



**SWMM predictions
of urban headwater
stream hydrological
response**

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This discussion paper is/has been under review for the journal Hydrology and Earth System Sciences (HESS). Please refer to the corresponding final paper in HESS if available.

Using the Storm Water Management Model to predict urban headwater stream hydrological response to climate and land cover change

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Received: 15 May 2013 – Accepted: 17 May 2013 – Published: 4 June 2013

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Published by Copernicus Publications on behalf of the European Geosciences Union.

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Abstract

Streams are natural features in urban landscapes that can provide ecosystem services for urban residents. However, urban streams are under increasing pressure caused by multiple anthropogenic impacts, including increases in human population and associated impervious surface area, and accelerated climate change. The ability to anticipate these changes and better understand their effects on streams is important for developing and implementing strategies to mitigate potentially negative effects. In this study, stream flow was monitored during April–November (2011 and 2012), and the data were used to apply the Storm Water Management Model (SWMM) for five urban watersheds in central Iowa, USA representing a gradient of percent impervious surface (IS, ranging from 5.3 to 37.1 %). A set of three scenarios was designed to quantify hydrological responses to independent and combined effects of climate change (18 % increase in precipitation), and land cover change (absolute increases between 5.2 and 17.1 %, based on separate projections of impervious surfaces for the five watersheds) for the year 2040 compared to a current condition simulation. An additional set of three scenarios examined stream response to different distributions of land cover change within a single watershed. Hydrological responses were quantified using three indices: unit-area peak discharge, flashiness (R-B Index), and runoff ratio. Stream hydrology was strongly affected by watershed percent IS. For the current condition simulation, values for all three indices were five to seven times greater in the most developed watershed compared to the least developed watershed. The climate change scenario caused a 20.8 % increase in unit-area peak discharge on average across the five watersheds compared to the current condition simulation. The land cover change scenario resulted in large increases for all three indices: 49.5 % for unit-area peak discharge, 39.3 % for R-B Index, and 73.9 % for runoff ratio, on average, for the five watersheds. The combined climate and land cover change scenario resulted in even greater increases for all three indices: 80.1 % for unit-area peak discharge, 43.7 % for R-B Index, and 74.5 % for runoff ratio, on average, for the five watersheds. The scenarios for different distributions of land

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cover change within one watershed resulted in changes for all three indices, with an 18.4% increase in unit-area peak discharge for the midstream scenario, and 17.5% (downstream) and 18.1% (midstream) increases in R-B Index, indicating sensitivity to the location of potential additions of IS within a watershed. Given the likelihood of increased precipitation in the future, land use planning and policy tools that limit expansion of impervious surfaces (e.g. by substituting pervious surfaces) or mitigate against their impacts (e.g. by installing bioswales) could be used to minimize negative effects on streams.

1 Introduction

The hydrology of urban streams is responsive to human activities in the surrounding landscape (Arnold and Gibbons, 1996; Walsh et al., 2005; Wenger et al., 2009). Compared to streams in more natural settings, urban streams are located in landscapes associated with less infiltration and more surface runoff, often leading to greater peak discharge and shorter peak discharge lag times (Anderson, 1970; Arnold and Gibbons, 1996). Declines in stream water quality and ecological condition, such as increases in pollutant and nutrient concentrations (e.g. Hatt et al., 2004; Pekarova and Pekar, 1996), and shifts in organismal assemblages to more eutrophic species (e.g. Black et al., 2011; Walsh et al., 2001) have commonly been reported for urban streams. Accelerated climate change (Denault et al., 2006), land use and land cover change (Grimm et al., 2008), and combinations of such changes (Nelson et al., 2009; Tong et al., 2012) are thought to be among the major driving factors leading to rapid degradation of urban stream systems.

Computer-based hydrological models have been used to better understand urban stream responses to potential stressors, such as projected changes in climatic conditions and land cover. Frequently used models include the Storm Water Management Model (or SWMM; Rossman, 2010; US EPA, 2011), the Hydrological Simulation Program-Fortran model (HSPF; Bicknell et al., 1997), and the Soil and Water

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Assessment Tool (SWAT; Neitsch et al., 2002), which can be used to predict hydrological responses to user-designed scenarios at relatively low cost. Of these models, SWMM has been applied in studies of urban streams because of its ability to simulate the hydraulic dynamics of artificial drainage systems that are prevalent in urban areas (e.g. Denault et al., 2006; Hsu et al., 2000; Meierdiercks et al., 2010). SWMM was developed to enable appropriate design of drainage systems (e.g. sizing for detention features, evaluating effectiveness of different runoff control strategies) and can be used to simulate dynamics of single events or for modeling on a continuous basis (US EPA, 2011). The model incorporates precipitation data to simulate surface runoff and pollutant outputs for sub-catchment areas which are then conveyed to the watershed outlet by a user-designated drainage system (US EPA, 2011).

Studies using hydrological models to predict stream response to changes in precipitation amounts and delivery patterns have used a variety of techniques to generate future scenarios. For example, in assessments of streamflow responses to climate change, Global Climatic Models (GCMs) and Regional Climate Models (RCMs) have been used (e.g. Jha et al., 2004; Nelson et al., 2009; Takle et al., 2010; Quilbe et al., 2008). However, the grid scales commonly used in GCMs (hundreds of km; Boyle, 1998) and RCMs (~ 50 km; Takle et al., 2010) and their time intervals (\geq hourly; Kendon et al., 2012) are not suitable for predictions at finer spatio-temporal scales. Other approaches that have been used at more local scales and for shorter time intervals include linear regression based on historical precipitation records (Denault et al., 2006; Takle and Herzmann, 2010) or projections based on likely proportional changes, for example, $\pm 20\%$ of current precipitation (Tong et al., 2012).

Similarly, a variety of approaches have been used to create projections for land cover change, including those based on Markov-chain probability models that generate both quantity and location of additional impervious surfaces (such as the software package *Land Change Modeler*, Eastman, 2012, as used by Bowman et al., 2012). Alternatively, logistic regression-based methods that incorporate historical land cover analyses combined with socioeconomic (population, land value) driver variables have also been used

(Gunalp et al., 2012; Serneels and Lambin, 2001). A third approach that has been used to predict land cover change is based on simple regression of historical changes in percent of developed land projected to a specified time in the future (Tu, 2009). Unlike the first two methods, land cover change using the third method is influenced only by the historical land cover characteristics, such as percent impervious surface (IS).

