

Modelling groundwater-dependent vegetation patterns using ensemble learning

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Received: 17 September 2007 – Published in Hydrol. Earth Syst. Sci. Discuss.: 2 October 2007

Revised: 12 February 2008 – Accepted: 14 February 2008 – Published: 19 March 2008

Abstract. Vegetation patterns arise from the interplay between intraspecific and interspecific biotic interactions and from different abiotic constraints and interacting driving forces and distributions. In this study, we constructed an ensemble learning model that, based on spatially distributed environmental variables, could model vegetation patterns at the local scale. The study site was an alluvial floodplain with marked hydrologic gradients on which different vegetation types developed. The model was evaluated on accuracy, and could be concluded to perform well. However, model accuracy was remarkably lower for boundary areas between two distinct vegetation types. Subsequent application of the model on a spatially independent data set showed a poor performance that could be linked with the niche concept to conclude that an empirical distribution model, which has been constructed on local observations, is incapable to be applied beyond these boundaries.

1 Introduction

Ecosystems are complex, evolving structures whose characteristics and dynamic properties depend on many interrelated links between direct gradients (nutrients, moisture, temperature), their environmental determinants (climate, geology, topography) and potential natural vegetation, and the processes that mediate between the potential and actual vegetation cover (Baird and Wilby, 1999). Riparian wetlands in

particular exhibit a complex interplay between meteorological, hydrological and biological processes and interactions with the surrounding terrestrial and aquatic systems resulting in a high spatial and short-term variability (Dall’O’ et al., 2001). The conceptual representation shown in Fig. 1 illustrates the relationships between hydrology, the physicochemical environment and vegetation at the local scale. The direct effect of site hydrology on physicochemical site properties, such as soil moisture content, oxygen and nutrient availability determines the productivity and species composition of the site (Venterink et al., 2001; Wassen et al., 2003). Vegetation, however, is not passive to the abiotic setting, but affects site hydrology and physicochemical properties through feedback processes of which transpiration (Engel et al., 2005), soil aeration (Mainiero and Kazda, 2005) and alterations in nutrient loadings (Hill, 1996; Fisher and Acreman, 2004) are just some examples. These localized direct and feedback processes result in spatial and temporal distributions of the abiotic constraints at a higher scale level (Schröder, 2006). Together with intraspecific, interspecific and anthropogenic interactions these distributed abiotic constraints result in vegetation patterns.

Exploring vegetation patterns is a central goal in ecology. Numerous studies examined environmental gradients in relation to vegetation type distributions in various ecosystems (Schulze et al., 1996; Famiglietti et al., 1998; Molina et al., 2004; Rudner, 2005), and different techniques have been developed to quantify vegetation-environment relationships. Canonical ordination (Jongman et al., 1995) for example, is widely applied in ecological studies to detect patterns of variation in vegetation data and quantify the main relations between vegetation and environmental variables. Generalized



