006), owing to their very high contrast to the surrounding arid landscape. F_A could not be estimated using our automated method when partial cloud 20 21 cover was present in the satellite imagery, or for the ETM-SLC off series of Landsat 7 (169 22 images from a total of 493). Therefore, F_A was estimated in these years by calculating the 23 midpoint between the most recent "before and after" F_A estimates. This approach also 24 allowed us to capture the largest F_A estimates as they were often partly obstructed by 25 clouds.

26

To account for registration error across the temporal and satellite series, the F_A estimate and its associated error (*estimation errors*) were obtained from three water features extracted for every image using a lower, mid and upper threshold of reflectance values. The three consecutive threshold values (either 10, 20, 30, 40, 50, or 60 value of reflectance) were selected to include the highest frequency distribution of water pixels while providing the smallest F_A estimate error. We also calculated *resolution errors* in extracting F_A to account for the use of 30m x 30m pixels values. Here, we applied a 15m

1 buffer inside and outside the water-only polygon for all thresholds. Thus, *estimation* and 2 resolution errors were largest when F_A was small owing to an increase in the "edge length" 3 to size ratio, and differences in F_A less than 6 km₂ should be considered with caution. A 4 simple linear regression obtained between the automated F_A and its buffer was used to 5 calculate the resolution error for these shapes. The resolution error for shape-estimated F_A 6 was calculated using linear regression formulas obtained between F_A and inside buffer (R_2 = 0.99, p value < 0.001) and outside buffer (R^2 = 0.99, p value < 0.001). Strong congruency 7 8 between elevation contours and the shape of flooded area estimates on the Fortescue 9 Marsh indicate that our thresholding methodology accurately detected standing water. 10 Neither *estimation* nor *resolution errors* were found to follow a seasonal or overall temporal 11 trend. However, we cannot discount that areas of waterlogged ground also contributed to 12 the estimates of flooded area (Castañeda et al., 2005).

- 13
- 14

15 Appendix B: Climate variables

16

While 17 meteorological stations have been intermittently recording daily rainfall data in 17 18 the upper Fortescue River catchment, only six are currently still in operation, forming a too 19 sparse and temporally inconsistent network for direct use in this study (Fig. 1; Table A2). 20 Explanatory hydroclimatic variables were thus generated using monthly gridded datasets 21 resolved at either 0.5 or 1° cell size weighted for their relative contribution to the upper 22 Fortescue River catchment (Table A3). Total rainfall and mean temperature were obtained 23 from the Australian Bureau of Meteorology (www.bom.gov.au/cgi-24 bin/silo/cli_var/area_timeseries.pl), the Climatic Research Unit (CRU) and the Global Precipitation Climatology Centre (GPCC) via the Koninklijk Nederlands Meteorologisch 25 26 Instituut (KNMI) Climate Explorer (climexp.knmi.nl). Potential evapotranspiration (PET), 27 calculated using Penman–Monteith parameterization and based on the actual vegetation 28 cover, was from van der Schrier et al. (2013). The mean number of rain days \cdot month⁻¹ 29 $(R_{\rm d})$ was calculated from daily rainfall records obtained from the four meteorological 30 stations still in operation, located within or close to the upper Fortescue River catchment, 31 relatively evenly distributed across the vast geographic area and with the longest records 32 (i.e., Noreena Downs, Bulloo Downs, Marillana and Mulga Downs) (Fig. 1; Table A2). 33

Author contributions. A. Rouillard wrote the paper with input from P. F. Grierson, G.
 Skrzypek, C. Turney and S. Dogramaci. A. Rouillard collected and processed the satellite
 imagery and conducted the statistical analyses. A. Rouillard and G. Skrzypek developed
 the modelling approach after discussion with co-authors. The study was conceived by P.
 F. Grierson, A. Rouillard, G. Skrzypek, S. Dogramaci and C. Turney.