Some of the previously described prediction methods have been applied to investigate hydrological responses to climate change alone (Denault et al., 2006; Jha et al., 2004; Takle et al., 2010), land cover change alone (Meierdierks et al., 2010; Nagy et al., 2012; Rose and Peters, 2001) or to combined climate and land cover changes (e.g. Cuo et al., 2009; Hamdi et al., 2011; Tong et al., 2012; Yang et al., 2010). In general, these researchers reported greater variability in discharge (flashiness) and greater pollutant loading in response to increases in IS and/or precipitation. In a growing number of studies, urban stream responses to climate and/or land cover change have been examined for multiple watersheds. For example, using the ArcView Generalized Watershed Loading Function (AVGWLF) model, Tu (2009) predicted future climate and land cover changes and examined responses for streamflow and nitrogen loads at seven study sites in and near Boston, Massachusetts. Model outputs indicated that the greatest impacts from climate and land cover change were related to the seasonal distribution of discharge rate and nitrogen loading in the future, which were projected to be greater in fall and winter and lower in the summer, rather than affecting the average total annual amounts. In another study, Nagy et al. (2012) observed hydrological and water quality differences among 13 small watersheds located along the Florida Gulf Coast with different watershed percent IS. These researchers reported increases in peak flow, flashiness, pH, and specific conductance as impervious surface in the watershed increased (Nagy et al., 2012).

To date, however, relatively few studies have been conducted in Midwest USA to quantify responses of multiple urban streams to potential changes in both climate and land cover using a hydrological model specifically designed for use in urban environments (e.g. SWMM) and to examine a suite of variables to describe stream responses.

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In the research described in this paper, we examined five watersheds from among 20 small urban headwater streams for which we collected water quality and water quantity data over two years, 2011 and 2012. The five watersheds were purposefully selected to represent a gradient of percent IS, ranging from 5.3 to 37.1 %. Climate and land cover change were projected to the year 2040 using regressions based on historical data. We then used SWMM to create hydrological models for these watersheds to answer the following questions:

1. What hydrological differences can be detected among the five urban headwater streams along a % IS gradient?
2. What are the hydrological responses to projected climate change, land cover change, and combined climate and land cover change for these stream systems?
3. How might different distributions of land cover change affect urban headwater stream hydrology?

2 Methods

2.1 Study area

Five headwater stream watersheds located in Polk County, Iowa, USA were included in this study (WS1 to WS5, Fig. 1). These watersheds were within the corporate boundaries of four cities: Altoona, Ankeny, Johnston and Pleasant Hill. These cities are located close to Iowa's state capital (Des Moines), and have experienced rapid population growth in recent years. Between 2000 and 2010, populations in Altoona, Ankeny, Johnston, and Pleasant Hill increased by 41, 68, 100, and 73 %, respectively, compared to a 4 % increase for Iowa as a whole (State Data Center of Iowa, 2012). The five watersheds were located within the Upper Midwest climatic region, with average annual precipitation of 838 mm (National Climatic Data Center, 2012). About 75 % of annual precipitation typically occurs between April and September. The watersheds

were located along the southern edge of Des Moines Lobe landform region, a recently glaciated area (14 kyr BP) in which stream network development is ongoing. The watersheds were approximately 280 m a.s.l. and had an average slope of about 8 % (Table 1).

Among the five watersheds, watershed area ranged from 61.8 to 269.8 ha, % IS ranged from 5.3 to 37.1 %, and average slope ranged from 5.4 to 10.5 % (Table 1). Two of the watersheds (WS1 and WS2) were located in Pleasant Hill (Fig. 1). Land cover in WS1 included residential development (clustered in the upstream area), agricultural land (midstream area), and pasture land (downstream area). The second watershed (WS2) contained a segment of US Highway 65 and was otherwise dominated by agricultural and forested areas. The third watershed (WS3) was in northeastern Altoona, along the eastern boundary of the city, and it contained residential, commercial, and agricultural land. The fourth watershed (WS4) was located in Johnston, and the fifth (WS5) was in Ankeny. Both WS4 and WS5 contained primarily residential areas. Although their drainage densities were similar, WS4 was characterized by a lower % IS than WS5.

2.2 Stream monitoring methods and other data sources

Flow rates were measured twice per month from April to October at defined channel cross-sections near each stream outlet using a FLO-MATE 2000 Water Current and Flowmeter™ (Hach Company, Loveland, CO). Discharge was then determined using the cross-section method of Rantz (1982). Area of the defined channel cross-section was determined by measuring stream depth at evenly distributed points (varying from one to seven) and multiplying by width. The maximum distance between adjacent measuring points was 0.5 m.

In addition, stream stage was continuously recorded at 5 min intervals from mid-May to October at each cross-section using HOBO U20 water level data loggers (Onset Computer Corporation, Inc., Pocasset, MA). An additional data logger was used to measure barometric pressure for correction of stream stage data. Continuous stage

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data were converted to discharge using a stage-discharge rating curve developed for each stream.

Digital Elevation Models (DEMs) at one-meter resolution were generated from Light Detection and Ranging (LiDAR) data available from the Iowa LiDAR Mapping Project (GeoTREE, 2011). Impervious surface cover for each watershed was manually digitized based on 2011 aerial photo images. Storm sewer GIS layers were obtained from staff members in the four cities. Five-minute interval precipitation data were obtained from Iowa SchoolNet system (Iowa Environmental Mesonet, 2012). Three weather stations were selected based on proximity to the five watersheds, “SDRI4” in Des Moines (for WS1, WS2, and WS3), “SGRI4” in Johnston (for WS4), and “SAKI4” in Ankeny (for WS5).

2.3 Calibration and validation of the Storm Water Management Model

The Storm Water Management Model (SWMM Version 5.0.022; Rossman, 2010; US EPA 2011) was used to simulate current and projected watershed conditions. Sub-catchments were delineated to collect precipitation and the kinematic wave method (Rossman, 2010) was used to route water through designated channels or pipes. Individual models were constructed for each of the five watersheds using 5 min interval precipitation data to simulate surface runoff and channelized discharge in road gutters, storm sewers, and surface channels. The embedded groundwater module was activated for all models using the same precipitation event (6.4 mm delivered to each watershed 5 h before each simulation) to “recharge” groundwater. Because we were interested in flow dynamics associated with single precipitation events, the effects of evaporation were not included (Gironás et al., 2009). Infiltration processes were simulated using Horton’s equation within the model (Green, 1986).