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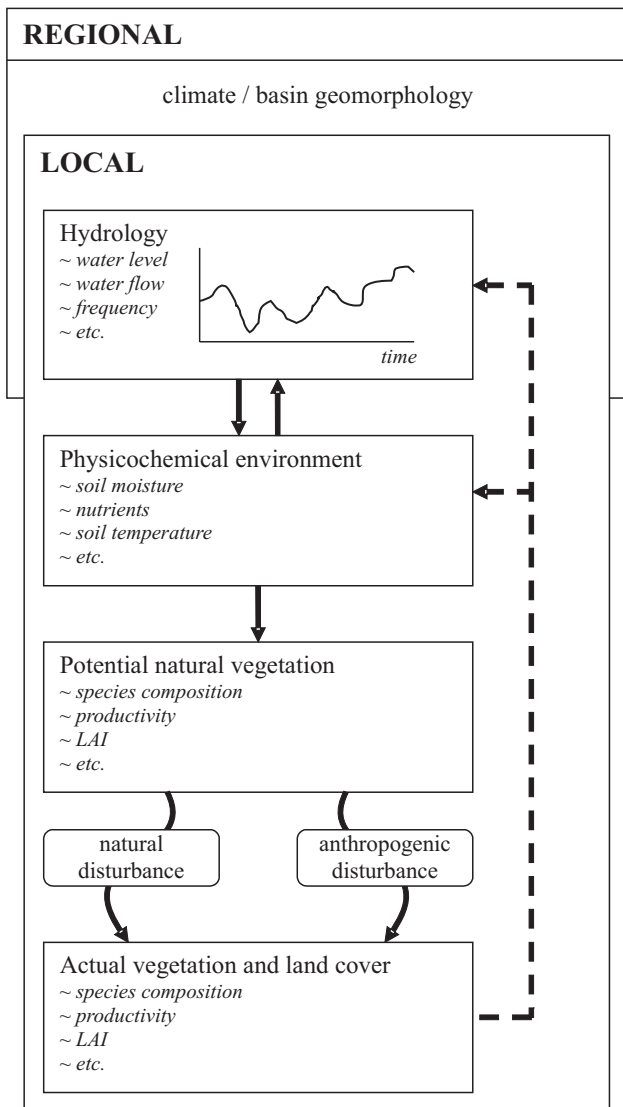


Fig. 1. Conceptual model illustrating the relationships between hydrology, the physicochemical environment and vegetation at the local scale. Legend: full arrows indicate direct effects, broken arrows indicate vegetation feedbacks, and rounded squares and bent arrows indicate exogenous disturbances. Figure adapted from Franklin (1995); Baird and Wilby (1999); Mitsch and Gosselink (2000).

linear models (e.g. multiple logistic regression (Hosmer and Lemeshow, 2000)) are frequently applied to construct distribution models (Austin, 2002; Bio et al., 2002, among others). Distribution models tend to predict spatial distributions of species based on environmental variables (Guisan and Zimmerman, 2000; Guisan and Thuiller, 2005). In this study, an ensemble learning technique named random forests (Breiman, 2001; Prasad et al., 2006), is applied to a spatially distributed data set containing information on environmental conditions and vegetation type distributions. The random forest distribution model was assessed in terms of: (i) its clas-

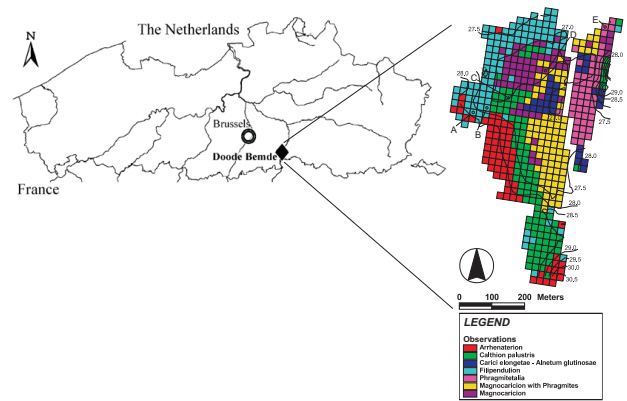


Fig. 2. The Doode Bemde is situated in the valley of the river Dijle. A detailed overview of the topography and the vegetation distribution at the site are shown. The positions of 5 (A-E) piezometers located along a topographical transect are symbolized by o.

sification accuracy, (ii) its applicability on a similar alluvial floodplain, and (iii) its potential to model vegetation distributions based on a reduced number of important environmental variables in groundwater-dependent ecosystems.

