

1 **Contribution of the Multi Attribute Value Theory to conflict**
2 **resolution in groundwater management. Application**
3 **to the Mancha Oriental groundwater system, Spain.**

4
5 **Apperl B.^{1,*}, Pulido-Velazquez M.¹, Andreu J.¹, Karjalainen T.P.²**

6 [1]{Research Institute of Water and Environmental Engineering (IIAMA), Universtat
7 Politècnica de València, Camino de Vera s/n, 46022 Valencia, Spain }

8 [2]{ Thule Institute, University of Oulu, P.O. Box 7300, University of Oulu, 90014 Oulu,
9 Finland }

10 [*]{now at: Institute of Water Management, Hydrology and Hydraulic Engineering,
11 University of Natural Resources and Life Sciences, Muthgasse 18, Vienna, Austria }

12 Correspondence to: M. Pulido-Velazquez (mapuve@hma.upv.es)

13
14 **Abstract**

15 The implementation of the EU Water Framework Directive demands participatory water
16 resource management approaches. Decision making in groundwater quantity and quality
17 management is complex because of the existence of many independent actors, heterogeneous
18 stakeholder interests, multiple objectives, different potential policies, and uncertain outcomes.
19 Conflicting stakeholder interests have been often identified as an impediment to the
20 realization and success of water regulations and policies. The management of complex
21 groundwater systems requires clarifying stakeholders' positions (identifying stakeholders
22 preferences and values), improving transparency with respect to outcomes of alternatives, and
23 moving the discussion from the selection of alternatives towards definition of fundamental
24 objectives (value-thinking approach), what facilitates negotiation. The aims of the study are to
25 analyse the potential of the multi attribute value theory for conflict resolution in groundwater
26 management and to evaluate the benefit of stakeholder incorporation in the different stages of
27 the planning process to find an overall satisfying solution for groundwater management. The
28 research was conducted in the Mancha Oriental groundwater system (Spain), subject to an
29 intensive use of groundwater for irrigation. A complex set of objectives and attributes were

1 defined, and the management alternatives were created by a combination of different
2 fundamental actions, considering different implementation stages and future changes in water
3 resources availability. Interviews were conducted with representative stakeholder groups
4 using an interactive platform, showing simultaneously the consequences of changes of
5 preferences to the alternative ranking. Results show that the approval of alternatives depends
6 strongly on the combination of measures and the implementation stages. Uncertainties of the
7 results were notable but did not influence heavily on the alternative ranking. The expected
8 reduction of future groundwater resources by climate change increases the conflict potential.
9 The implementation of the method to a very complex case study, with many conflicting
10 objectives and alternatives and uncertain outcomes, including future scenarios under water
11 limiting conditions, illustrate the potential of the method for supporting management
12 decisions.

13

14 **1 Introduction**

15 Groundwater is a vital natural resource for the reliable and economic provision of potable
16 water supply in both urban and rural environments (Foster et al., 2002). It serves as a basis for
17 life and social prosperity. The limited availability of clean water, whether it originates from
18 rivers or from aquifers, produces conflicts. Groundwater resources in La Mancha Oriental
19 suffer increasing pressure due to water abstraction for irrigation and urban water supply
20 (López Sanz, 2010). Conflicts related to large scale groundwater management develop in
21 many cases into intractable conflicts, which are typically very complex, involving many
22 parties and interests, a long history, and even strong emotions (e.g. Llamas and Martínez
23 Santos, 2005; Bromley et al., 2010).

24 To undertake a comprehensive decision-making approach, the complexity of groundwater
25 management demands needs to represent all stakeholder interests, while being understandable
26 for stakeholders in a participative context (Karjalainen et al. 2013). The guarantee of a
27 balanced use of the available water is attempted to be achieved by legislative restrictions, but
28 also by increasing the awareness and participation of society (López-Gunn and Martínez,
29 2006). With the implementation of the European Water Framework Directive, WFD
30 (European Commission, 2000), the ecological function of water becomes more relevant for
31 decision-making. To guarantee the fulfilment of the WFD while maximizing social benefits,
32 an effective management of the available resources is required. The WFD requires that

1 Member States take the necessary measures to “protect, enhance and restore all bodies of
2 groundwater”. The objective for groundwater bodies is to reach a “good” groundwater status,
3 which implies both, a good quantitative and a good chemical status. The Directive requires
4 the definition and implementation of cost-effective combinations of measures that should be
5 implemented to achieve good groundwater status. The high complexity und uncertainty
6 demands decision support tools that help the decision makers to find optimal solution by
7 assessing the trade-offs among economic, social and ecological objectives.

8 The use of Multi Criteria Decision Analysis (MCDA) can help facilitate the negotiation
9 process among stakeholders by changing their preferences towards more consensus-orientated
10 decisions (Hostmann et al., 2005; Marttunen et al 2013). A comprehensive approach is
11 required to face the multiple objectives and alternatives (Karjalainen et al. 2013;
12 Stefanopoulos et al. 2014). As MCDA’s are often technically oriented, complex and difficult
13 to understand for laymen (Kangas et al. 2008), the challenge exists in finding an evaluation
14 process of alternatives that are comprehensible for stakeholders while representing the
15 complexity of environmental decision processes.

16 Decision-making methods share common characteristics, such as the presence of multiple,
17 non-commensurable and conflicting criteria, different units of measuring among the criteria
18 and the presence of quite different alternative policies (Bogardi and Nachtnebel, 1994). This
19 work analyses the different management alternatives by focusing first on the objectives
20 (value-thinking approach) and secondly, on evaluating the alternatives using the Multi
21 Attribute Value Theory (MAVT). The MAVT represents a value measurement model in
22 which numerical scores are constructed in order to represent the degree to which one decision
23 option may be preferred over another (Keeney and Raiffa 1976). The approach has been
24 proven to provide a transparent and systematic framework to analyze problems with multiple
25 criteria and alternatives when working with stakeholders (Mustajoki et al. 2011; Karjalainen
26 et al. 2013).

27 The main objective of this work is to test the aptitude of MAVT to analyse the complex
28 system of groundwater management and to enhance conflict mediation in the Mancha
29 Oriental (MO) aquifer, in Eastern Spain. In the MO aquifer, management conflict is a long
30 lasting and complex problem in which collective actions play an essential role (Lopez-Gunn,
31 2003).

1 This study applies the MAVT approach to facilitate conflict resolution for a sustainable
2 management of the MO aquifer, and estimates the conflict potential for different management
3 alternatives considering stakeholder preferences and values. The method considers all
4 possible alternative policies, identifies different objectives and elicits stakeholder preferences
5 on the objectives. Key questions of the analyses are:

- 6 ▪ How do stakeholders rank management alternatives and are the results of our method
7 coherent with holistic rankings?
- 8 ▪ What are the principal points of conflict?
- 9 ▪ How do different future scenarios influence the preferences of stakeholders?
- 10 ▪ What are the advantages of stakeholder inclusion for a sustainable groundwater
11 management?

12

13 **2 Method**

14 The MAVT is a Multi Criteria decision analysis (MCDA) tool to solve complex real world
15 decision problems, judging different amelioration alternatives for finding a well-accepted
16 solution (Mustajoki et al. 2011, Bogardi and Nachtnebel, 1994). Different alternatives are
17 ranked by evaluating the fulfilment of set objectives. Stakeholder interviews and workshops
18 are used to elicit their preferences.

19 For the evaluation of the alternatives we adopted the additive value function (weighted sum of
20 single attribute function) (Belton and Stewart, 2002; Hostmann et al, 2005):

$$21 \quad V(A) = \sum w_i * v_i(a_i) \quad (1)$$

22 with a_i as the level of attribute i resulting from alternative A , $v_i(a_i)$ the single attribute
23 function, w_i the weights of the attributes and $V(A)$ as the total value of the alternative.

24 The value function has the vector of attribute levels that quantifies the effects of an alternative
25 as an argument and converts it into a single value that expresses the satisfaction for that
26 alternative in regard to an objective (Soncini-Sessa, 2007). The single value ranges from 0 to
27 1 for every objective and allows different objectives to be compared. The single value
28 function is unique for every stakeholder and objective. The weights express the importance of
29 each criterion compared to other criteria. The value of the weights depends on the relative
30 importance that the stakeholder associates to each attribute. The goal is to attain lumped

1 measurement of the attractiveness or utility of the outcome of a set of alternatives by
2 stakeholder.

3 **2.1 General approach**

4 The general approach of the evaluation process is shown in Fig 1. The applied approach is
5 almost similar to DAI approach (the decision analysis approach; Marttunen and Hämäläinen
6 1995, Karjalainen et al., 2013) that provides strong interactivity with the stakeholders through
7 facilitated meetings and personal computer-aided interviews of stakeholders.

8 In a very first step the problem has to be clearly defined. The second step involves a complete
9 identification of the stakeholders who are related to the problem. The snowball approach is
10 used for that purpose. A crucial step (step 3) is the identification of all stakeholder objectives.
11 Objectives are identified with stakeholders, representing their values and interests. They are
12 ordered in a hierarchy tree with different levels of detail, considering the different scope,
13 inconsistency and explicitness (Keeney & Raiffa, 1976). We arranged lower-level objectives
14 in ecological, economic and social objectives (Bogardi et al., 1982). Afterwards (step 4),
15 potential measures are defined and combined into alternatives (step 5).