Model parameters were obtained in three different ways. The first set of parameters were defined based on existing data for sub-catchment and drainage structure parameters, % IS, and slope. The second set of parameters were calibrated using available discharge observations, sub-catchment width, coefficients for groundwater equations,

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and Manning's roughness coefficient (n) for impervious surfaces, pervious surfaces, and channels. All other parameters were set to default values or values suggested by the SWMM application manual (Gironás et al., 2009). For example, the initial infiltration rate was 100 mm h^{-1} , the constant infiltration rate was 7 mm h^{-1} , and the decay constant was 3.5 for Horton's infiltration equation.

Precipitation events (26 June 2011) recorded at the three weather stations were chosen to use for the calibration process. Precipitation depths were 8.9 mm (WS1, WS2, and WS3), 10.4 mm (WS4), and 12.7 mm (WS5). The models were then manually calibrated for best fit to the continuous discharge data derived from field monitoring for the same events. Precipitation events from the three weather stations on another date (25 May 2011) were used to validate the models. Precipitation depths for the validation procedure were 21.3 mm (WS1, WS2, and WS3), and 23.1 mm (WS4 and WS5).

Model performance was quantified using the coefficient of determination (R^2) and Nash–Sutcliffe model efficiency coefficient (NSE). The coefficient of determination ranges from 0 to 1, where greater values indicate a closer relationship between predicted and observed values for discharge. The NSE statistic has a range of $-\infty$ to 1. A greater value indicates a better prediction of discharge, shown in Eq. (1):

$$\text{NSE} = 1 - \frac{\sum(Q_{o,t} - Q_{m,t})^2}{\sum(Q_{o,t} - \overline{Q_o})^2} \quad (1)$$

where NSE is the Nash–Sutcliffe model efficiency coefficient, $Q_{o,t}$ is the observed discharge ($\text{m}^3 \text{ s}^{-1}$) at time t , $Q_{m,t}$ is the modeled discharge ($\text{m}^3 \text{ s}^{-1}$) at time t , and $\overline{Q_o}$ is the average for the observed discharge ($\text{m}^3 \text{ s}^{-1}$).

2.4 Climate change projection

A precipitation event occurring on 10 June 2011 was used as the current climate condition for all five watersheds. This event delivered 16.8 mm of precipitation in one hour. To create a future (2040) precipitation event, annual precipitation for the region was

obtained for the period 1895 to 2011 (National Climatic Data Center, 2012). Using linear regression (similar to method of Denault et al., 2006; Takle and Herzmann, 2010), annual precipitation in 2040 was projected to be 18% more than it was during 2011. This proportional increase is consistent with results reported for this region of study by previous researchers (e.g. Jha et al., 2004; Takle et al., 2010). The projected precipitation event was designed to have the same duration and time distribution as the precipitation event on 10 June 2011.

2.5 Land cover change projection

Percent IS in 2011, calculated by dividing IS area within each watershed by the corresponding total watershed area, was used as the current land cover condition. Land cover was projected to the year 2040 using separate regression models based on quantification of total IS areas for each of the four cities (Altoona, Ankeny, Johnston, and Pleasant Hill) in 1940, 1961, 1990, 2002, and 2011 (method adapted from Tu, 2009). Binomial curves provided the best fit for the four cities, with coefficients of determination (R^2) of 0.997 (Altoona), 0.984 (Ankeny), 0.973 (Johnston), and 0.994 (Pleasant Hill). The increase of IS area within each watershed (Table 1) was assumed to be the same as that for each city, unless the projected % IS for the watershed exceeded 45% (which we set as a maximum according to the % IS for other fully developed residential areas in our study area). For the land cover change simulation that included all five watersheds, increases in IS area were evenly distributed across each watershed.

2.6 SWMM simulations with independent and combined effects of climate and land cover change

We conducted current condition SWMM simulations using a single precipitation event (10 June 2011, 16.8 mm precipitation in one hour) and 2011 land cover data in all of the calibrated SWMM models (Table 2). A set of three different climate and land cover change scenarios were designed for this part of the study. In the first scenario,

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we used a precipitation event projected for 2040 with actual 2011 land cover. In the second scenario, we used the actual 2011 precipitation events and the projected land cover for 2040. In the third scenario, we used 2040 projections for both climate and land cover.

2.7 SWMM simulations with different distributions of land cover change in a single watershed

To assess effects of different distributions of future land cover changes, one watershed (WS4) was divided into three sections (Fig. 2). The three sections were downstream, midstream, and upstream areas within the watershed, and characterized by similar size and initial % IS (2011). In this watershed, existing impervious surfaces (2011) were relatively evenly distributed between 10 % and 95 % distances to the watershed outlet. The same increase in urban land cover (to cause a 16.7 % absolute increase for each section) was applied within one of the three sections, changing the impervious surface distribution from upstream to downstream. The projected precipitation event for 2040 in WS4 was used for each simulation in this set of scenarios.

2.8 Quantifying hydrological indices

To quantify hydrological responses, three indices were calculated for current condition and scenario simulations, including unit-area peak discharge, Richards–Baker Index (hereafter R-B Index; Baker et al., 2004), and runoff ratio. Unit-area peak discharge is the quotient of peak discharge divided by watershed area, and indicates the greatest amount of discharge generated by a unit area in a single precipitation event. A greater value of peak discharge indicates greater potential for flooding (Huong and Pathirana, 2013). The R-B Index measures oscillations in discharge relative to total discharge, also referred to as “flashiness”. A higher value of the R-B Index indicates a greater difference between high and low flows, which may be linked to changes in channel morphology, water quality and habitat structure of stream ecosystems (Shields et al.,

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2010; Violin et al., 2011). We calculated the R-B Index based on 5 min interval discharge data, as per Eq. (2):

$$\text{R-B Index} = \frac{\sum_{i=1}^n |q_i - q_{i-1}|}{\sum_{i=1}^n q_i} \quad (2)$$

where n is the total number of discharge records, and q_i is the i th measured discharge of a stream, and q_{i-1} is the $i - 1$ th measured discharge of a stream.

The runoff ratio is the total discharge depth divided by total precipitation depth, which indicates the proportion of precipitation that is discharged in surface channels. A higher value of runoff ratio indicates increase of surface runoff and may result in decreases in groundwater level because of less infiltration (Foster and Chilton, 2004).

3 Results

3.1 Calibration and validation of the Storm Water Management Model

Calibrated model parameters (Table 3) were used to generate hydrographs for each watershed (Fig. 3). Generally, calibration and validation hydrographs demonstrated acceptable fit between observed and simulated discharge, although some differences in timing (WS5, calibration) and/or magnitude (WS1, validation) were observed. For calibration simulations, coefficients of determination ranged from 0.73 to 0.89, and NSEs ranged from 0.25 to 0.80 (Fig. 3). Coefficients of determination for validation simulations ranged from 0.39 to 0.92, and NSEs ranged from 0.23 to 0.91. Among the models, validation statistics were somewhat lower (although still acceptable as per Moriasi et al., 2007) for watershed WS1.