2 Description of the study site

A lowland river valley in Belgium called “Doode Bemde” was the research area of this study (Fig. 2). The site is an alluvial floodplain mire in the middle course of the river Dijle, situated approximately 30 m above sea level. The area is bordered by the river Dijle in the west, the Molenbeek, a tributary of the Dijle, in the north and the valley slope with a number of permanent springs in the east (De Becker et al., 1999). The climatic conditions at the site are typically temperate, with an average yearly rainfall of ≈ 800 mm distributed evenly over the year (Verhoest et al., 1997; De Jongh et al., 2006), an average annual pan evaporation of 450 mm, and an average yearly air temperature of 9.8°C (Van Herpe and Troch, 2000). Local conditions at the Doode Bemde have been extensively described by De Becker et al. (1999) and Joris and Feyen (2003).

2.1 Ecohydrological monitoring scheme

During the summer of 1993 and the spring of 1994, plant species occurrences were mapped in the study area. Therefore, the total area of 21.08 ha was subdivided in 519 regular and adjacent 20 m by 20 m grid cells. Mapping was restricted to a selection of 56 plant species of which 45 were typically groundwater dependent (phreatophytes, *sensu* Londo (1988)) and 11 were differential species for several vegetation types at the Doode Bemde. Based on these species cover data, De Becker et al. (1999) applied TWINSPAN (Hill, 1979) in order to define vegetation types. Seven different types were

Table 1. Summary of the vegetation types: abbreviation, name, short description and area.

Nr.	Name	Short description	Characteristic species (English names)	area [ha] (number of grid cells)	
				DB	SN
Ar	<i>Arrhenatherion</i>	High yield potential pasture. Characteristic species include <i>Arrhenatherum elatius</i> (L.) J. & C. Presl, <i>Anthriscus silvestris</i> (L.) Hoffm. and <i>Trifolium dubium</i> Sibth.	Tall Oat Grass Cow Parsley Lesser Trefoil	2.80 (70)	0.83(83)
Cp	<i>Calthion palustris</i>	Species-rich mesotrophic fen meadow dominated by <i>Caltha palustris</i> L., <i>Lychnis flos-cuculi</i> L., and many <i>Carex</i> species.	March Marigold Ragged Robin Sedges	4.24 (106)	0.93 (93)
Ce	<i>Carici elongetae</i> – <i>Alnetum glutinosae</i>	Mesotrophic forest type with dominance of <i>Alnus glutinosa</i> (L.) Gaertn. and a herb layer with <i>Carex elongata</i> L., <i>Carex acutiformis</i> Ehrh. and <i>Lycopus europaeus</i> L.	Alder Elongated Sedge Lesser Pond Sedge Gipswort	1.20 (30)	1.21 (121)
Fi	<i>Filipendulion</i>	Tall herb fen with <i>Filipendula ulmaria</i> (L.) Maxim., <i>Alopecurus pratensis</i> L., <i>Cirsium oleraceum</i> (L.) Scop. and <i>Heracleum sphondylium</i> L.	Meadowsweet Meadow Foxtail Cabbage Thistle Hogweed	4.16 (104)	1.07 (107)
Ph	<i>Phragmitetalia</i>	Highly fertile reedswamps dominated by <i>Phragmites australis</i> (Cav.) Steud.	Reed	2.12 (53)	0.19 (19)
MP	<i>Magnocaricion</i> with <i>Phragmites</i>	<i>Magnocaricion</i> vegetation with <i>Phragmites australis</i> (Cav.) Steud.	Reed Slender Tufted Sedge Lesser Pond Sedge Skullcap	3.72 (93)	0.78 (78)
Ma	<i>Magnocaricion</i>	Tall sedge swamp with <i>Carex acuta</i> L., <i>Carex acutiformis</i> Ehrh., <i>Scutellaria galericulata</i> L. and <i>Phalaris arundinacea</i> L.	Reed Canary Grass Slender Tufted Sedge Lesser Pond Sedge Skullcap Reed Canary Grass	2.52 (63)	–