16 The description of impacts is carried out through a consequence table (step 6) showing the
17 consequence that a given alternative will have on a given objective. A consequence matrix is
18 built up through various sub steps (Keeney and Raiffa, 1976). First, the physical impact of an
19 alternative to the hydrological system is analysed and quantified using measureable units.
20 Then, its socioeconomic and economic impact is quantified. To include the possible impacts
21 of climate change on preferences two impact matrices are created, where one depicts the
22 status quo of available water resources and the other one includes changes in future water
23 resources availability. In step 7, interviews are conducted to elicit stakeholders' preferences
24 which are described in detail below. The results from step 6 and 7 are combined with eq.1 to
25 rank the alternatives (step 8) among the stakeholders. Afterwards they are presented in a
26 workshop (step 9), where stakeholders are also asked to rank the measures independently of
27 any prior objectives. The results (8 and 9) are assessed and interpreted in step 10 and
28 differences in rankings of future scenarios are compared (step 11). Finally, a sensitivity
29 analysis is conducted to test the robustness of the obtained results towards uncertainties in the
30 attribute levels of the alternatives (consequence matrix) and in the attribute valuation of the
31 stakeholders.

1 **2.2 Stakeholders' Preferences**

2 Stakeholders are asked in the interviews to value different levels of attributes for every
3 objective on a standardized scale. The attribute level range must cover all the attribute levels
4 of the alternatives, limited by the highest and the lowest level of all of them (Keeney and
5 Raiffa, 1976).

6 The single attribute value function $v_i(a_i)$ is assessed with the direct rating method, which is a
7 numerical estimation method (von Winterfeldt and Edwards, 1986). The evaluation is done
8 with Microsoft Excel© on an interactive evaluation platform, so that the decider is able to see
9 a visualization of answers (Fig. 2). The range of attribute levels is limited by the highest and
10 lowest level of the attribute in all alternatives of both scenarios. The respondent is asked to
11 estimate the strength of preferences of every attribute level on a numeric scale between 1 and
12 0, with 1 as the most preferred level and 0 for the least preferred level. The remaining levels
13 have to be rated between 1 and 0, considering the space between two attributes levels as the
14 strength of preference between them. The value function is a transformation of the attribute
15 levels of the objective on a comparable scale between 1 and 0.

16 In addition to the value functions, the weights of the different objectives have to be elicited
17 for every representative. To avoid the risk of stakeholder behavioral biases it is of high
18 importance to understand the possible influences and impacts to the result of different
19 methods (Roberts and Goodwin, 2002). Different weighting methods lead to different results,
20 although they are based on the same theoretical assumptions (Pöyhönen and Hämäläinen,
21 1999). Therefore it is important to use a method which on the one hand is easily
22 understandable for the interviewed persons and on the other hand statistically applicable and
23 traceable. In this study we have applied the SWING method (von Winterfeldt and Edwards,
24 1986). In this approach the attribute ranges are explicitly incorporated in the elicitation
25 questions, being proven as a successful method on convergence tests (Pöyhönen and
26 Hämäläinen, 1999).

27 The calculation of the results was realized with Web-HIPRE© Version 1.22, a web version of
28 HIPRE 3+ software for decision analytic problem structuring, multi criteria evaluation and
29 prioritization (Mustajoki and Hämäläinen, 2000). The further processing of the data was
30 realized with Microsoft Excel©.

31 The preferences have been ranked and the different fundamental actions have been analyzed.
32 Additionally, a sensitivity analysis has been conducted to deal with uncertainty in the

1 valuations and also for uncertainty in the attribute levels. The conflict potential has been
2 interpreted and the ranking discrepancy, expressed by the mean standard deviation of the
3 average ranking of all stakeholders. The influence of the dynamic variables has been analyzed
4 by comparing the two scenarios, observing the changes of the rankings. Finally, the aptitude
5 of the MAVT method has been evaluated, comparing the results of the model with the results
6 of the holistic alternative ranking, which have been evaluated by some catch questions in the
7 interviews.

8

9 **3 The case study**

10 **3.1 Description**

11 The Mancha Oriental (MO) Aquifer is located in south-eastern Spain at the eastern part of the
12 Mancha plain, mainly in the provinces of Albacete and Cuenca, in the Castilla- La Mancha
13 region (central Spain), with small areas in the Valencian Community and Murcia (see Fig. 3).
14 It is part of the Júcar river basin district as one of its 52 groundwater bodies, with a total
15 extension of approximately 8,000 km² (CHJ, 2009) and is consequently the biggest aquifer of
16 this system and one of the largest carbonate aquifer systems in Spain. The major part of the
17 system belongs to the catchment of the Júcar River, which is strongly connected with the
18 aquifer.

19 The region is characterized by a semiarid continental climate, with an effective average
20 rainfall of about 350 mm/year varying between 150 mm in dry years and 750 mm in humid
21 years (López-Fuster, 1999). The average net precipitation of the system from 1945-1975 was
22 about 338 Mm³/year (CHJ, 2009). However, in the last decade it decreased to 292 Mm³
23 (86%).

24 With its high agricultural activity and semiarid climate about 90% of water in the Eastern
25 Mancha region is demanded by agriculture (CHJ, 2009), of which the major part is used for
26 irrigation and a small part for livestock breeding. Urban water demand accounts only for
27 about 10% of the total demand and plays a secondary role. The intensive expansion of
28 irrigation since the early 1970's and the cultivation of high water consumption crops led to a
29 significant increase of water demand and groundwater pumping. In the last years the gross
30 extractions have stabilized between 300 and 450 Mm³. However, the renewable resources are

1 assessed between 280 and 330 Mm³/year (CHJ, 2009) and consequently, the aquifer balance
2 is still negative.

3 The intensive groundwater pumping has been led to a significant drop of groundwater level
4 and the piezometric level is locally about 35 m or more under the level of natural regime
5 (Sanz, et al. 2009). As a consequence, the springs, whether they are permanent, temporal or
6 ephemeral, are suffering a notable change in the discharge and a 13% of them have been dried
7 up (López-Sanz, 2010). Also the discharge of Júcar River is heavily influenced, as stream
8 aquifer interaction in the upper reach has changed from a gaining river to a losing or even to a
9 non-connected river.

10 The Júcar River is the main surface water course in the Mancha Region. In the upstream limit
11 of the aquifer region, the Alarcón reservoir (with storage capacity of 1,112 Mm³) serves the
12 main streamflow regulator, providing water supply for urban demand in the region and
13 partially for agriculture. El Molinar reservoir represents the downstream geographic limit of
14 the Júcar river reach within the MO region.

15 The MO aquifer system suffers since the early 1970s a continuous drop of groundwater levels
16 due to intense groundwater pumping. This pumping has been provoked by an important
17 transformation from dry land to irrigated land and consequently an increasing demand of
18 irrigation water for agriculture. Promoted by economic incentives, the development of an
19 intensive agriculture has led to a total irrigated area of about 100,000 ha, whereof the
20 predominant part is supplied by groundwater. Agriculture is the most important economic
21 factor in MO region. However, the overexploitation of the aquifer produces important
22 ecological impacts, ranging from the drying of springs and wetlands to the disappearance and
23 regime alteration of the rivers and the pollution of groundwater from nitrogen leaching due to
24 the intense fertilizer use. Nitrate concentrations in groundwater of up to 125 mg/l have been
25 measured at certain locations (Moratalla et al., 2009), and the aquifer has been designed as nitrate
26 vulnerable zone by the Castilla-La Mancha regional government.

27 Yet, water use from the aquifer is limited by the River Basin Authority (Confederación
28 Hidrográfica del Júcar) and the Jucar River Basin Management Plan. The regulation and
29 control of water abstractions is managed by the Junta de Regantes de la Mancha Oriental,
30 with about 800 members and an irrigated area of 90,000 ha. Groundwater abstractions and
31 water use are controlled by remote sensing and personal inspections (Castaño et al., 2010).
32 The total irrigated area fluctuates between 100,000 ha and 110,000 ha.

1 Various measures have already being proposed to halt overexploitation and achieve the goal
2 of a sustainable aquifer management, by controlling the quantitative overexploitation of the
3 aquifer through collective actions (Lopez-Gunn, 2003) and achieving a good chemical status
4 of the aquifer through the use of fertilizer standards and fertilizer taxes (Peña-Haro et al.,
5 2010 and 2014).

6 This study intends to identify the points of conflict between stakeholders to create a basis for
7 further planning, and also to sensitize the stakeholders to possible impacts caused by different
8 managements. A general approach for finding well-accepted measures should be found by
9 testing the MAVT method, incorporating the stakeholders. And finally, the method's aptitude
10 for conflict resolution in water management has been evaluated. The impact assessment of the
11 measures and the evaluation of alternatives have been realized for the year 2027, which is
12 homogeneous with the provided deadline of fulfilling the goals set in the WFD in second
13 instance. To evaluate the robustness to external changes, two future scenarios were
14 considered: a static and a dynamic one, including possible changes in the available water
15 resources in future. The application of the MAVT in the case study of MO aquifer should give
16 an idea of the possible impacts of various alternatives, its fulfilment of the objectives for the
17 different stakeholders and the conflict potential of the alternatives between the stakeholders.
18 Furthermore, the aptitude of the MAVT as tool for conflict resolution in water management
19 has been tested.