6

7 Acknowledgements. We thank the two anonymous referees and the Editors for their 8 constructive comments, which have helped focus and improve the quality of the 9 manuscript. This research was supported by the Australian Research Council (ARC) in 10 partnership with Rio Tinto (LP120100310). A. Rouillard was supported by the Australian Government and UWA via an International Postgraduate Research Scholarships (IPRS) 11 and an Australian Postgraduate Awards (APA), as well as by the Canadian and Québec 12 13 governments via a Natural Sciences and Engineering Research Council (NSERC) and a 14 Fonds québécois de la recherche sur la nature et les technologies (FQRNT) via graduate scholarships. G. Skrzvpek participation is supported by an ARC Future Fellowship (FT110 15 100 352). We thank the Fortescue Metals Group Ltd for access to ortho-images and Digital 16 Elevation Model of the Fortescue Marsh. The authors thank Jeremy Wallace (CSIRO) and 17 18 Victoria Marchesini (UWA) for advice on image processing, Gerard van der Schrier (KNMI) 19 for sharing the PET dataset, Gavan McGrath (UWA) for council on tropical cyclones, 20 Caroline Bird (Archae-Aus) for help with archival research and the following researchers 21 for their assistance with the modelling: Yun Li (CSIRO), Jérôme Chopard (UWA), Michael 22 Renton (UWA), Edward Cook (UColumbia), Jonathan Palmer (UTAS) and Jason Smerdon 23 (UColumbia). We also acknowledge the kind support of Lee and Sue Bickell (Marillana 24 Station), Barry and Bella Gratte (Ethel Creek Station), Victor and Larissa Gleeson (Mulga 25 Downs Station), and Murray and Ray Kennedy (Roy Hill Station). We are also grateful to Alison O'Donnell and Gerald Page for providing useful comments on the early and final 26 27 versions of the manuscript.

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- 25
- 26

- Table 1: Model parameter estimates and standardised statistics for the final linear model
- to reconstruct historical flood area on the Fortescue Marsh, NW Australia

Driver	β (km ²)	Effect	<i>p</i> value
R	144.729	+	< 0.001
R_{d}	-62.950	-	< 0.001
F _{At-1}	-29.157	±	< 0.001
Int	-7.650	-	0.070
Intercept	-8.040	-	0.816

- Note: β = Weighted contribution; Effect = gain (+) or loss (-) effect of each variable on change in flood area (ΔF_A) ; R = total rainfall · month⁻¹ on the upper Fortescue; R_d = number of days with > 0 mm of rain · month⁻¹; F_{At-1} = flood area of the previous month; Int = the time interval between observations; Intercept = equation 6 7 intercept

- **Table A1**: Temporal coverage of all official stream gauging stations in the Upper
- 2 Fortescue River catchment and maximum recorded daily discharge

Site number	Stream Name	Name	Operational date	Last measurement	Max discharge (m ³ sec ⁻¹)	Total discharge (GL)		
708001	Marillana Ck	Flat Rocks	15/08/1967	23/02/1983	1327	72		
708006	Fortescue River	Goodiadarrie Crossing	01/12/1972	01/10/1986	*	*		
708008	Fortescue River	Roy Hill	01/09/1973	29/09/1986	*	*		
708011	Fortescue River	Newman	09/01/1980	Present	1730	78		
708013	Weeli Wolli Ck	Waterloo Bore	30/11/1984	Present	4137	142		
708014	Weeli Wolli Ck	Tarina	10/05/1985	Present	2100	62		
708016	Weeli Wolli Ck	Weeli Wolli Springs	08/10/1997	14/07/2008	423	10		
Note: * Only daily stage height available; location of stations marked on Fig. 1								

- 1 **Table A2:** Australian Bureau of Meteorology (BoM) rainfall stations
- 2 (www.bom.gov.au/climate/data/) located within and nearby the upper Fortescue River
- 3 catchment, NW Australia.
- 4