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3.2 Current condition simulation

Current condition simulations describe existing (2011) hydrological dynamics for the five watersheds (Table 4). Compared to other watersheds, WS1 was characterized by the lowest values for unit-area peak discharge ($0.29 \times 10^{-6} \text{ m s}^{-1}$), R-B Index (flashiness; 0.015), and runoff ratio (0.053). As % IS increased, watershed simulations generated consistently greater values for all three indices. In WS5, the three indices were $2.33 \times 10^{-6} \text{ m s}^{-1}$, 0.130, and 0.355, respectively, five to seven times greater than those for WS1.

3.3 Independent and combined effects of climate and land cover change simulations

In the climate change scenario, an 18% increase in precipitation generated increases in all three hydrological indices compared to the current condition (e.g. Fig. 4 for WS4). Specifically, unit-area peak discharge for the five watersheds ranged from $0.35 \times 10^{-6} \text{ m s}^{-1}$ to $2.82 \times 10^{-6} \text{ m s}^{-1}$, an increase of 20.8% on average compared to current condition values (Table 4). The change in unit-area peak discharge was greatest in WS3 (23.4%). Increases in R-B Index and runoff ratio compared to current condition values were much smaller than those for unit-area peak discharge. On average, the R-B Index increased 3.7%, ranging from 0.015 to 0.136, with the greatest proportional increase occurring in WS4 (8.3%). Watersheds with higher % IS generally had greater values for R-B Index. Runoff ratio increased 0.4% on average, ranging from 0.053 to 0.358, with the greatest proportional increases occurring in WS4 and WS5 (0.7%).

The land cover change scenario led to a greater hydrological response compared to the current condition than did the climate change scenario (e.g. Fig. 4 for WS4). All three response indices increased to a greater degree than they did in the climate change scenario (Table 4). Unit-area peak discharge ranged from 0.58 to 2.80 (an average increase of 49.5%). The greatest increase compared to the current condition

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(99.7%) occurred in WS1. Increases in the R-B Index ranged from 0.019 to 0.148, with an average increase of 39.3%. The greatest increase (68.4%) occurred in WS3. Runoff ratio ranged from 0.105 to 0.430, a 73.9% average increase, with the greatest increase (97.9%) detected for WS2.

The combined effects of climate and land cover changes generated the largest changes in the hydrograph (e.g. Fig. 4 for WS4) and for all three indices (Table 4). Unit-area peak discharge increased from 0.69 to 3.40, with an average increase of 80.1%. The greatest increase (138.0%) occurred in WS1. The R-B Index ranged from 0.019 to 0.156 (an average increase of 43.7%), with the largest increase (66.9%) in WS3. The runoff ratio ranged from 0.105 to 0.433 (an average increase of 74.5%), with the largest increase (98.9%) in WS2.

3.4 Simulations of different distributions of land cover change for WS4

For different distributions of land cover change in WS4, we observed consistent increases but different patterns of change for each of the hydrological indices (Table 5). The unit-area peak discharge responded most strongly to additional % IS in the mid-stream area, increasing by 18.4%, compared to the other two scenarios (16.1 and 15.6%; Table 5). The response of the R-B Index to greater % IS in the upstream area was much lower (12.0%) compared to the other two scenarios (17.5 and 18.1%). Increases in runoff ratios were similar for the three scenarios (ranging from 19.3 to 20.2%; Table 5).

When urban land cover was added to the downstream portion of the watershed, the three indices increased (16.1, 17.5, and 19.3%, respectively) compared to the current condition. When urban land cover increased in the midstream area of the watershed, the unit-area peak discharge and R-B Index increased the most of the three scenarios tested (18.4 and 18.1%, respectively). When urban land cover increased in the upstream portion of the watershed, runoff ratio increased the most among the three scenarios, from 0.268 to 0.322 (a 20.2% increase).

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Hydrographs for the initial climate change scenario and the three different distributions of land cover change (Fig. 5) indicated that: (1) discharge rates increased more rapidly during the early stages (at 0.33 h) of the downstream scenario (Box A), and discharge recession was slower for the upstream scenario (Box E); (2) lower peak discharge occurred for the upstream scenario at the first (Box B) and the third peak (Box D), linked to smaller increases in R-B Index; and (3) the greatest overall unit-area peak discharge occurred during the second peak for all three scenarios, among which midstream scenario generated the greatest unit-area peak discharge (Box C).

4 Discussion

In this study, we calibrated and validated SWMM for each of five urban headwater stream watersheds that represent a gradient in % IS. Calibration and validation of the five models demonstrated acceptable fit between measured and simulated discharge. The current condition simulation illustrated the influence of increased % IS among these watersheds. Of the response metrics we analyzed, simulations of climate change had the greatest effect on unit-area peak discharge. Simulations of land cover change led to relatively large increases in all three hydrological indices. Simulations for combined climate and land cover change caused greater changes to the three hydrological indices than either climate or land cover change alone. Unit-area peak discharge and R-B Index were the most sensitive to different spatial distributions of additional IS.

4.1 Calibration and validation of the storm water management model

Calibration and validation results indicated that SWMM was able to simulate hydrological processes in the small watersheds we studied, and was sensitive to the gradient in initial % IS among them (Fig. 3). Of the five watersheds, WS1 and WS3 were separated by the greatest distance from a weather station (Table 1), which resulted in time lags

between recorded and actual precipitation (Fig. 3) which led to the generally lower (but still acceptable) validation statistic values for these watersheds.

4.2 Current condition simulations

Current condition hydrological responses indicated that unit-area peak discharge, R-B Index, and runoff ratio all increased along the % IS gradient that characterized these watersheds (Table 4), a result that corroborates the hypotheses and work of a number of researchers (e.g. Arnold and Gibbons, 1996; Nagy et al., 2012; Schueler, 1994; Yang et al., 2010).