DB = Doode Bemde; SN = Snoekengracht

distinguished (Table 1), and their spatial distribution can be seen in Fig. 2. All vegetation types are herbaceous, except for *Carici elongetae* – *Alnetum glutinosae* where a tree layer of Common Alder is present. The similarity in species composition between grid cells was compared using the Jaccard index of similarity $JS=c/(a+b+c)$ where c is the number of species shared by both cells, and a and b are the numbers of species unique to each of the cells (Jaccard, 1912). The Jaccard similarity of two grid cells expresses their ecological resemblance concerning species composition, and ranges between 0 (when both cells have unique species) and 1 (when both cells have equal species composition). Averaged JS values are given in Table 2 for the seven different vegetation types. The values of the diagonal elements in Table 2 are a measure of similarity between grid cells of the same vegetation type. Based on these values, patches of *Phragmitetalia*, *Magnocaricion* with *Phragmites* and *Magnocaricion* can be concluded to be more homogeneous in species

composition compared to the other vegetation types which have lower values. Between the different vegetation types, marked differences in similarity can be observed. *Magnocaricion* with *Phragmites* has high similarities with *Phragmitetalia* and *Magnocaricion*. Between the other vegetation types, similarities are generally lower, but nevertheless differences can be observed. *Arrhenatherion* for example, has twice as much species in common with *Filipendulion* than with *Magnocaricion*.

A groundwater monitoring network consisting of 25 piezometers was installed in 1989. Groundwater depths were measured every fortnight during the period 1 January 1991–31 December 1993. Time series of linear interpolated groundwater depths measured at several piezometers (A–E, locations can be seen in Fig. 2) along a topographical transect are plotted in Fig. 3a. A yearly pattern of high summer depths and low winter depths was observed at all piezometers. Based on these time series, hydrological duration lines expressing the probability (%) that a groundwater

Table 2. Jaccard index of similarity between the vegetation types in the Doode Bemde.

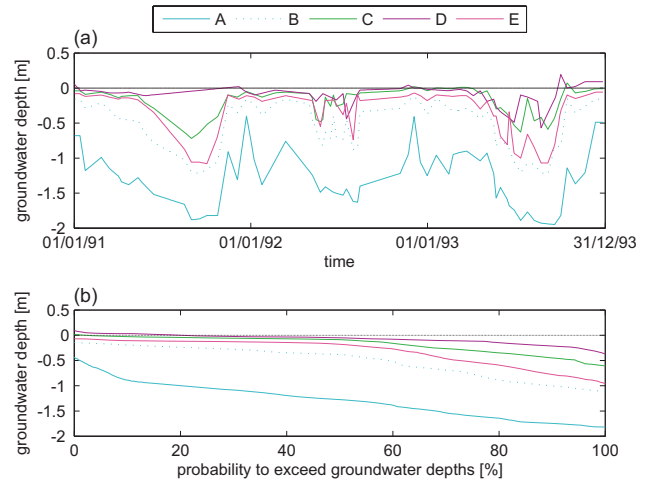
	<i>Ar</i>	<i>Cp</i>	<i>Ce</i>	<i>Fi</i>	<i>Ph</i>	<i>MP</i>	<i>Ma</i>
<i>Ar</i>	0.40						
<i>Cp</i>	0.18	0.37					
<i>Ce</i>	0.11	0.17	0.46				
<i>Fi</i>	0.24	0.21	0.20	0.39			
<i>Ph</i>	0.09	0.19	0.35	0.22	0.55		
<i>MP</i>	0.10	0.19	0.30	0.23	0.44	0.51	
<i>Ma</i>	0.11	0.24	0.30	0.33	0.38	0.42	0.54