No	Station name	BoM number	Lat (°N)	Long (°E)	Status	Year open	Year closed
	Mulga Downs	5015	-22.10	118.47	Open	1898	
	Bulloo Downs	7019	-24.00	119.57	Open	1917	
	Marillana	5009	-22.63	119.41	Open	1936	
	Noreena Downs	4026	-22.29	120.18	Open	1911	
1	Balfour Downs	4003	-22.80	120.86	Closed	1907	1998
2	Wittenoom	5026	-22.24	118.34	Open	1949	
3	Auski Munjina Roadhouse	5093	-22.38	118.69	Open	1998	
4	Kerdiadary	5047	-22.25	119.10	Closed	1901	1910
5	Warrie	5025	-22.40	119.53	Closed	1927	1964
6	Bonney Downs	4006	-22.18	119.94	Open	1907	
7	Poondawindie	4063	-22.20	120.20	Closed	1930	1938
8	Sand Hill	5064	-22.78	119.62	Closed	1971	1984
9	Roy Hill	5023	-22.62	119.96	Closed	1900	1998
10	Ethel Creek	5003	-22.90	120.17	Closed	1907	2003
11	Packsaddle Camp	5089	-22.90	118.70	Closed	1989	2002
12	Rhodes Ridge	7169	-23.10	119.37	Open	1971	
13	Rpf 672 Mile	4065	-22.70	121.10	Closed	1913	1947
14	Billinnooka	13029	-23.03	120.90	Closed	1960	1974
15	Jigalong	13003	-23.36	120.78	Closed	1913	1991
16	Minderoo	7172	-23.40	119.78	Closed	1913	1931
17	Newman Aero	7176	-23.42	119.80	Open	1971	
18	Capricorn Roadhouse	7191	-23.45	119.80	Open	1975	
19	Murramunda	7102	-23.50	120.50	Closed	1915	1949
20	Sylvania	7079	-23.59	120.05	Open	1950	
21	Prairie Downs	7153	-23.55	119.15	Open	1968	
22	Turee Creek	7083	-23.62	118.66	Open	1920	
23	Mundiwindi	7062	-23.79	120.24	Closed	1915	1981
24	Rpf 561 Mile	13013	-23.90	120.40	Closed	1913	1947
	Newman	7151	-23.37	119.73	Closed	1965	2003

Note: * Numbers correspond to location of stations marked on Fig. 1

- **Table A3**: Climate variables used in the development of a linear model to reconstruct
- 2 historical flood area on the Fortescue Marsh, NW Australia.

Interval	Variable	Res.	Location	Period	Source
d	R		Bulloo Downs	1917–2012	www.bom.gov.au/climate/data/
d	R		Marillana	1936–2012	www.bom.gov.au/climate/data/
d	R		Mulga Downs	1907–2012	www.bom.gov.au/climate/data/
d	R		Noreena Downs	1911–2012	www.bom.gov.au/climate/data/
m	R	1°	UF	1900–2012	www.bom.gov.au/cgi-bin/silo/cli_var/area_timeseriespl
m	R	0.5°	UF	1901–2012	GPCC V6 rain gauge precipitation dataset
m	R	0.5°	UF	1901–2009	CRU time-series (TS) version 3.10.01 (land)
m	Т	1°	UF	1910–2012	www.bom.gov.au/cgi-bin/silo/cli_var/area_timeseriespl
m	Т	0.5°	UF	1901–2009	CRU TS 3.10 (land)
m	PET	0.5°	UF	1901–2009	(Schrier et al., 2013)

 Note: Res. stands for the resolution of gridded data, d is daily weather station rainfall data, m is monthly gridded climate data, *R* is total rainfall (mm), *T* is mean temperature (°C) and PET is Penman-Monteith potential evapotranspiration index. <u>UF</u> is the upper Fortescue River catchment (31,000 km²).