Among much earlier studies documenting the effects of % IS on urban streams, distinct hydrological impacts were reported to occur when watershed % IS reached a threshold of 10 % (Arnold and Gibbons, 1996; Booth and Jackson, 1997; Schueler and Holland, 2000). We found important differences for all three hydrological response indices that were detectable between 5.30 % IS (WS1) and 7.96 % IS (WS2), especially for unit-area peak discharge (Table 4). These results indicate that important hydrological responses can occur below the often-cited threshold, in our case below 8 % IS. Other researchers have also recently reported hydrological changes at relatively low % IS in lower-order watersheds, suggesting threshold criteria of about 5 % IS (based on 13 watersheds studied by Nagy et al., 2012 along the Florida Gulf Coast), or even 3 % IS (based on 16 watersheds examined by Yang et al., 2010 in the White River Basin in Indiana). It may be that focusing on small headwater streams allows detection of these effects at lower % IS thresholds, or that these relatively small stream systems are more sensitive to % IS.

A number of other factors important to urban stream hydrology include slope, distribution of IS, and storm sewer system density and structure (Booth and Jackson, 1997; Dingman, 2008; Meierdiercks et al., 2010; Mejia and Moglen, 2010). Given variation in these watershed parameters among the five study watersheds (Table 1), the consistent trend of changes in the three hydrological response indices corroborates other

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evidence that % IS can be a robust indicator of the effects of urban land cover on stream hydrology (as earlier suggested by Paul and Meyer, 2001; Schueler et al., 2009).

4.3 Climate change simulations

In previous studies, hydrological impacts have been interpreted by determining changes in continuous discharge over relatively long time scales (e.g. Franczyk and Chang, 2009; Jha et al., 2004). In this study, however, we used single precipitation events to examine more detailed hydrological responses to climate change. The SWMM simulations we conducted indicate that climate change (increased precipitation with other factors held constant) will have greatest effects on unit-area peak discharge (20.8 % average increase) compared to R-B Index (a 3.7 % average increase) or runoff ratio (a 0.4 % average increase). Thus, for already developed watersheds, stormwater mitigation strategies that address peak discharge rates should be a priority. A possible strategy that addresses peak discharge is stormwater system retrofitting to delay delivery (e.g. Karamouz et al., 2011), such as integration of wet ponds in a stormwater treatment train (e.g. Villarreal and Bengtsson, 2004). Increased precipitation resulted in greater response for all three indices along the gradient of % IS from WS1 (5.30 % IS) to WS3 and WS4 (18.21 to 28.28 % IS). Thus, at higher initial % IS, stormwater management strategies to increase infiltration to mitigate potential increases in flashiness and runoff ratio become more important.

4.4 Land cover change simulations

Our watershed simulations for land cover change indicated that, compared to the climate change scenario, land cover change resulted in increased values for all three hydrological indices. Our findings are generally consistent with previous reports for increases in peak discharge (from 30 to 100 %, e.g. Rose and Peters, 2001), runoff ratio (from 21 to 45 %, but more sensitive to watershed slope; Rose and Peters, 2001) and R-B Index (15 %, Yang et al., 2010) in response to greater amounts of impervious land

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cover. In their examination of small watersheds along the Gulf Coast, Nagy et al. (2012) reported greater effects of increasing % IS on peak discharge, but lesser effects on R-B Index. It is likely that the relative importance of changes in climate and land cover on stream hydrology are subject to the specifics of scenario designs (e.g. differing interpretations offered by Hamdi et al., 2011; Tong et al., 2012; Tu, 2009). Notwithstanding, our land cover change simulations provide further support for the contention that important hydrological changes occur at thresholds below 10 % IS, based on consistent increases in all three response indices for the predicted impervious surface change from 5.30 to 10.47 % in WS1 and from 7.96 to 15.74 % in WS2.

4.5 Combined climate and land cover change simulations

The combined climate and land cover change simulations led to changes in response indices that were slightly greater than additive effects for climate and land cover change considered independently (Table 4). Although previous work has documented additive effects of combined climate and land cover changes (e.g. Nelson et al., 2009; Tong et al., 2012; Tu, 2009), conclusions about which influence might be stronger (either climate or land cover change) have been inconsistent, possibly owing to differences in prediction methods. For example, Tong et al. (2012) reported that climate change (assuming a 2 % increase in temperature, and 20 % increase in precipitation) led to a greater change in mean daily flow than did land cover change when both were projected to 2050. However, when their climate change assumption was altered (to a 4 % increase in temperature with associated increases in evaporation and evapotranspiration, and 20 % increase in precipitation), the increase in mean daily flow due to climate change was less than that due to land cover change.

In our study, we projected a 1 h precipitation event with an 18 % increase from a 16.8 mm (current condition) event to 19.8 mm. The projected 2040 precipitation event was equivalent to a 1 h, 4 month recurrence interval event in this region at the present time (Huff and Angel, 1992). We selected and projected this event relatively conservatively to avoid flooding during the simulations, which would preclude detection of

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changes in the response variables. Thus, although it appears that land use change has a stronger influence in our case, it may be only because the climate change scenario was constrained by the relatively small change to predicted precipitation.

4.6 Simulations of different distributions of land cover change in WS4

5 The differentially allocated land cover change scenarios for WS4 indicated that, although all three hydrological indices increased from the initial climate change scenario, only unit-area peak discharge and R-B Index appeared to respond differently to increases in % IS applied in the downstream, midstream, or upstream areas of the watershed. Thus, it appears that the location of IS additions primarily affects the timing, rather than amount, of discharge conveyed to the watershed outlet (e.g. WS4 curve in Fig. 5, where Box A indicates greater unit-area discharge early in the storm event for the downstream scenario due to the short path to the watershed outlet, and Box E indicates a slower recession for the upstream scenario due to a longer flow path).

15 Assuming a given increase of % IS in a watershed, any scenario that minimizes increases in peak discharge and flashiness would be better for the in-stream environment, since greater values for these indices are related to stream dynamics that can cause bank erosion and lead to poorer quality aquatic habitat (Walsh et al., 2005; Wenger et al., 2009). Thus, given the existing IS distribution and stormwater conveyance structures, addition of impervious surfaces in the upstream section of WS4 would likely have less impact on stream hydrology and ecology. Using an urban hydrological model (e.g. SWMM in our study) to assess scenarios for future urban development in a range of watershed types could provide land use planners with critical information for decision making to better protect urban streams.