depth is exceeded are calculated (Fig. 3b). Groundwater depths corresponding to a probability of exceedance of 50% are yearly average groundwater depths. They differed considerably along the transect (Fig. 3b). At the levee near the river an average value of 1.27 m was measured (piezometer A), which decreased gradually moving further down toward the depression (piezometer B→C→D), with a minimal yearly average groundwater depth of 0.05 m measured at piezometer D in the center of the depression. Fig. 3b also shows different periods of superficial groundwater depths (<0.3 m) in all piezometers, ranging from 75% of the year in piezometer C to 35% of the year in piezometers B and D. Groundwater depths measured in piezometer A are never <0.3 m. Additional to the monitoring of groundwater dynamics, all 25 piezometers were sampled on several groundwater quality variables during a sampling campaign in September 1993 with respect to pH, Cl^- , Ca^{2+} , Fe_{tot} , K^+ , Mg^{2+} , NO_3^- -N, NH_4^+ -N, H_2PO_4^- and SO_4^{2-} . All values are in $[\text{mg L}^{-1}]$ except for pH [-]. A soil type map was made based on 60 drillings to a depth of 1 m, evenly distributed over the study area. Management regime was assessed for each grid cell separately. Four different regimes could be distinguished:

- Yearly mowing in early summer, followed by grazing or mowing of the aftermath;
- Cyclic mowing (once every 5 to 10 years) or not mown at all since at least 5, and up to 10 years;
- No management for at least 10 years;
- Transition from yearly to cyclic mowing.

2.2 Data set

Groundwater depth measurements were used to calculate a dynamic groundwater variable, the mean groundwater depth (MGD) below surface [m]. Values of this variable, together with the groundwater quality variables, were assigned to each grid cell by spatial interpolation of measurement data over the entire area using block kriging (for details, see Bio et al. (2002)).

**Fig. 3.** (a) Time series of the groundwater depth, as monitored by piezometers A–E along a topographic transect (see Fig. 2). (b) Hydrological duration lines expressing the probability that measured groundwater depths are exceeded. The line colours correspond to the vegetation types wherein these piezometers were installed (see Fig. 2).

The spatially explicit variables were structured into a data set. The data set contains $N=519$ measurement vectors $\mathbf{x}_i=(x_{i1}, x_{i2}, \dots, x_{ip})$ consisting of the values of $p=13$ variables describing the abiotic environment:

- Groundwater dynamics: mean groundwater depth (continuous variable);
- Groundwater quality: pH, Cl^- , Ca^{2+} , Fe_{tot} , K^+ , Mg^{2+} , NO_3^- -N, NH_4^+ -N, H_2PO_4^- and SO_4^{2-} . All these variables are continuous;
- Soil: soil type (silt/peat, categorical);
- Management: yearly mowing, cyclic mowing, no management, transition (categorical).

Seven different vegetation types c_1, \dots, c_7 are considered. To each measurement vector \mathbf{x}_i a unique vegetation type $l_i \in \{c_1, \dots, c_7\}$ is assigned. The data set will be denoted as:

$$L = \{(\mathbf{x}_1, l_1), \dots, (\mathbf{x}_N, l_N)\}. \quad (1)$$

2.3 Independent evaluation data set

A spatially independent ecohydrological data set L_{ev} was constructed for a similar valley ecosystem, ‘‘Snoekengracht’’. The Snoekengracht is an alluvial floodplain of the river Velp, situated approximately 15 km from the Doode Bemde. The climatic setting of both nature reserves is very much alike, and local environmental conditions and floral composition are very similar (Bio et al., 2002). The monitoring scheme was largely the same as in the Doode Bemde (Huybrechts and De Becker, 1999), and a grid-based (with a grid size of

10 m by 10 m) data set consisting of $M=501$ elements was constructed, which will be denoted as:

$$L_{ev} = \{(y_1, l_1), \dots, (y_M, l_M)\}. \quad (2)$$

where l_i is the vegetation type assigned to measurement vector y_i . Most vegetation types coincide with those found at Doode Bemde, except for *Magnocaricion* which was not found at Snoekengracht (see Table 1).