20 **3.2 Application**

21 **3.2.1 Identification of stakeholders**

22 The very first step was the characterization of the problem of sustainable groundwater
23 management in the case study and the identification of all the relevant stakeholders involved.
24 In a large and complex groundwater system as the MO, a broad range of interests are
25 involved. The identification of the main stakeholders involved in this study was based on
26 interviews with experts with regional knowledge and the application of the snowball
27 principle, which involves asking yet identified stakeholders to identify new ones (Hostmann,
28 2005) Eight stakeholder groups were finally selected (Table 1).

29 During the identification process, we checked if a stakeholder's involvement was reasonable
30 for avoiding biased results. Therefore three exclusion criteria were defined:

- 1 ▪ Lack of knowledge about the aquifer system
- 2 ▪ Lack of willingness to cooperate
- 3 ▪ Missing empathy to the aquifer management problem

4 3.2.2 Identification of objectives and attributes

5 A hierarchy tree (see Fig. 4) was defined by experts and afterwards discussed and adapted
6 with experts and stakeholders until the constructed hierarchy of objectives was accepted by all
7 of them. This was absolutely crucial for gaining acceptance of the results. All the objectives
8 contribute to the overall goal of a sustainable management of the groundwater system.

9 Afterwards, attributes with measurable units were assigned to each objective on a quantitative
10 or qualitative scale in order to assess the performance of the different alternatives
11 in relation to that objective. A qualitative scale was assigned when a high uncertainty existed
12 in the impact assessment, and also, if comprehensibility of qualitative data in the valuation
13 process was easier for stakeholders. One of the criterion was that a clear association of the
14 attribute to the objective is given (Keeney and Raiffa, 1976). Another criterion was the
15 validity of the attribute in all the study area, as the MO aquifer management issues are not
16 only local. Especially physical impacts were expressed all over the area representative value
17 (e.g.: groundwater depletion cannot be used as attribute, as the impact varies over the area). A
18 division of the objectives into geographical zones by lower leveled sub-objectives was
19 contemplated, but finally rejected, because of the danger of appearance of local interests in
20 the valuation process and the loss of the objective character. A list of identified objectives
21 with attributes can be found in table 2.

22 3.3 Definition of management alternatives

23 Management alternatives were not defined by experts, but created from already proposed
24 measures. Given that the overexploitation of the Mancha Oriental aquifer has been a problem
25 for many years, different measures already have been developed by different organisations,
26 varying in the basic approach for solving the aquifer management problem (López-Gunn,
27 2003; Santa Olalla, et al. 1999; Peña-Haro et al., 2010 and 2014). After all potential measures
28 were identified they were grouped in fundamental actions (FA):

29

- 30 1. FA1 Control/Restriction of groundwater use

- 1 a. Reduction of irrigated agricultural area, change to dry farming
- 2 b. Reduction of water allotment in drought periods
- 3 c. Change of crops
- 4 d. Improvement of extraction controls
- 5 e. Improvement of irrigation efficiency
- 6 2. FA2 Increased Surface Water Use / Groundwater substitution
 - 7 a. Groundwater substitution by surface water for agricultural and urban water
 - 8 supply
- 9 3. FA3 Water demand reduction by economic instruments
 - 10 a. Implementation of fertilizer standards
 - 11 b. Implementation of water taxes and fertilizer taxes

12

13 To generate alternatives a system generating approach was applied, creating for every
 14 fundamental action different levels of implementation (Bogardi et al., 1982), from slow
 15 (Level1) to high (Level 3).

16

17 FA1: (W1, W2, W3)

18 FA2: (S1, S2, S3)

19 FA3: (E1, E2, E3)

20

21 In a further step the actions were combined with all different implementation stages
 22 considering various restrictions (R). In total 27 discrete alternatives (Ai) were defined by

23

$$24 \quad A_i = A_i (W_j, S_k, E_l; R) \quad (2)$$

25 The alternative with the lowest implementation stage for all fundamental actions represents
 26 the status-quo alternative. The compatibility of the different measures was checked. Two
 27 exclusion criteria to check for compatibility have been set: (1) two measures are partially

1 dependent and have contradictory outcomes and (2) a measure makes another measure
2 redundant. Restrictions (R) referred to external or immutable variables that influence on the
3 available water resources. In the MO aquifer these restrictions come from river discharge
4 constraints for downstream adjacent regions, but also reliability of urban water supply.
5 Putting restrictions helped to focus on an inner solution of the aquifer management.

6 Although the directly ranking of a high number of complex alternatives with different
7 outcomes would be very difficult, unrealistic and probably meaningless task for the
8 stakeholders, the MAVT will allow to indirectly rank the alternatives for each stakeholder by
9 obtaining their values and preferences in terms of objectives, and assessing the performance
10 of the alternatives on those objectives.

11

12 **3.4 Impact assessment.**

13 For assessing how well each alternative meets the objectives, the impact assessment was
14 forecasted for a certain time horizon. As the principal objectives are based on the guidelines
15 of the WFD, the deadlines for its implementation were chosen. The date of evaluation was set
16 in 2027, since this is one of the stages to be addressed. The base situation described the
17 average situation between 2000 and 2008 to avoid possible anomalies of one specific year.
18 Some of the proposed measures have already been implemented or are in progress of
19 realization.

20 To analyse the influence of climate change and other changes in the hydrological system two
21 scenarios with different assessment (consequences) matrices were introduced:

- 22 • Scenario 1: Static scenario
- 23 • Scenario 2: Dynamic scenario

24 Scenario 1 assumes no changes in external influences or restrictions to the base situation
25 between 2000 and 2008. This scenario focuses on changes and influence of agricultural
26 management measures.

27 Scenario 2 incorporated changes of urban water demand and climate. The possible impact of
28 climate change on water resources (Chirivella, 2010) and the possible increase of urban water
29 supply, handled as a restriction, were introduced in the calculation of the impact assessment.
30 By comparing the results the vulnerability of the acceptance of different measures and

1 changes in preferences could be evaluated as well as the importance on the outcome by
2 considering such variables.

3 For the assessment of the physical impact assessment we used an existing groundwater model
4 (Sahuquillo et al., 2008), groundwater budgets from the annual statistics of the Jucar River
5 Basin Agency (CHJ) and piezometric level observations (Lopez Sanz, 2010). Fitted
6 regression equations (Sanz et al., 2009) and nitrate measurements (Moratalla et al., 2009)
7 were used to estimate the influence of management options to nitrate concentration. For the
8 assessment of possible climate change impacts on the water resources we used estimations
9 from Chirivella-Osma (2010) and Chirivella-Osma et al. (2014).

10 For the assessment of economic and socioeconomic impact we used input-output matrices
11 from the statistic institute of Castilla La Mancha, fitted regressions from previous
12 hydroeconomic models of the groundwater system (Peña-Haro et al., 2010 and 2014) to
13 estimate the effects of different policies such as fertilizer taxation and water prices, and other
14 relevant data from previous studies in the same case study (López Fuster, 1999; Santa Olalla,
15 1999)

16

17 **3.5 Quantification of the stakeholders' preferences**

18 The value function and the weights of the alternatives were elicited through interviews with
19 representatives of every stakeholder group. The valuation of the attributes was realized for the
20 range between the lowest and the highest level of the attribute for all alternatives in both
21 scenarios. This range differed slightly between the two scenarios and it would have required
22 separate valuation interviews for every scenario. However, it was confusing for stakeholders
23 to realize twice an interview with almost identical ranges of attributes. Because of this, a little
24 error has been accepted, as it was not of high significance to the outcomes. Also, the use of
25 qualitative scales of attribute levels reduced this problem.

26 Personal computer-aided interviews (section 2.2) were conducted with representatives of the
27 stakeholder groups. Either the interview was with the official spokesman of a stakeholder
28 group or with at least two stakeholders per stakeholder group to maintain the objective
29 character. Within one stakeholder group, the values of the representatives were averaged. If a
30 stakeholder was not available for a personal interview, it was conducted by electronic
31 interviews, although personal interviews were preferred to avoid misinterpretations or errors

1 in the valuation. In addition to the elicitation of the value function and weights, some general
2 questions about the aquifer management were asked. Through the interview questions the
3 interviewee`s holistic preferences were evaluated. Afterwards, a workshop in Albacete
4 (Spain) was held where the results were presented and stakeholders were asked to evaluate
5 their prior valuation of preferences.

6 **4 Results**

7 **4.1 Ranking of alternatives and interpretation of results**

8

9 Based on the total value of the alternatives for the stakeholder groups we proceeded to rank
10 the different alternatives between 1 (best) and 36 (worst). The different alternatives were
11 ranked quite similarly (see Figure 5) across the stakeholder groups, although some clear
12 discrepancies can be also detected. According to the results of the MAVT method,
13 stakeholders tend to prefer a mix of different fundamental actions for problem resolution. The
14 potential of combined measures might be preferred due to economic reasons, and might cause
15 smaller impacts, distributed on various sectors instead of one rigorous measure. For an
16 alternative to be considered good, ecological, economic and also social interests have to be
17 considered. Of course among the different stakeholder groups the focus differs, but the best-
18 ranked alternatives are still the same. Figure 6 shows the evaluation of the alternatives and the
19 range of variation between stakeholders. The best overall alternatives were chosen in two
20 steps. First, all dominated alternatives were eliminated. Secondly, the sum of the stakeholder
21 rankings was calculated for every alternative. The lower the sum of ranking, the better the
22 overall evaluation.