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5 Conclusions

We used SWMM to simulate hydrological responses of five headwater streams to increases in precipitation and urban land cover based on projections to the year 2040. We conclude that:

1. The current condition simulations indicated that watersheds that have greater % IS are also characterized by greater values in unit-area peak discharge, R-Blindex, and runoff ratio. Given variation in other important characteristics among the study watersheds (e.g. average slope, watershed size, and drainage density) % IS was a reliable indicator of impacts on urban stream hydrology. Efforts to mitigate negative impacts to stream hydrology in urban areas should include specific attention to strategies that limit additional IS and that minimize connectivity among existing impervious surfaces. We also detected important hydrological responses below the often-cited 10 % IS threshold, in our case below 8 % IS.
2. The climate change scenario in this study had strongest effects on unit-area peak discharge in these watersheds. All three hydrological indices were affected by the land cover change scenarios used in this study. The combined climate and land cover change scenarios resulted in slightly more than additive effects from climate and land cover change alone. These findings confirmed that urban stream hydrology, especially for unit-area peak discharge, is highly sensitive to expected changes in climate and land cover. The capacity of existing stormwater drainage/infiltration systems should be continuously evaluated and incorporated in comprehensive planning at municipal and regional levels.
3. Simulations for different distributions of land cover change demonstrated that the location of IS additions has a greater effect on the timing of delivery than on total amount of discharge. The ability to detect hydrological impacts associated with specific placement of impervious surfaces indicates that this simulation method

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could be very useful in identifying locations for development that would minimize stream degradation in small urban watersheds.

The result that hydrological responses of these streams to projected land cover changes were much greater than those generated by projected climate change is due, in part, to the conservative approach to climate change that we used to develop the models. In spite of this limitation, simulations for this set of watersheds indicated that plausible changes in both climate and land cover will have strong effects on urban stream hydrology. In addition, simulations for different distributions of IS in WS4 were also constrained by pre-existing IS and stormwater conveying systems in that watershed. In spite of these limitations, this study extends our knowledge of the hydrological dynamics of lower-order urban streams, and elucidates several potential cause and effect relationships that can be used to manage urban landscapes to reduce negative impacts to urban stream ecosystems.

Acknowledgements. We would like to thank Anna Whipple, City of Des Moines, for access to land cover data, and city staff members in Altoona, Ankeny, Johnston, and Pleasant Hill for their help in obtaining storm sewer data. This research was supported by the Chinese Scholarship Council, the Center for Global and Regional Environmental Research at the University of Iowa, the State of Iowa, and McIntire–Stennis funding.

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Table 1. Geographic characteristics, drainage density, distance to nearest weather station, and projected percent impervious surface (IS, 2040) for five urban watersheds (WS1 to WS5) in central Iowa representing a percent impervious surface gradient. Total drainage density is the sum of road gutter, storm sewer and surface channel densities.

	WS1	WS2	WS3	WS4	WS5
Area (ha)	194.9	61.8	269.8	89.5	92.0
Current percent IS	5.3	8.0	18.2	28.3	37.1
Average slope (%)	10.3	5.4	5.6	10.5	8.8
Total drainage density (km km ⁻²)	3.2	10.7	5.1	9.0	8.8
Road gutter density (km km ⁻²)	–	–	0.6	0.3	0.3
Storm sewer density (km km ⁻²)	0.2	–	1.3	6.7	6.5
Surface channel density (km km ⁻²)	3.0	10.7	3.3	2.0	1.9
Distance to nearest weather station (km)	10.4	5.8	15.0	5.3	3.1
Land cover projection					
Percent IS in 2040	10.5	15.7	35.3	45.0	45.0
Absolute change (%)	5.2	7.8	17.1	16.7	7.9
Relative change (%)	97.6	97.7	93.7	59.1	21.2

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Table 2. Current and predicted scenarios for SWMM model simulation using current (2011) and future (2040) climate and land cover conditions for five watersheds in central Iowa. The different distributions of land cover change (partial) were only applied to the fourth watershed (WS4).

	Watersheds included	Scale of simulation	Land cover year	Climate year
Current condition	All	Whole watershed	2011	2011
Predicted scenarios				
Climate	All	Whole watershed	2011	2040
Land cover	All	Whole watershed	2040	2011
Climate and land cover	All	Whole watershed	2040	2040
Land cover (partial)	WS4	Subsection, downstream area	2040	2040
Land cover (partial)	WS4	Subsection, midstream area	2040	2040
Land cover (partial)	WS4	Subsection, upstream area	2040	2040

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Table 3. Calibrated SWMM model parameters for five watersheds (WS1 to WS5) in central Iowa. *A1*, *B1*, *A2*, and *B2* are coefficients for the groundwater equation in SWMM.

Parameters and Statistics	WS1	WS2	WS3	WS4	WS5
Number of sub-catchments	29	32	49	52	60
Average width of sub-catchments (m)	704	259	324	101	123
Manning's <i>n</i> (impervious surfaces)	0.017	0.05	0.05	0.05	0.05
Manning's <i>n</i> (pervious surfaces)	0.18	0.5	0.2	0.4	0.4
Manning's <i>n</i> (channels: natural)	0.05	0.05	0.04	0.05	0.04
Manning's <i>n</i> (channels: storm sewer)	–	–	0.012	0.012	0.012
Manning's <i>n</i> (channels: gutter)	0.015	–	0.015	0.015	0.015
Parameters for groundwater equation					
<i>A1</i>	0.0005	0.050	0.0070	0.050	0.008
<i>B1</i>	1.00	1.00	2.00	1.00	1.00
<i>A2</i>	0.0005	0.007	0.0010	0.050	0.008
<i>B2</i>	1.00	1.00	1.00	1.00	1.00

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Table 4. Hydrological response characteristics for current condition, climate change, land cover change, and combined scenarios for five watersheds (WS1 to WS5) in central Iowa. The relative changes in percent for climate, land cover and combined scenarios are calculated compared to the current condition.

Scenarios	Indices	WS1	WS2	WS3	WS4	WS5
Current condition	Unit-area peak discharge (10^{-6} ms^{-1})	0.29	0.88	1.37	1.62	2.33
	R-B Index	0.015	0.027	0.060	0.061	0.130
	Runoff ratio	0.053	0.077	0.181	0.266	0.355
Climate	Unit-area peak discharge (10^{-6} ms^{-1})	0.35	1.02	1.69	1.97	2.82
	Change (%)	21.1	16.8	23.4	21.8	21.1
	R-B Index	0.015	0.027	0.061	0.066	0.136
	Change (%)	1.7	2.0	2.0	8.3	4.7
	Runoff ratio	0.053	0.077	0.181	0.268	0.358
	Change (%)	0.0	0.6	0.1	0.7	0.7
Land cover	Unit-area peak discharge (10^{-6} ms^{-1})	0.58	1.17	2.03	2.37	2.80
	Change (%)	99.7	33.7	47.8	46.4	20.1
	R-B Index	0.019	0.037	0.102	0.091	0.148
	Change (%)	27.2	38.5	68.4	48.5	14.1
	Runoff ratio	0.105	0.152	0.350	0.424	0.430
	Change (%)	97.8	97.9	93.6	59.1	21.0
Climate and land cover	Unit-area peak discharge (10^{-6} ms^{-1})	0.69	1.37	2.46	2.92	3.40
	Change (%)	138.0	56.9	79.8	80.0	46.0
	R-B Index	0.019	0.038	0.101	0.098	0.156
	Change (%)	29.9	40.9	66.9	60.0	20.7
	Runoff ratio	0.105	0.153	0.351	0.427	0.433
	Change (%)	97.8	98.9	93.8	60.2	21.9

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Table 5. Hydrological response characteristics for different distributions of land cover change in three sections (downstream, midstream, and upstream areas) of one watershed, WS4. The three simulations use the predicted (2040) climate parameters, percent changes for indices are compared to the initial climate change scenario for that watershed.