3 Distribution model

The distribution model used in this study applies the random forest technique (Breiman, 2001). Random forest is an ensemble learning technique which generates many classification trees (Breiman et al., 1984) that are aggregated to compute a classification. Each classification tree is grown using another bootstrap subset L_i of the original data set L and the nodes are split using the best split variable among a subset of m randomly selected variables (Liaw and Wiener, 2002). The pseudo-code for growing a random forest is given in Appendix A1. The number of trees (k) and the number of variables used to split the nodes (m) are two user-defined parameters required to grow a random forest. An unbiased estimate of the generalization error (the so called out-of-bag error, oob error) is obtained during the construction of a random forest (Appendix A2). Breiman (2001) proved that random forests produce a limiting value for the oob error. As the number of trees increases, the generalization error always converges. The number of trees (k) needs to be set sufficiently high to allow for this convergence. The oob error can be used to optimize the other user-defined parameter m , in order to get a minimal random forest error (Peters et al., 2007). The model outcome is an ensemble of k classification trees which are aggregated based on majority votes to compute the final classification. Since every classification tree votes for a certain vegetation type c_j based on the measurement vector x_i of grid cell i , the probability of occurrence of vegetation type c_j is given by $P(c_j) = N_{c_j}/k$, where N_{c_j} is the number of trees voting for vegetation type c_j , and k the total number of trees. The highest probability of occurrence ($P(c_j)_{\max}$) determines the predicted vegetation type c_j .

Additionally, the random forest algorithm can estimate variable importances (Appendix A3), i.e. variables can be ranked according to their importance in determining vegetation distributions at the study site.

4 Modelling vegetation distributions

4.1 Model construction and results

At first instance the data set L was randomly split into 3 data subsets for 3-fold cross-validation. The model was constructed using the random forest program provided by Breiman and Cutler (2005). User-defined parameters m , the

Table 3. Confusion matrix of the classification made by the random forest distribution model. Predicted vegetation types are compared with the observations at the Doode Bemde.

		Observed						
		<i>Ar</i>	<i>Cp</i>	<i>Ce</i>	<i>Fi</i>	<i>Ph</i>	<i>MP</i>	<i>Ma</i>
Predicted	<i>Ar</i>	55	4	0	4	0	0	0
	<i>Cp</i>	6	89	0	7	0	5	4
	<i>Ce</i>	0	1	19	0	1	4	4
	<i>Fi</i>	9	2	0	82	1	0	7
	<i>Ph</i>	0	2	7	1	45	4	2
	<i>MP</i>	0	2	3	1	4	68	9
	<i>Ma</i>	0	6	1	4	2	12	37

number of randomly selected variables to split the nodes, and k , the number of trees within the random forest, were optimized using the oob error, and suitable parameter values were $m=3$ and $k=1000$. The results include an ensemble of $k=1000$ predictions, one made by each classifier, which are aggregated based on majority votes into a final classification. A confusion matrix summarizing the final classification is given in Table 3, and results are shown in Fig. 4a.

4.2 Model evaluation

4.2.1 Classification accuracy

Out of the 519 grid cells included in the study, the model classified 395 (76.1%) correctly, and 124 (23.9%) incorrectly (Table 3). A κ (Cohen, 1960) value of 0.716 was calculated, indicating a substantial agreement between observations and predictions. A threshold-independent evaluation using receiver operating characteristic (ROC) graphs was performed (Hosmer and Lemeshow, 2000). ROC graphs are useful for visualizing classifier performances (Fawcett, 2006). ROC graphs are two-dimensional graphs in which the true positive rate, tp , is plotted on the y-axis, and the false positive rate, fp , on the x-axis, where

$$tp = \frac{\text{positives correctly classified}}{\text{total positives}} \quad (3)$$

$$fp = \frac{\text{negatives incorrectly classified}}{\text{total negatives}} \quad (4)$$

The area under the ROC curve, abbreviated AUC, is a scalar value between 0 and 1 representing the classifier performance (Fawcett, 2006). Since random guessing produces a diagonal line between (0,0) and (1,1) in ROC space, with an AUC value of 0.5, a classifier with a higher AUC value than 0.5 does better than random guessing. For multi-class ROC graphs, which should be applied here since 7 vegetation types are considered, a methodology described in Fawcett (2006) is used. For each class a different ROC curve is produced, with ROC curve j plotting the classification performance using