23 Through analysing the composition of the total preference value it could be seen that
24 economic, ecological and social interests had to be fulfilled by an alternative to achieve a
25 good ranking.

26 The representatives of all stakeholder groups agreed that a restriction of water access is
27 necessary to obtain a sustainable aquifer management. The impediment of a further increase
28 of the water demand by legal instruments was fundamental for a high ranking of an
29 alternative. It can be reasoned, that any other measures will be declined, if there will be no
30 restriction to water access. Such measures are the reduction of the irrigated agricultural area,

1 the reduction of the water allotment in drought periods, restrictions for high water needing
2 crops, the improvement of extraction controls and the improvement of the irrigation
3 efficiency. Alternatives 1-9 do not include such measures to limit the water access and
4 received consequently a low ranking. Also, the compliance of the ecological objectives relied
5 mainly on this restriction.

6 Groundwater substitution by surface water was evaluated as an appropriate measure, although
7 the preferred implementation factor differed among the stakeholders. Especially the high costs
8 and the long realization time of a full implementation affected negatively on the results of
9 environmental organizations and rural development. To agricultural representatives it
10 represented a necessary measure to guarantee water supply and economic activity.

11 Regarding the economic instruments, fertilizer taxation is considered as the most adequate
12 solution, especially because of its positive influence on the nitrate concentration in
13 groundwater. However, the influence of water taxation on the results is quite low and is
14 considered less adequate than other measures.

15 In general, it can be observed that the main question being considered a good alternative is
16 not only “what is the best measure”, but “what is the implementation stage of a measure” and
17 “how is the combination with other measures”.

18 **4.2 Conflict potential**

19 A conflict potential might arise when there is a large variation in the ranking of certain
20 alternative across all the stakeholder groups. Analyzing the conflict potential helps to focus on
21 the proper measures for reaching a consensus solution. The conflict potential expressed by the
22 mean deviation of the evaluation of alternatives by the stakeholders is shown in figure 7.

23 Analyzing the fundamental actions, conflict potential could be found in the full
24 implementation of the planned measures. High implementation stages of measures are just
25 accepted by all stakeholder groups, if they are combined with other fundamental measures.
26 For instance, the option of full implementation of groundwater substitution is assigned a low
27 ranking by ecologists unless the option also includes water use restrictions to prevent an
28 overall increase in irrigation. On the other hand, agricultural representatives would accept a
29 water use restriction provided that groundwater substitution measures are also implemented to
30 avoid economic drawbacks.

1 The acceptance of economic measures varies significantly and has high conflict potential.
2 According to the results, water taxes are not decisive for a good alternative ranking. Fertilizer
3 taxes contain high conflict potential, but they are provided in the best-ranked alternatives, due
4 to their regulation capacity of nitrate concentration in groundwater. However, in reality, the
5 acceptance of taxes, especially by agricultural representatives (obtained by direct interviews)
6 might be low, because of the economic losses as a consequence.

7 Uncertainties in the future by dynamic variables (Scenario 2) are another source of conflicts.
8 This might be because of the uncertainty of future development. Static external variables
9 (Scenario 1) provide more precise results, but they might not represent the future reality.

10

11 **4.3 Scenario comparison**

12 The scenario comparison served to show up conflict points coming up because of
13 uncertainties on variables in future. In Scenario 2, due to the inclusion of a greater urban
14 water demand and climate change, the achievement of the objective of a sustainable aquifer
15 management was more difficult than in Scenario 1. Figure 8 shows changes in ranking for the
16 three best evaluated scenarios 1. Main questions were if there were significant changes in the
17 preferences of outcomes and how the alternatives change between the two scenarios.

18 Looking at the best evaluated alternatives in both scenarios it could be seen that the most
19 preferred alternatives are not entirely the same, but the focal point of good evaluated
20 alternatives were rather similar. The following similarities could be detected:

- 21 • Alternatives with focus on just one measure are evaluated worse than those with
22 several different measures of the fundamental actions
- 23 • To obtain a good ranking, a restriction to the water access has to be implemented
- 24 • A substitution of groundwater by surface water on a medium level of implementation
25 is recommended
- 26 • A control of nitrate contamination by fertilizer taxation is desired to fulfil ecological
27 objectives of WFD

28 Focusing on the differences between the scenarios the following shifts could be observed:

- 1 • Alternatives with combination of various fundamental actions on a medium level of
2 implementation show up with the highest losses (although their absolute rank still can
3 be high)
- 4 • The highest improvement of alternative ranks can be observed in the alternatives with
5 more rigorous measures as a consequence to the higher necessity of actuating, due to
6 the bigger water resource availability problem
- 7 • Also the status-quo alternative improves the acceptance. This might be because of the
8 decreased cost efficiency of the measures. In other words, the cost of measures is not
9 justified by the result.

10 Differences between the scenarios could be detected, but nevertheless the tendency of the best
11 ranked alternatives were the same and alternatives with a mix of different measures were the
12 best ranked in both scenarios. Nevertheless, in MO case the difference in the preferences of
13 outcome between the scenarios was of moderate importance.

14 **4.4 Sensitivity Analysis**

15 The sensitivity analysis was applied to test the robustness of the results towards uncertainties
16 in the inputs. The results of the MAVT depend on two principle input factors:

- 17 • The attribute levels of the alternatives (consequence matrix)
- 18 • The valuation of the attributes by the stakeholders

19 The uncertainty of attribute levels in the consequence is expressed by a possible variation of
20 the forecasted attribute levels for every alternative. The sensitivity analysis was realized by
21 varying the attributes separately in the possible fluctuation range and analysing the influence
22 on the outcome of the alternatives. The fluctuations arise from uncertainties of the underlying
23 models of attributes (e.g.: groundwater model Júcar Basin, Sahuquillo, et al., 2008), but also
24 from expert rated uncertainty ranges in qualitatively evaluated attributes. The attributes high
25 irrigated area, low implementation costs and maintenance and management costs, are
26 excluded from the sensitivity analysis. They include fundamental characters, defining the
27 consequences of the alternatives and are consequently treated as fixed values. The robustness
28 of the preferences of the alternative varies within the stakeholder groups. Despite the
29 influence on the preference value, the effect on the total ranking is quite low, and the basic
30 structure of the ranking dos not change (table 3). The most preferred alternatives also obtain a

1 high ranking considering the uncertainty range. Referring to the uncertainties in the attribute
2 levels, there are no significant negative variation observable.

3 The uncertainty in the valuation of the attributes is given by different value functions and
4 weights assigned by different representatives within one stakeholder group, which results in
5 different preference values and alternative ranking. A similar valuation within a stakeholder
6 group creates homogeneity and uncertainty becomes low. The sensitivity analysis for
7 uncertainty in valuations was done for stakeholder groups with more than one representative.

8 Uncertainties within the stakeholder groups were significant for ‘environmental
9 organizations’, (Fig. 9) meanwhile municipalities (Fig.10) and rural development had a quite
10 similar evaluations. This was interesting, as one would expect lower discrepancy between
11 ecologists than between other stakeholder groups.

12

13 **4.5 Discussion and Conclusions**

14 Finding the optimal solution with the Multi Attribute Value Theory in water resource
15 problems seems to be difficult, requiring that complex structures are reduced to one value,
16 expressing the acceptance or negotiation of an alternative. More complex methods like for
17 instance ELECTRE (San Cristóbal Mateo, 2012) might be more appropriate to find the best
18 solution, but in the application of these there is a risk of non-transparency and lack of
19 understanding about the method among the participating stakeholders.

20 Nevertheless, the applied decision analysis framework based on the MAVT is a useful method
21 to find possible conflict points between the multiple stakeholders, helping also to identify
22 possible consensus solutions. Furthermore, it is possible to define basic criteria for alternative
23 planning, to guarantee a high acceptance of measures and avoid future conflicts.

24 The acceptance of the method is quite high (see Fig.11), because of its simplicity (see also
25 Marttunen et al., 2013). The involvement of stakeholders at the beginning of the planning
26 process, especially in the setting of the objectives, is considered important to obtain high
27 acceptance. It is important to weigh the necessary complexity of the model with the
28 comprehensibility. Stakeholders have to be chosen carefully in terms of their knowledge of
29 the issue, and they should have a good overview of the problem. If not, the valuations will be
30 made without basing them on facts. If stakeholders understand the method, the acceptance of

1 the results will be higher, and their contribution to conflict resolution too, since the results
2 become more acceptable.

3 The holistically assessed preferences given by some stakeholders are mostly coherent with the
4 results of the MAVT method. In comparison to a holistic ranking of options, the MAVT
5 method has the advantage of creating a more detailed evaluation framework, which enables
6 more informative analysis to be undertaken. This includes a more detailed analysis of conflict
7 potential and the ability to undertake uncertainty and sensitivity analysis.