- Table A4: Sensitivity analysis of the four variables included in the statistical model

Model												
R^2_{adj}	0.79											
p	< 0.001											
E_{RMS}	52											
ਕੂ Driver	Coeff	Error	р									
Driver	2.9	0.2	<0.001									
R_{d}	-18.8	2.4	<0.001									
F_{At-1}	-0.13	0.02	<0.001									
Int	-0.99	0.54	0.070									
Intercept	4.3	18.3	0.816									
Model												
R^2_{adj}	0.64			0.28								
p	< 0.001			<0.001								
 	68			96								
ep E _{RMS} Priver Priver Priver	Coeff	Error	p	Coeff	Error	p						
र्षे <i>R</i>	1.86	0.11	<0.001	-	-	-						
	-	-	-	18.29	2.29	<0.001						
F _{At-1}	-	-	-	-	-	-						
Int	-	-	-	-	-	-						
Intercept	-67.158	6.42	<0.001	-58.351	9.898	<0.001						
Model												
R^2_{adj}	0.35			0.71			0.72			0.79		
p	< 0.001			<0.001			<0.001			<0.001		
	91			61			59			52		
4 Variable bla Duine <i>B</i> B ^{BWR}	Coeff	Error	p	Coeff	Error	p	Coeff	Error	p	Coeff	Error	р
n R	-	-	· -	1.8		, <0.001	3.0		, <0.001	2.9		, <0.001
$\frac{1}{4}R_{d}$	18.0	2.2	<0.001	-	-	-	-18.8		< 0.001	-18.4		<0.001
F _{At-1}	-0.14		< 0.001	-0.13	0.02	<0.001	-		-	-0.13		< 0.001
Int	0.18			-0.57		0.370	-0.93	0.62	0.133		-	-
Intercept			0.216			0.220			0.311		6.4	<0.001
Nutri D												

Note: $R = \text{total rainfall·month}^{-1}$ on the upper Fortescue (mm); $R_d = \text{number of days with } > 0 \text{ mm of rain·month}^{-1}$ (days); $F_{\text{At-1}} = \text{flood area of the previous month}$ (km²); *Int* = the time interval between observations (days); Std Error = Standard error; p = significance level

- **Table A5:** Analysis of predictive accuracy of the final model based on the full 1988–2012
- 2 calibration period and the subsets for the 1998–2012, 1988–1997; 2005–2012, 1988–2004
- 3 periods.

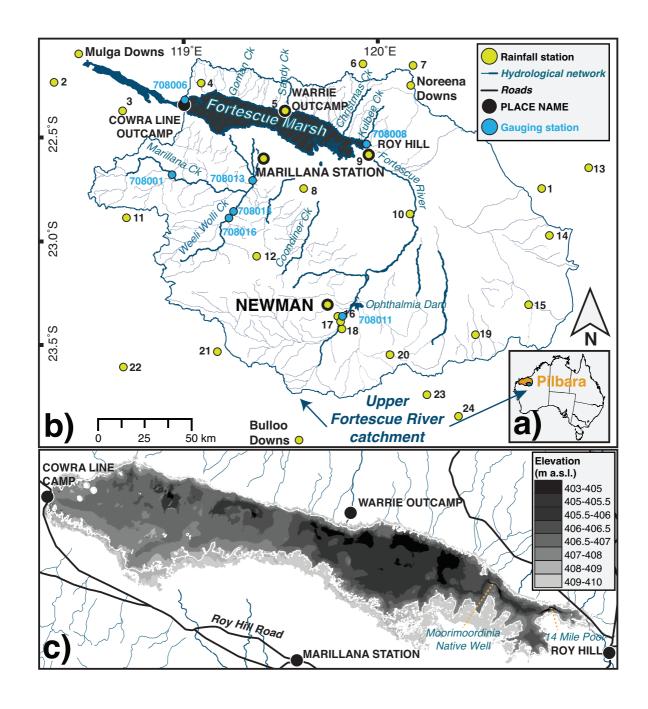
Period	1988-20	12		1998-20	12		1988-19	97; 200	5–2012	1988-20	04	
R^2_{adj}	0.79)		0.82			0.64	Ļ		0.81		
p	< 0.001			< 0.001			<0.001			<0.001		
E _{RMS}	52	2		51			46	5		56		
ERMSP	-			58			86	;		56		
Driver	Coeff	Error	р	Coeff	Error	р	Coeff	Error	р	Coeff	Error	р
R	2.9	0.2	<0.001	2.9	0.2	<0.001	2.1	0.2	<0.001	3.3	0.2	<0.001
R_{d}	-18.8	2.4	<0.001	-17.7	2.6	<0.001	-10.7	3.0	<0.001	-24.0	3.4	<0.001
F _{At-1}	-0.13	0.02	<0.001	-0.13	0.02	<0.001	-0.17	0.03	<0.001	-0.12	0.02	<0.001
Int	-0.99	0.54	0.070	-0.73	0.71	0.874	-1.24	0.61	0.044	-1.02	0.67	0.134
Intercept	4.3	18.3	0.816	-3.7	23.4	0.305	15.3	20.9	0.467	4.3	23.0	0.853