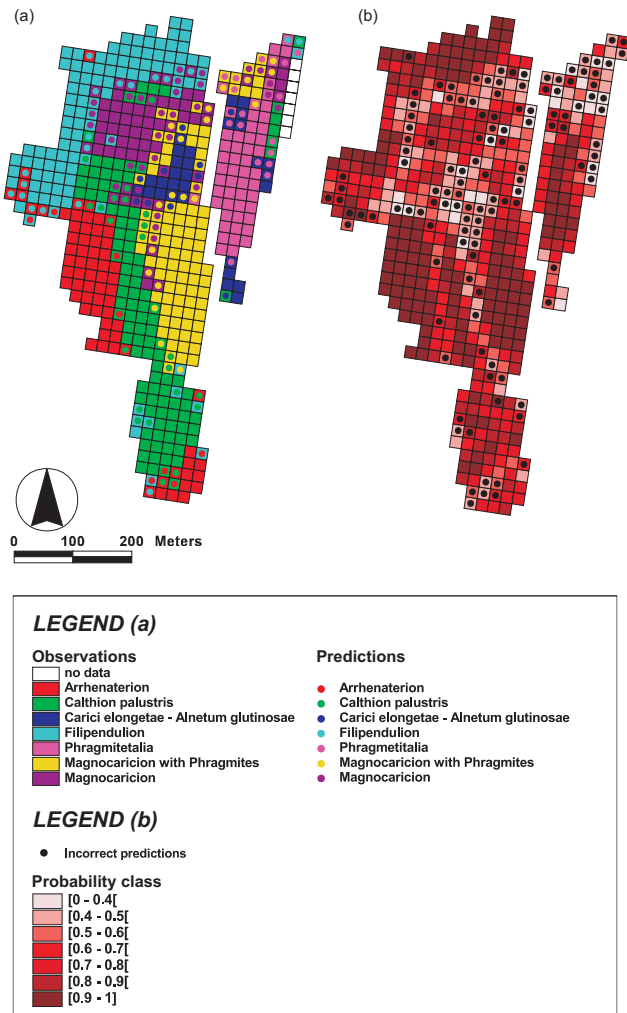


Fig. 4. (a) Observed vegetation types overlaid by the classification made by the random forest distribution model. (b) Modelled probabilities ($P(c_j)_{\max}$) on which the classification is based.

vegetation class c_j as positive and all other classes as negative. For each ROC curve, the AUC can be calculated and averaged over the different classes using class weights based on class prevalences in the test data (Provost and Domingos, 2001):

$$\text{AUC}_{\text{total}} = \sum_{c_j \in C} \text{AUC}(c_j) \cdot w(c_j) \quad (5)$$

where $\text{AUC}(c_j)$ is the area under the class reference ROC curve for c_j , and $w(c_j)$ a weighing factor. Weighing factors are obtained from Table 1. Figure 5 visualizes the ROC curves for each vegetation type. The $\text{AUC}_{\text{total}}$ value equals 0.96 and the random forest distribution model is concluded to perform well.