8 The approach has been applied to the analysis of sustainable management of the Mancha
9 Oriental aquifer, allowing to elicit stakeholder groups' values and evaluating groundwater
10 management options. Stakeholders clearly preferred combined measures due to economic
11 reasons, and cause smaller impacts, because impacts are distributed on various sectors. For an
12 alternative to be considered good in MO, ecologic, economic and also social interests have to
13 be considered. Logically among the different stakeholder groups the focus differs, but the
14 best-ranked alternatives are still the same due to their basic structure. The representatives of
15 all stakeholder groups agree that a restriction of water access is necessary to obtain a
16 sustainable aquifer management. Other measures just have efficiency if there is no
17 additionally increase of the water use. Also, the compliance of the ecological objectives relies
18 mainly on this restriction. Groundwater substitution by surface water is evaluated as an
19 appropriate measure, although the preferred implementation factor can differ among the
20 stakeholders. Especially the high costs and the long realization time of a full implementation
21 affect negatively on the results of "environmental organizations" and "rural development". To
22 "agricultural representatives" groundwater substitution represents a necessary measure to
23 guarantee water supply and economic activity. According to economic instruments, fertilizer
24 taxation is considered as the most adequate solution to achieve a good ecological status,
25 especially because of its ability to limit the concentration of nitrate in the groundwater. The
26 influence of water taxation on the results however is quite low and is considered less adequate
27 than other measures.

28 Although the amount of alternatives seemed to be unmanageable for the stakeholders at the
29 beginning, the MAVT approach helped them to get more sensitized to the complexity of
30 groundwater management in this aquifer and to elicit stakeholders' preferences and potential
31 conflict points. In summary, the applied MAVT method is a useful support tool for planning
32 processes, not to find the best solution, but to avoid future conflicts and find potential

1 consensus solutions by a detailed analysis of the measures rank based on the stakeholder
2 preferences and values. Also it serves to sensitize stakeholders to competing interests in
3 environmental problems.

4

5 **Acknowledgements**

6 The authors would like to thank all stakeholders for the cooperation and participation in this
7 study. We extend special thanks to Dr. Alfonso Calera, of IDR (Instituto de Desarrollo
8 Regional) of University of Castilla-La Mancha for the organization of the workshop with the
9 stakeholders to introduce the study, as well as to the Jucar River Basin Agency
10 (Confederacion Hidrografica del Júcar), represented in the meeting by Luis Garijo, and the
11 Junta Central de Regantes de la Mancha Oriental (represented by its president, Francisco
12 Belmonte) for all the data and information provided. This work has been partially funded by
13 the European Community 7th Framework Project GENESIS (n. 226536) on groundwater
14 systems and the Plan Nacional de I+D+I 2008-2011 of the Spanish Ministry of Science and
15 Innovation (projects CGL2009-13238-C02-01 and CGL2009-13238-C02-02 on climate
16 change impacts and adaptation).

17

1 **References**

- 2 Belton, V., Stewart, T.J.: Multiple Criteria Decision Analysis. An Integrated Approach.
3 Springer US, Boston, MA, 2002.
- 4 Bogardi, J.J., Nachtnebel, H.P.: Multicriteria decision analysis in water resources
5 management. UNESCO, Paris, 1994.
- 6 Bogardi, J.J., Nachtnebel, H.P., Duckstein, L. Multicriterion Analysis for Regional Water
7 Resource Development, Part I: Cost Effectiveness Approach, Environmental Systems
8 Analysis and Management, pp. 74–81.
- 9 Bromley, J., Cruces, J., Acreman, M., Martinez, L., & Llamas, M. R.: Problems of sustainable
10 groundwater management in an area of over-exploitation: the Upper Guadiana catchment,
11 central Spain. International Journal of Water Resources Development, 17(3), 379-396, 2001.
- 12 Castaño, S., Sanz, D. and Gómez-Alday, J.J.: Methodology for quantifying groundwater
13 abstractions for agricultura via remote sensing and GIS. Water Resource Management 24(4),
14 795-814, 2010.
- 15 Chirivella-Osma, V.: Tesis doctoral: Caracterización de los futuros escenarios climáticos en la
16 Comunidad Valenciana: propuestas de mejora para la evolución de la oferta y demanda de
17 recursos hídricos, Universidad Politécnica de Valencia, Valencia, Spain (in Spanish), 2010.
- 18 Chirivella-Osma , V., Capilla Romá, J. E., Pérez Martín M. A.: Modelling regional impacts of
19 climate change on water resources: the Júcar basin, Spain. Hydrological Sciences
20 Journal 60(1), 30-49, 2015.
- 21 European Commission: Directive 2000/60/EC establishing a framework for Community
22 action in the field of water policy. EU-WFD, 2000.
- 23 Foster, S., Garduño Héctor, Kemper, K., Tuinhof, A., Nanni Marcela, Dumars, C. Sustainable
24 Groundwater Management, Conept & Tools. Groundwater Quality Protection. Defining
25 strategy and setting priorities. GW Mate Briefing Notes Series 2002 (Note 8).
26 www.worldbank.org/gwmate, 2002.
- 27 Hostmann, M.: Decision support for river rehabilitation. ETH, Zürich, 2005.

- 1 Hostmann, M., Bernauer, T., Mosler, H.-J., Reichert, P., Truffer, B.: Multi-attribute value
2 theory as a framework for conflict resolution in river rehabilitation. *J. Multi-Crit. Decis. Anal.*
3 13 (2-3), 91–102, 2005.
- 4 Kangas, A., Kangas, J., Kurttila, M.: Decision support for forest management. Springer,
5 Berlin, 2008.
- 6 Karjalainen, T. P.; Rossi, P. M.; Ala-aho, P.; Eskelinen, R.; Reinikainen, K.; Kløve, B. et al.:
7 A decision analysis framework for stakeholder involvement and learning in groundwater
8 management. In: *Hydrol. Earth Syst. Sci.* 17 (12), S. 5141–5153. DOI: 10.5194/hess-17-5141-
9 2013, 2013.
- 10 Keeney, R.L., Raiffa, H.: Decisions with multiple objectives. Preferences and value tradeoffs.
11 Wiley, New York NY u.a., 1976.
- 12 Llamas, M. R., & Martínez-Santos, P.: Intensive groundwater use: silent revolution and
13 potential source of social conflicts. *Journal of Water Resources Planning and*
14 *Management*, 131(5), 337-341, 2005.
- 15 López Fuster, P.: Los regadíos en la Mancha Oriental, Albacete, 1999.
- 16 López Sanz, G.: Estudios de fuentes, manantiales y pequeños espacios del agua n la cuenca
17 media de los ríos Júcar y Cabriel, 2010.
- 18 Lopez-Gunn, E.: The Role of Collective Action in Water Governance: A Comparative Study
19 of Groundwater User Associations in La Mancha Aquifers in Spain. *Water International* 28
20 (3), 367–378, 2003.
- 21 Lopez-Gunn, E., Cortina, L.M.: Is self-regulation a myth? Case study on Spanish groundwater
22 user associations and the role of higher-level authorities. *Hydrogeol J* 14 (3), 361–379, 2006.
- 23 Marttunen, M., Hämäläinen, R.P.: Decision analysis interviews in environmental impact
24 assessment. *Eur J Oper Res* 87(3):551–563, 1995.
- 25 Marttunen, M., Hämäläinen, R.P.: Decision analysis interviews in supporting collaborative
26 management of a large regulated water course. *Environ Management* 42(6):1026–1042, 2008.
- 27 Marttunen, M., Mustajoki, J., Dufva, M., Karjalainen, TP.: How to design and realize
28 participation of stakeholders in MCDA processes? A framework for selecting an appropriate
29 approach. *EURO J Decis Process.* 1–28, 2013. DOI 10.1007/s40070-013-0016-3, 2013.

- 1 Moratalla, A., Gomez Alday, J., De las Heras, J., Sanz, D., & Castaño, S.: Nitrate in the
2 water-supply wells in the Mancha Oriental Hydrogeological System. *Water Resource*
3 *Management* 23,1621-1640, 2009.
- 4 Mustajoki, J. and Hämäläinen, R.P.: Web-HIPRE: Global decision support by value tree and
5 AHP analysis, *INFOR*, 38, 208–220, 2000.
- 6 Mustajoki, J., Saarikoski, H., Marttunen, M., Ahtikoski, A., Hallikainen, V., Helle, T. et al.:
7 Use of decision analysis interviews to support the sustainable use of the forests in Finnish
8 Upper Lapland. In: *Journal of Environmental Management* 92 (6), S. 1550–1563. DOI:
9 10.1016/j.jenvman.2011.01.007, 2011.
- 10 Peña-Haro, S., Llopis-Albert, C., Pulido-Velazquez, M., Pulido-Velazquez, D.,: Fertilizer
11 standards for controlling groundwater nitrate pollution from agriculture: El Salobral-Los
12 Llanos case study, Spain. *Journal of Hydrology* 392 (3-4), 174–187, 2010.
- 13 Peña-Haro, S., García-Prats, A., Pulido-Velazquez, M.,: Influence of soil and climate
14 heterogeneity on the performance of economic instruments for reducing nitrate leaching from
15 agriculture. *Science of the Total Environment* 499, 510-519, 2014.
- 16 Pöyhönen, M., Hämäläinen, R. On the convergence of multiattribute weighting methods,
17 *European Journal of Operational Research* 129, pp. 569±585. DOI: 10.1016/S0377-
18 2217(99)00467-1, 2001.
- 19 Reichert, P., Borsuk, M., Hostmann, M., Schweizer, S., Spörri, C., Tockner, K., Truffer, B.:
20 Concepts of decision support for river rehabilitation. In: *Environmental Modelling &*
21 *Software* 22 (2), S. 188–201. DOI: 10.1016/j.envsoft.2005.07.017, 2007.
- 22 Roberts, R., Goodwin, P.: Weight approximations in multi-attribute decision models. *J. Multi-*
23 *Crit. Decis. Anal.* 11 (6), 291–303, DOI: 10.1002/mcda.320, 2002.
- 24 Sahuquillo, A., Castaño, S., Cassiraga, E., Calera, A., Gómez-Alday, J. J., Peña, S. et al.: .
25 *Modelo de flujo subterráneo de los acuíferos de la Mancha Oriental y sus relaciones con los*
26 *ríos Júcar y Cabriel. Albacete, Valencia: UPVLC, UCLM, 2008.*
- 27 San Cristóbal, M., Ramón, J.: *Multi criteria analysis in the renewable energy industry.*
28 *London, New York: Springer (Green energy and technology), ISBN: 9781447123460, 2012.*
- 29 Santa Olalla Mañas, F. M., Brasa Ramos, A., Fabeiro Cortes, C., Fernando González, D., &
30 López Córcoles, H.: *Improvement of irrigation management towards the sustainable use of*

1 groundwater in Castilla-La Mancha, Spain. *Agricultural Water Management* , pages 195-205,
2 1999.