	Initial Climate Change Scenario	Downstream	Midstream	Upstream
Unit-area peak discharge ($\times 10^{-6} \text{ m s}^{-1}$)	1.97	2.29	2.33	2.28
Change (%)		16.1	18.4	15.6
R-B Index	0.066	0.078	0.078	0.074
Change (%)		17.5	18.1	12.0
Runoff ratio	0.268	0.320	0.321	0.322
Change (%)		19.3	19.6	20.2

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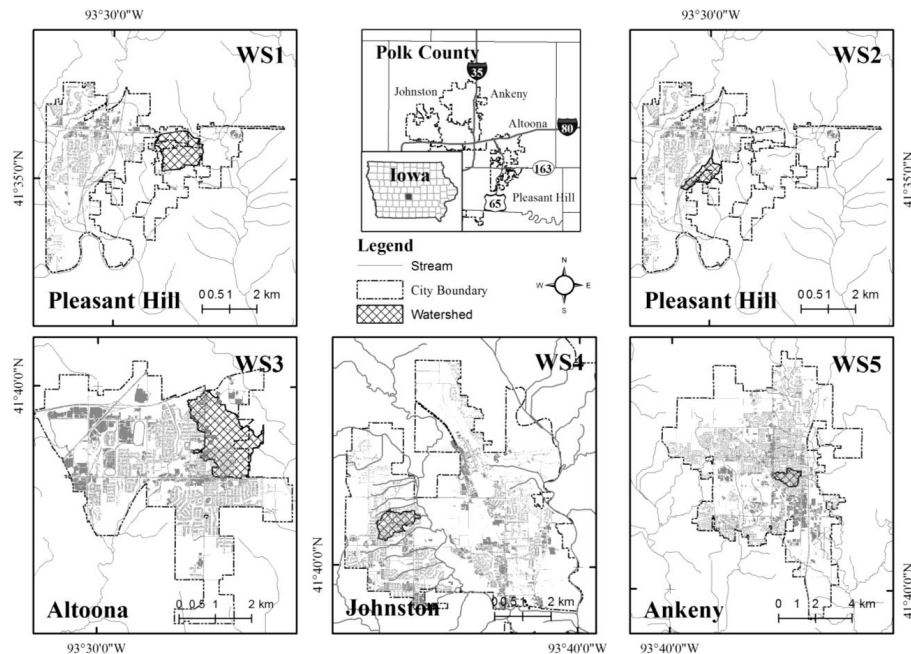


Fig. 1. Five headwater stream watersheds located in four cities (one each in Altoona, Ankeny, and Johnston, and two in Pleasant Hill) in Polk County, central Iowa. Shaded areas represent impervious surface in 2011, watershed areas for WS1 through WS5 are delineated with cross-hatching.

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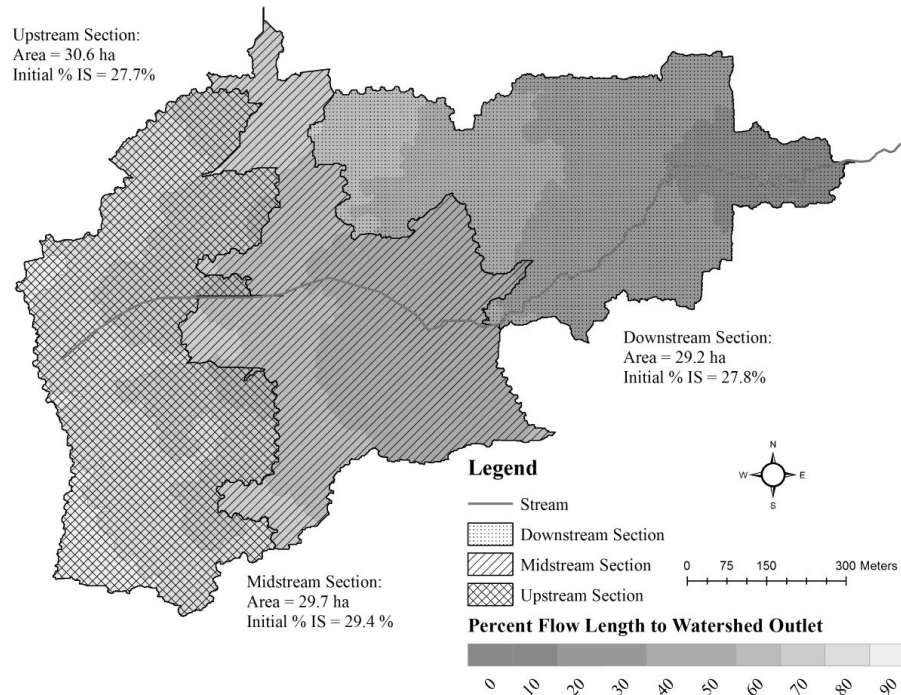


Fig. 2. Boundaries of the three sections for the different distributions of impervious land cover scenario for WS4. The three sections have similar area and percent impervious surface in 2011 (initial % IS). The downstream section includes 0 to 40 % of flow length to the outlet, the midstream section includes 40 to 70 %, and the upstream section includes 70 to 100 % of flow length to the outlet.