Note: R = total rainfall month⁻¹ on the upper Fortescue (mm); R_d = number of days with > 0 mm of rain·month⁻¹ (days); F_{At-1} = flood area of the previous month (km²); *Int* = the time interval between observations (days); Std Error = Standard error; p = significance level

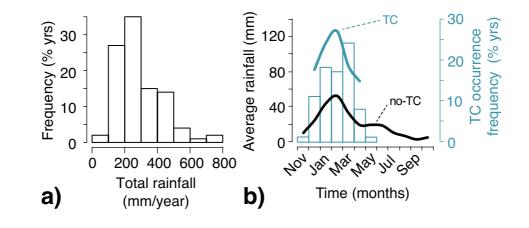
- **Table A6**: Pearson correlation matrix of the variables included in the final linear model to
- 2 reconstruct historical flood area on the Fortescue Marsh, NW Australia

	R	R_{d}	F_{At-1}	Int
R	1	-	-	-
R_{d}	0.8518 <i>p<</i> 0.001	1	-	-
F_{At-1}	-0.0361 <i>p</i> =0.6507	–0.0313 <i>p</i> =0.6943	1	-
Int	0.0703 p=0.3767	0.0089 p=0.9108	-0.0162 p= 0.8388	1

Note: R = total rainfall month⁻¹ on the upper Fortescue (mm); R_d = number of days with > 0 mm of rain month⁻¹ (days); F_{At-1} = flood area of the previous month (km²); Int = the time interval between observations (days)



3 Figure 1: The a) Pilbara region in northwest Australia, b) Upper Fortescue River 4 catchment and river network (blue lines; DoW, 2014), including the Fortescue Marsh's 5 floodplain area used in this study (black hatched section; < 410 m a.s.l. extracted from a 1 6 sec DEM-H, Geoscience Australia, 2011), stream gauging stations (blue circles, see full 7 list in Appendix A, Table A1; WIN, 2014) and meteorological stations (green circles, see 8 full list in Appendix A, Table A2; www.bom.gov.au/climate/data/) and c) elevation of the 9 study area (0.1 m vertical accuracy (RMS) LiDAR Survey DEM; Fortescue Metals Group Ltd, 2010) with roads and place name (black lines and circles; Geoscience Australia, 10 11 2001). Generated in ArcMap v. 9.2.



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- 2
- 3

4 **Figure 2**: The upper Fortescue River catchment 1912–2012 hydroclimate with **a**)

5 frequency distribution of total yearly rainfall and **b)** average monthly rainfall for months

6 recording at least one tropical cyclone (TC) within 500 km radius of the upper Fortescue