3 Sanz, D., Gómez-Alday, J.J., Castaño, S., Moratalla, A., las Heras, J. de, Martínez-Alfaro,
4 P.E.: Hydrostratigraphic framework and hydrogeological behaviour of the Mancha Oriental
5 System (SE Spain). *Hydrogeol J* 17 (6), 1375–1391, 2009.

6 Soncini-Sessa, R.: *Integrated and participatory water resources management-practice*, vol. 1b.
7 Elsevier, Amsterdam, 2007.

8 Stefanopoulos, K., Yang, H., Gemitzi, A., Tsagarakis, K. P.: Application of the Multi-
9 Attribute Value Theory for engaging stakeholders in groundwater protection in the Vosvozis
10 catchment in Greece. In: *Science of The Total Environment* 470-471, S. 26–33. DOI:
11 10.1016/j.scitotenv.2013.09.008, 2014.

12 Winterfeldt, D.v., Edwards, W.: *Decision analysis and behavioural research*. Cambridge Univ.
13 Press, Cambridge u.a., 1986.

14

1 Table 1: Stakeholder groups and representatives in La Mancha Oriental

Stakeholder	Representation	Interviewee
National Administration	Jucar River Basin Authority	1
Regional Administration	Regional Government	1
Municipalities	Local Government, municipal Association	2
Agricultural Representatives	Junta Central de Regantes de la Mancha Oriental	1
Environmental Organisation	Local environmental organizations	2
Industry	Power generation, water using Industry, fertilizer production	0
Recreational Organizations	Recreative and tourism organizations	0
Regional development organizations	ADIMAN, Institute of Regional Development	2

2

3

1 Table 2: Attributes and measurement units for the lowest level objectives

Objectives	Attribute	Unit
Good chemical groundwater status	Water quality	Groundwater nitrate concentration mg/l
Good quantitative status of aquifer	Budget	Mm ³ / year
Recuperation of springs and wetlands	Recuperation potential	qualitative scale
Good ecological status Júcar	Stream – aquifer interaction	qualitative scale
Max. utilisation of irrigable area	Irrigation area	ha
High crop profitability	Net Benefit per ha.	€
Industrial Productivity/ Energy potential	Influence on Energy production in Júcar river	qualitative scale
Short realization time	Time between planning and realization times	Years
Low implementation costs	Cost of measures, etc.	€/qualitative
Low maintenance and management cost	Cost of administration and control	€
Create Employment	Number of jobs	Number of jobs
Improve recreational opportunities	Recreational space	qualitative scale
Increase of regional productivity	Influence on per capita income of region	+/-
High cost-benefit ratio		%

2

3

4

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39

1 Table 3: Results of sensitivity analysis: Influence of changes of attribute levels on the
 2 preference value for all stakeholder groups in % respectively to the total value

POSITIVE VARIATION OF ATTRIBUTE LEVELS					
[%]	15a	15b	18a	18b	24b
MAX VARIATION	13.9	12.4	18.2	17.6	13
MIN VARIATION	-7.5	-7.7	-9.9	-9.6	-10.5
AVERAGE VARIATION	1.7	1.4	1.2	1.3	1.3
Negative variation of attribute levels					
[%]	15a	15b	18a	18b	24b
MAX VARIATION	3.2	1.2	11	10.7	8.4
MIN VARIATION	-17.2	-16.2	-22.9	-20	-19
AVERAGE VARIATION	-5	-5.3	-3.6	-3.3	-3.8