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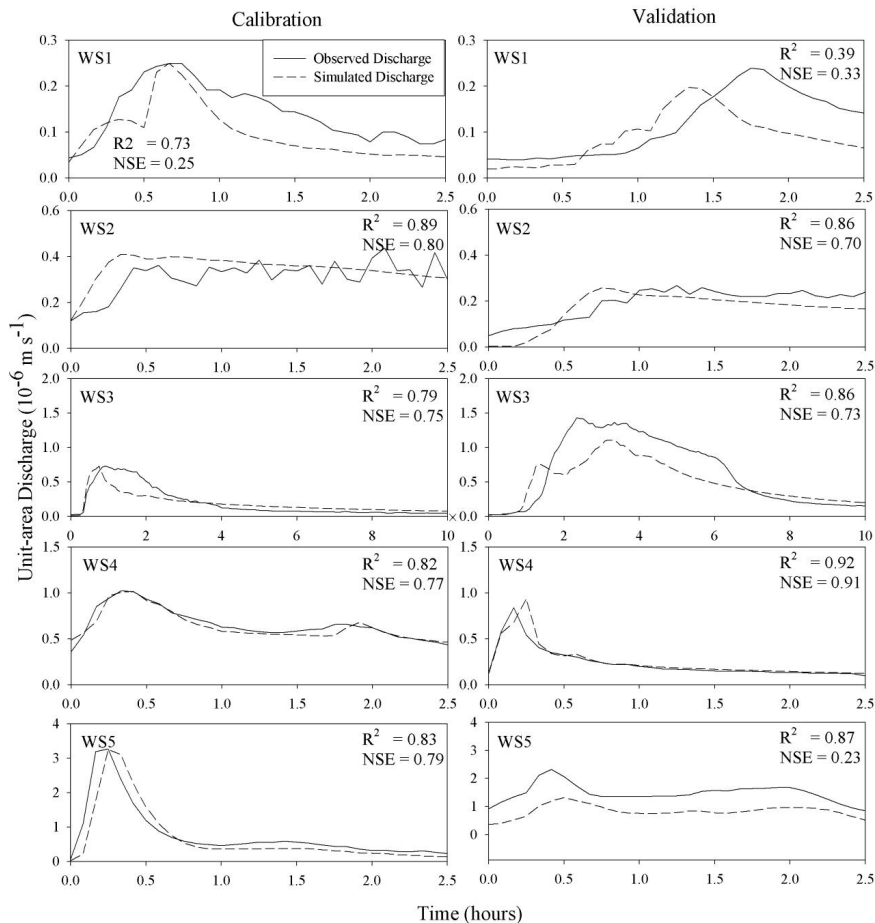


Fig. 3. Hydrograph segments for calibration (left column) and validation (right column) of the five models (WS1 to WS5). Discharge was standardized by watershed area.

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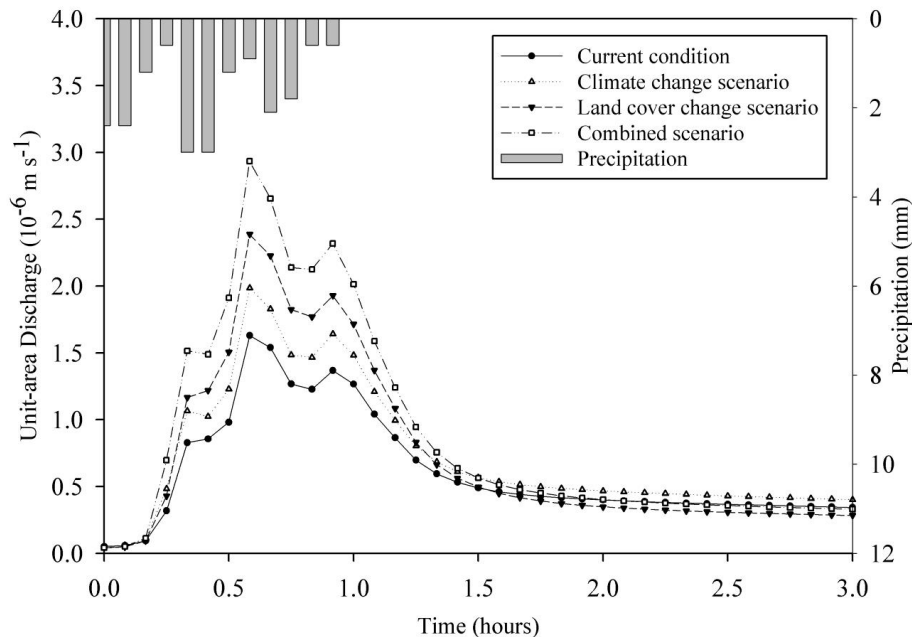


Fig. 4. Precipitation and hydrographs for the current condition, climate change, land cover change and combined climate and land cover change scenarios for WS4.

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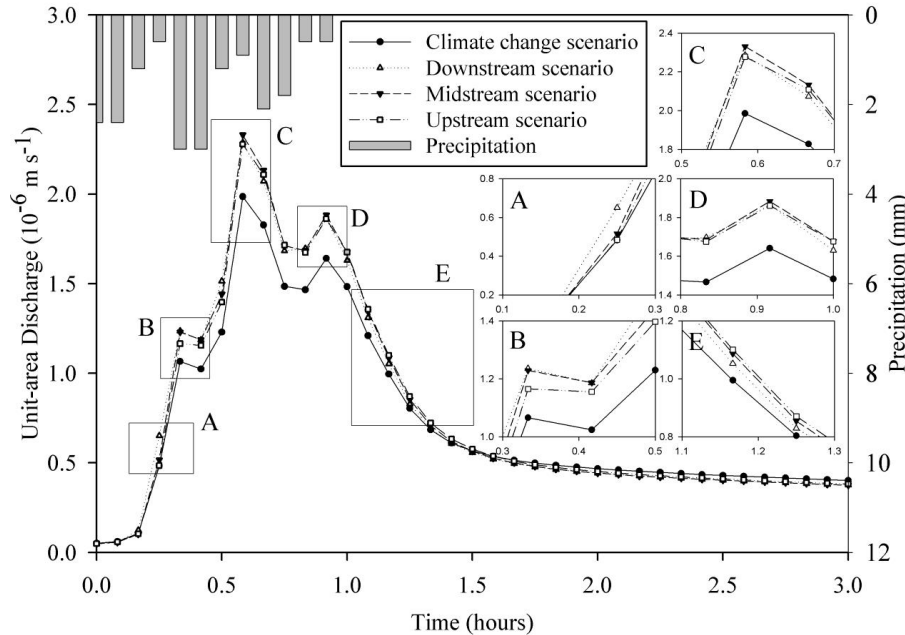


Fig. 5. Precipitation amounts and hydrographs for different distributions of land cover change for WS4. Discharge was standardized by watershed area. Box A illustrates greater discharge for the downstream scenario; Box B and Box D indicate lower peak discharge for the upstream scenario; Box C illustrates greatest peak discharge for the midstream scenario; and Box E indicates greater discharge for the upstream scenario.

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