7 River catchment (blue line) and without TC recorded (black line), with the number of years

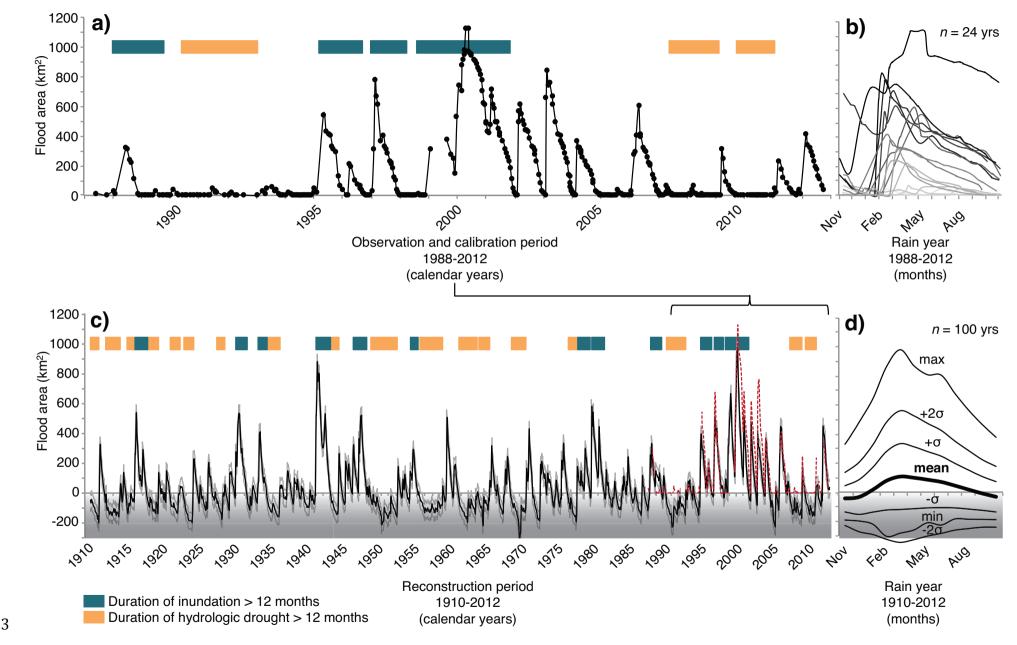
8 (frequency) where TC occurrence was recorded for each month of the water year (blue

9 columns); only one occurrence of TC was recorded in Nov and May for the last century

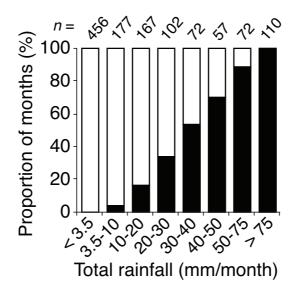
10 and thus rainfall averages for these months were not included. Source:

11 www.bom.gov.au/cgi-bin/silo/cli_var/area_timeseriespl &

12 www.bom.gov.au/cyclone/history/.



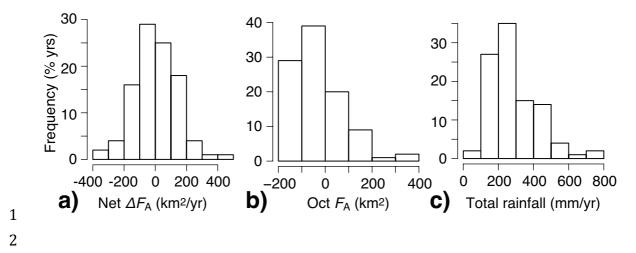
- 14
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- 16 Figure 3: 1988–2012 a) flood area observation and calibration dataset (solid black line with dots for each observation) and its b) timing of
- 17 seasonal change over the rain year (n = 24 yr); 1912–2012 **c)** flood area reconstruction (solid black line), monthly E_{RMSP} of $\Delta F_{\text{A}} = \pm 56$
- 18 km² (solid grey lines) and observation dataset plotted for the 1988–2012 period (red dashed line) for comparison and **d)** monthly mean,
- 19 minimum (min), maximum (max) and 1 and 2 σ ranges of variation over the rain year for the reconstructed period (n = 100 yr). Overlaid
- on **a**) and **c**) time-series are the supraseasonal dry and wet periods, where F_A was either < 0 km² or > 0 km² for over 12 consecutive
- 21 months. c) & d) $F_A \le 0 \text{ km}^2 = \text{no surface water evident on the Marsh (shaded).}$