3

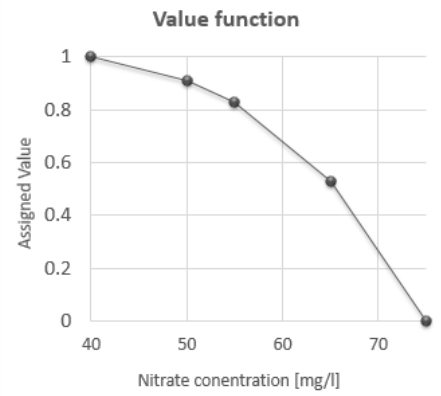
4

5

Objetivo: Good chemical groundwater status

Atributo: Water quality

ID	Nitrate concentration [mg/l]	Assigned Value
1	40	
2	50	
3	55	
4	65	
5	75	



1

2 Figure 2: : Interactive tool for stakeholders to define valuefunctions (developed in Microsoft

3 Excel)

4

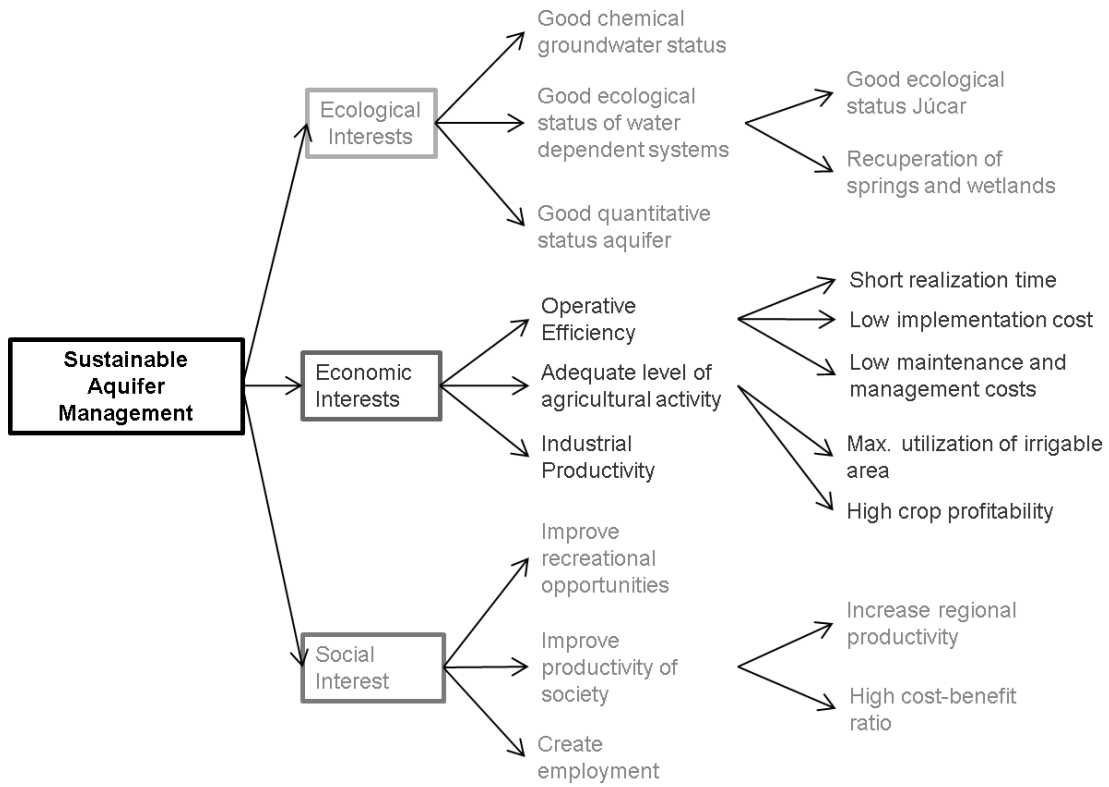


1

2 Figure 3: Case study area: La Mancha Oriental Aquifer in South Eastern Spain

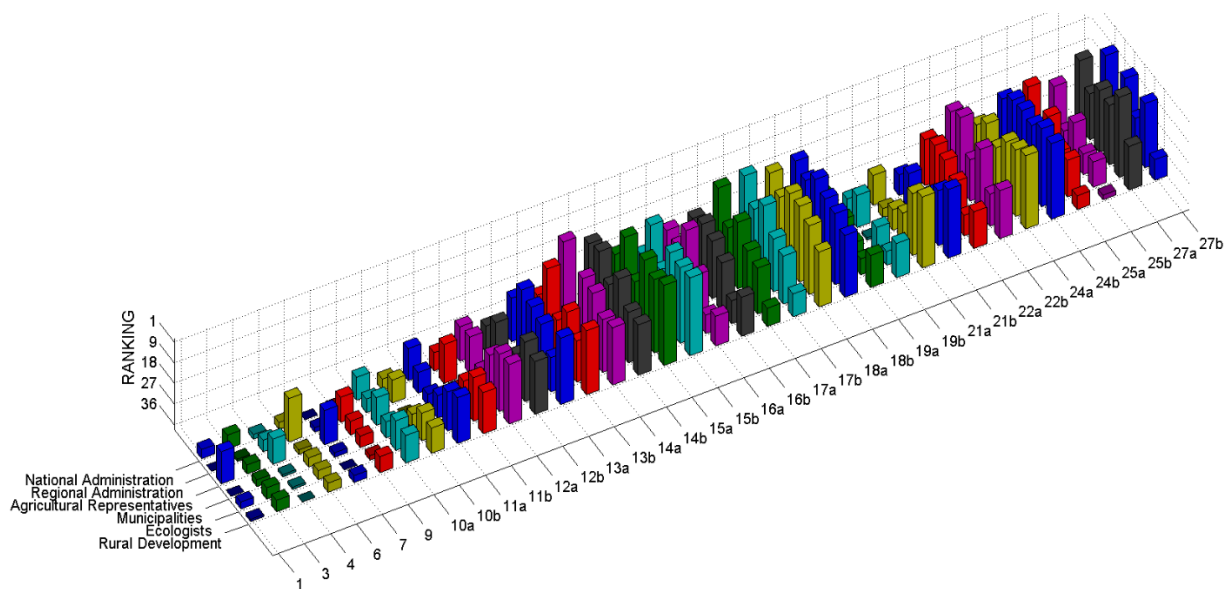
3

4



1
2
3
4

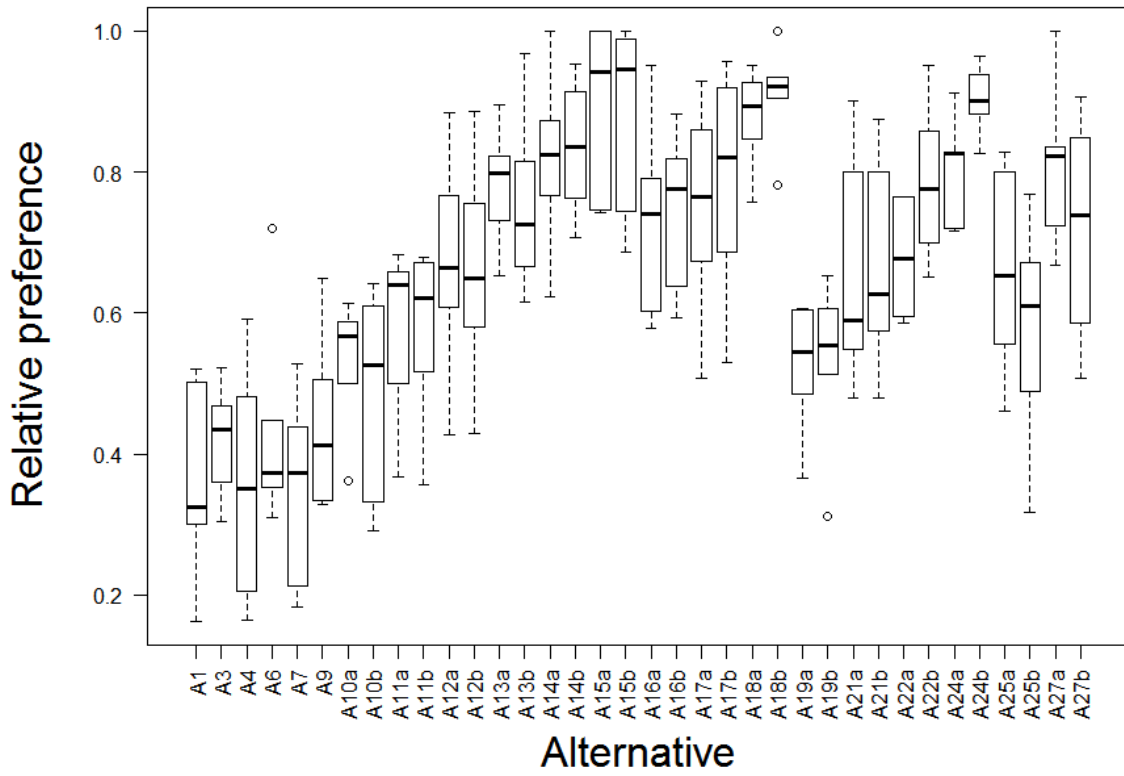
Figure 4: Hierarchy value-tree of objectives from interviews to experts and stakeholder



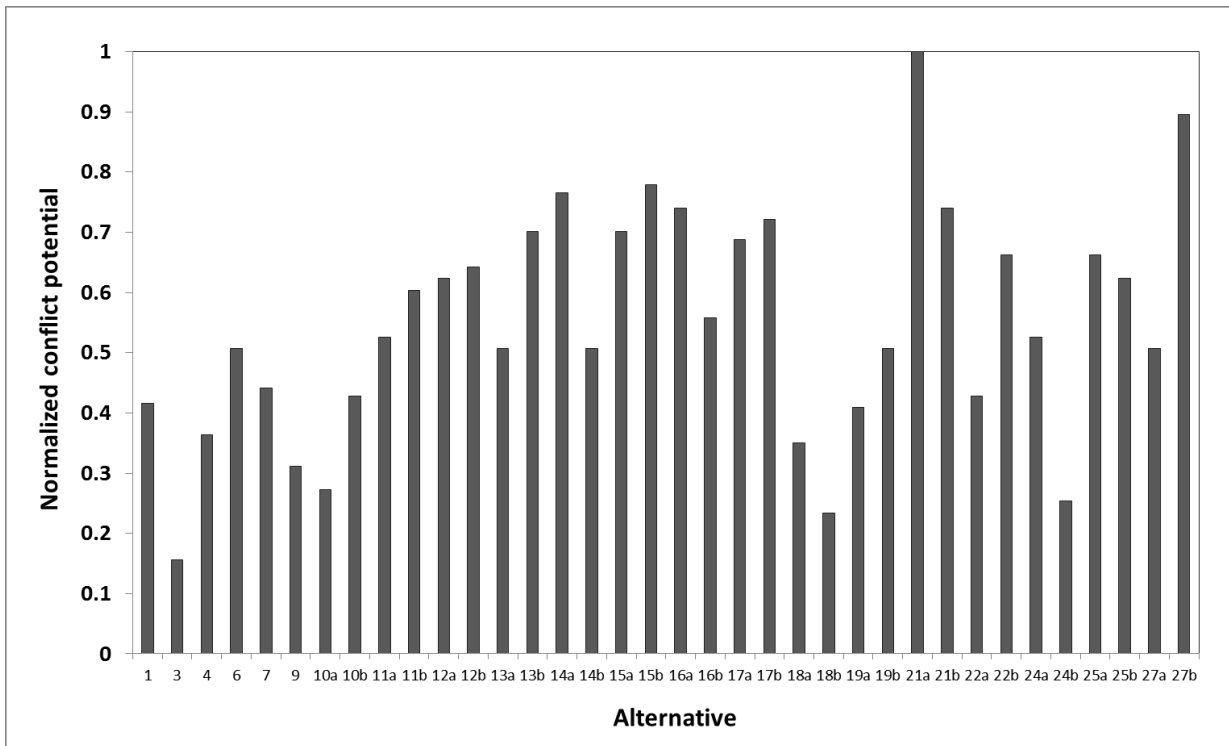
1

2 Figure 5: Ranking of alternatives by the stakeholders in scenario 1

3



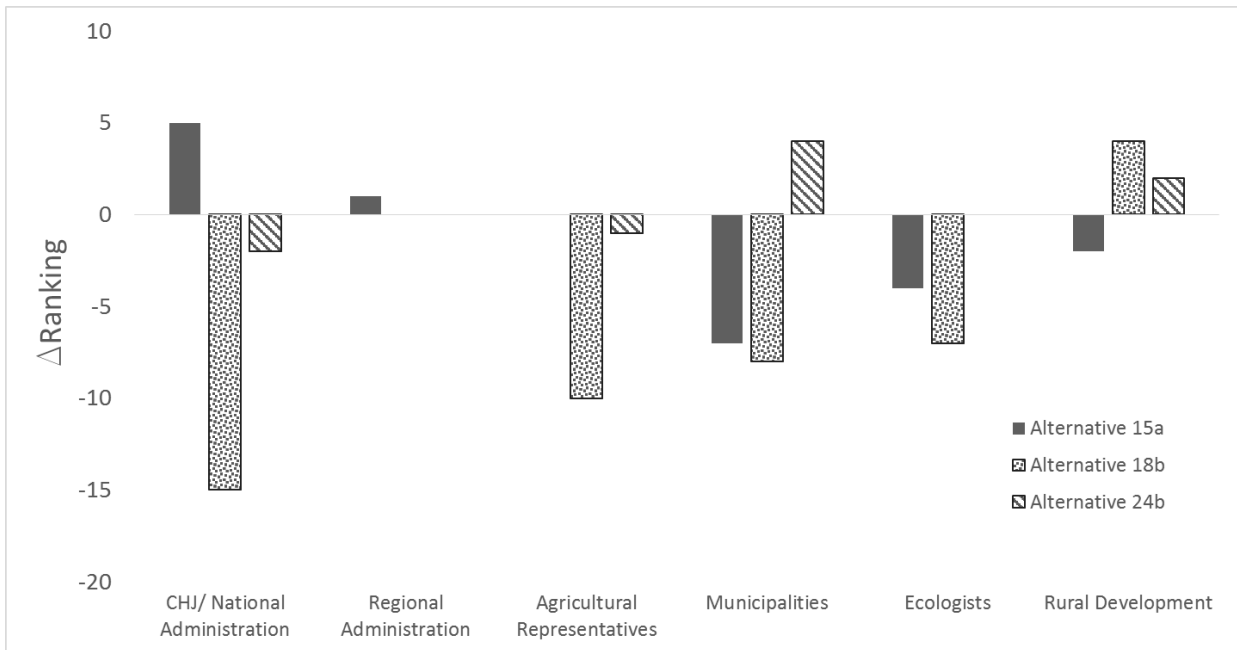
1
 2 Figure 6: Standardized values of preferences of the alternatives by the stakeholders (min,
 3 median, 75th percentile, max)
 4



1

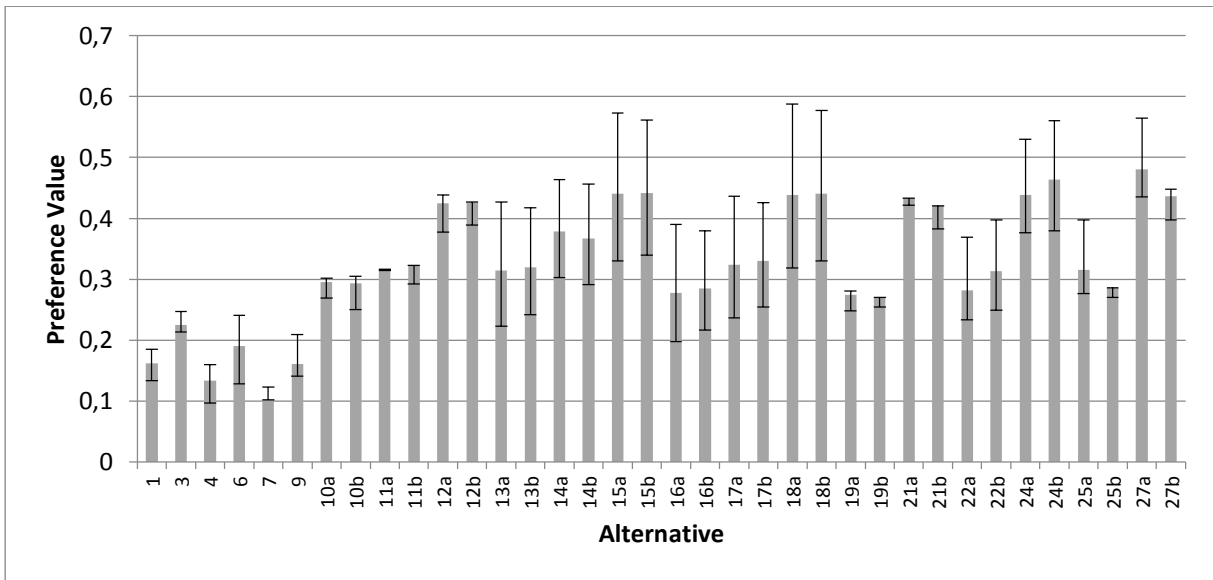
2 Figure 7: Normalized values ranking discrepancy expressed through mean deviation of
 3 preference relative values across stakeholders

4



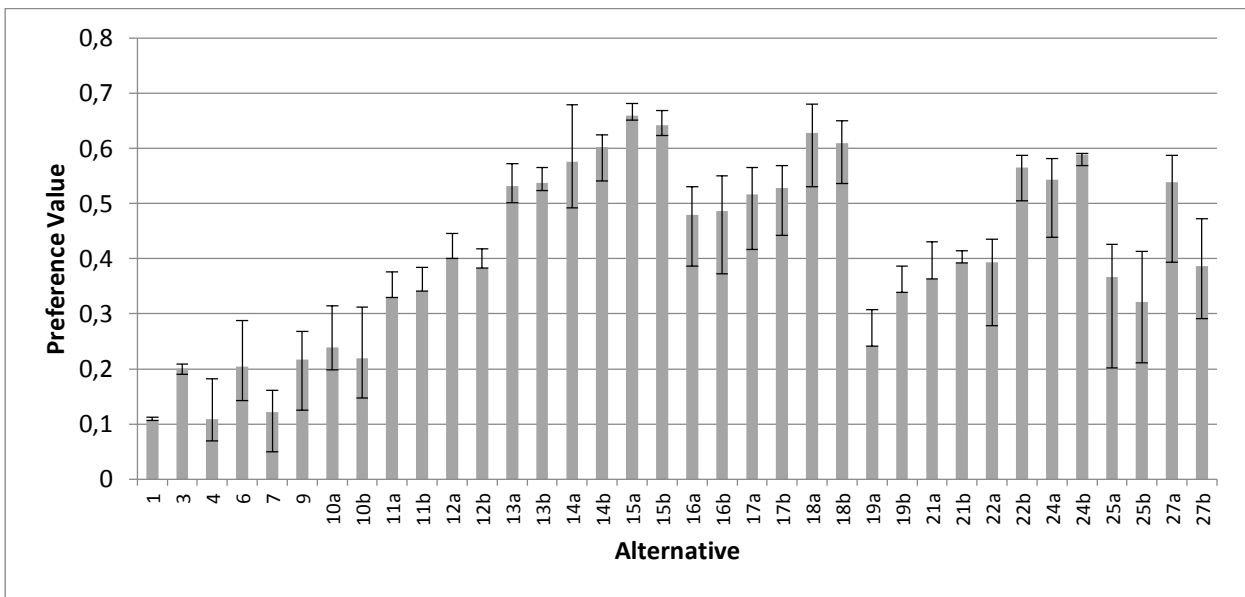
1
 2 Figure 8: Change of ranking between scenario 1 and scenario 2 of the three best evaluated
 3 alternatives in scenario 1

4



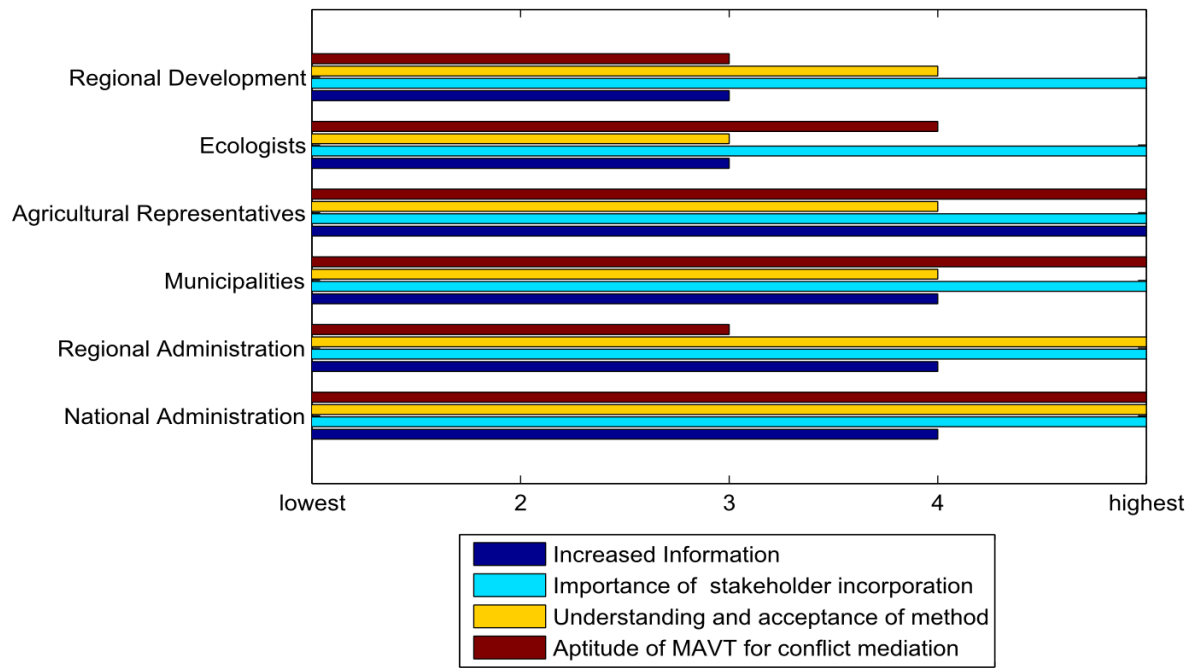
1
 2 Figure 9: Preference values of stakeholder group “Ecologists” (bars) and fluctuation range
 3 due to different valuations of attributes within the the stakeholder group

4



5
 6 Figure 10: Preference values of stakeholder group “Municipalities” (bars) and fluctuation
 7 range due to different valuations of attributes within the the stakeholder group

8



1

2 Figure 11: Average evaluation of the method based on stakeholders' feedback

3