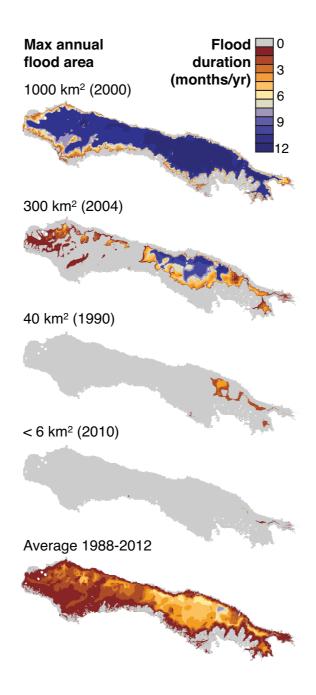
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- 2

3 **Figure 4**: Total monthly rainfall in the upper Fortescue River catchment (Total

- 4 rainfall) causing an increase in surface water area measured as the proportion
- 5 of months with net change in flood area or $\Delta F_A > 0 \text{ km}^2$ (black columns) at the
- 6 Fortescue Marsh (1912–2012).



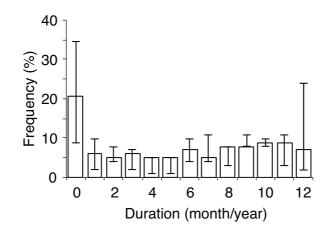
- 3 Figure 5: 1912–2012 frequency distributions of yearly a) net change in flood
- 4 area (ΔF_A) , **b)** end-of-the-year flood area (Oct F_A) and **c)** yearly maximum
- 5 flood area (F_{Amax} ; km²), n = 100 yr.



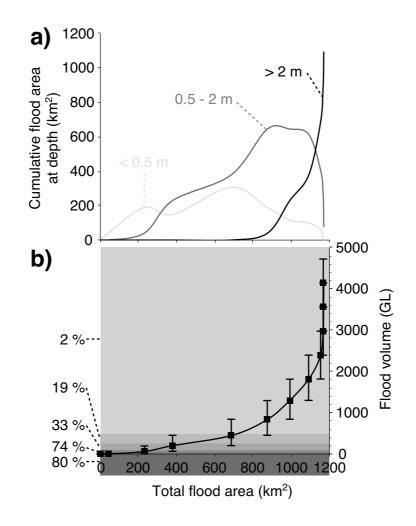
- 1
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- 4 isohyets over the rain year (Nov–Oct) representing examples of the main
- 5 connectivity thresholds: wettest year observed in 2000 ($F_{Amax} \sim 1000 \text{ km}^2$); a
- 6 very large flood year in 2004 ($F_{Amax} \sim 300 \text{ km}^2$); the long-term mean flood year
- 7 in 1990 ($F_{Amax} \sim 40 \text{ km}^2$) and a dry year in 2010 ($F_{Amax} < 6 \text{ km}^2$) and the
- 8 1988–2012 average.

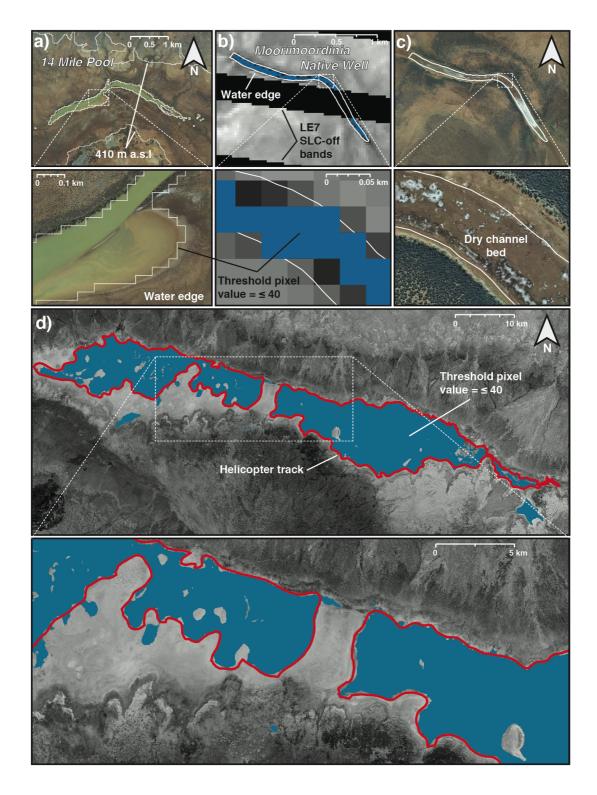


- 1
- 2
- 3 **Figure 7**: Frequency distribution of drought duration per annum (i.e.
- 4 consecutive month with $F_A < 0 \text{ km}^2$), with error bars representing the variation
- 5 in the distribution when threshold for drought duration is defined as $F_A < \pm 56$
- 6 km² for the last century (n = 100 yr).



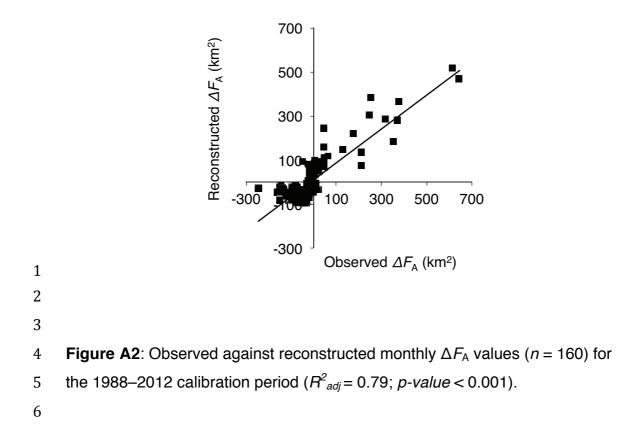
3 Figure 8: Total flood area at the Fortescue Marsh and a) its proportion

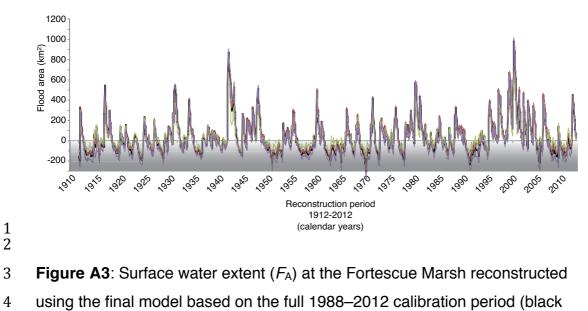
occupied by water depth shallower than 0.5 m (light grey), between 0.5 and 2
m (dark grey) and deeper than 2 m (black) and b) the volume of surface water
(black line) with century frequencies (% yr) at which different thresholds (grey
shading) were attained.



1 Figure A1: Validation and groundtruthing of standing water on the Fortescue 2 Marsh, including: a) standing water on the 14 Mile Pool extracted from Level 3 1T Landsat image (Jul 2010; solid white line = threshold pixel value \leq 40; LT5; 4 USGS) and close up against a 40-cm resolution ortho-photo (Jul 2010); 5 delineation by GPS route tracking while walking along the water edge (1-2 m 6 distance from standing water; solid white line) and close up against b) a Level 7 1T Landsat image of Moorimoordinia Native Well (Nov 2012; blue fill = 8 threshold pixel value \leq 40; LE7-SLC-off, USGS) and **c)** a RGB image showing 9 the extent of the dry channel bed (Dec 2006; SPOT-5); d) delineation of 10 standing water by GPS route tracking during a low altitude helicopter survey 11 along the water plume of the Fortescue Marsh (2012 Feb 12; solid red line) 12 and close up against standing water extracted from Level 1T Landsat image 13 (2012 Feb 14; blue fill = threshold pixel value \leq 40; corrected LE7-SLC-off; 14 USGS), overlain on a 2.5 m resolution RGB image taken during dry season 15 (Dec 2006; SPOT-5).

16





5 line) and the subsets for the 1998–2012 (red line) 1988–1997; 2005–2012

6 (green line), 1988–2004 (purple line) periods; $F_A \le 0 \text{ km}^2 = \text{no surface water}$

- 7 evident on the Marsh (shaded).
- 8