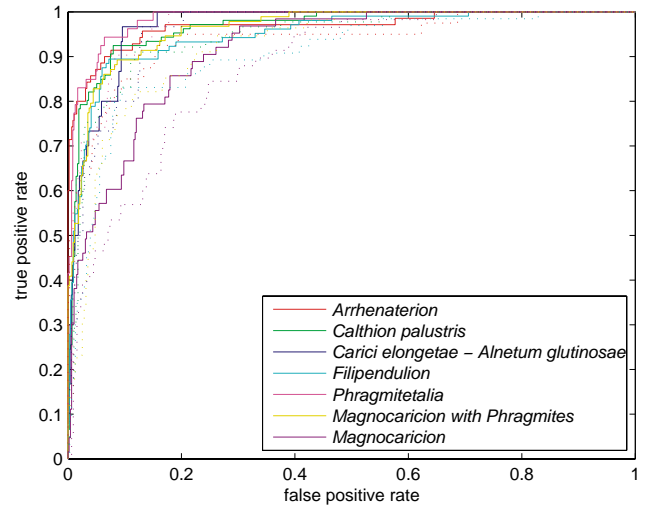


Fig. 5. Receiver operating characteristic (ROC) curves visualizing the classification performances of the 3-fold cross-validated random forest distribution model for the 7 vegetation types (full curves). The $\text{AUC}_{\text{total}}$ equals 0.96. Model performances for boundary cells only are summarized by the dashed ROC curves, yielding an $\text{AUC}_{\text{total}}$ value of 0.92.

4.2.2 Spatially explicit evaluation

For each grid cell, the ensemble of $k=1000$ classification results is aggregated by calculating probabilities of occurrence $P(c_j)$ for all j vegetation types of which the vegetation type with the highest $P(c_j)$ value ($P(c_j)_{\max}$) is the predicted one. As seen in Fig. 6 this decision rule leads to an increasing number of correct classifications with increasing $P(c_j)_{\max}$ values. Indeed, 252 elements are correctly classified with a probability higher than 0.7, whereas only 2 elements are correctly classified with a probability lower than 0.3. 50% of the correctly classified elements are based on probabilities >0.78 . The incorrect classifications show a maximum in the $[0.4,0.5]$ interval, with 1 element incorrectly classified with a probability lower than 0.3, and 28 elements incorrectly classified with probabilities higher than 0.7. 50% of the incorrectly classified elements are based on probabilities >0.55 .

Figure 4b shows the spatial distribution of $P(c_j)_{\max}$ values at the study site in graduated colours. Correctly classified grid cells with high $P(c_j)_{\max}$ values are situated within the central areas of homogeneous vegetation clusters, and $P(c_j)_{\max}$ values tend to decrease toward the boundaries of these areas (see also Fig. 4a). Incorrectly classified grid cells are mainly found where two adjacent vegetation types meet, and are based on low $P(c_j)_{\max}$ values at the central depression and the north-eastern side of the study site. The vegetation types found in these areas are *Carici elongetae-Alnetum glutinosae*, *Phragmitetalia*, *Magnocaricion with Phragmites* and *Magnocaricion*. A Jaccard similarity matrix was constructed for the boundary grid cells only (Table 4). The *JS*

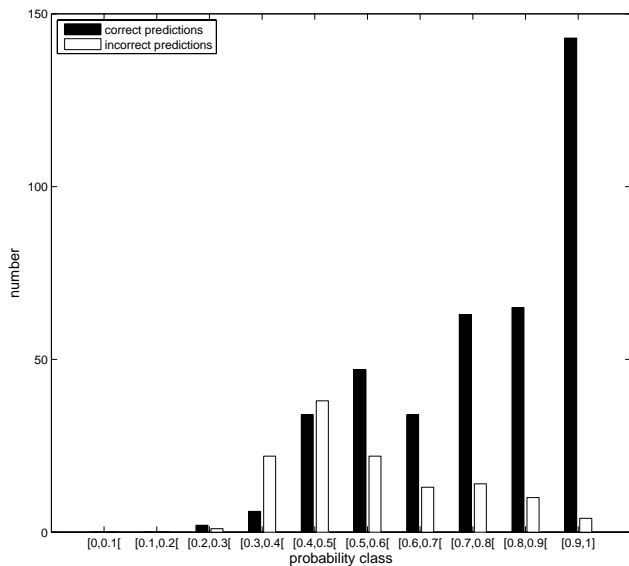


Fig. 6. Probability distribution of correct and incorrect classified grid cells of the Doode Bemde ($N=519$).

values in Table 4 express averaged resemblances in species composition of each boundary grid cell with its 8 neighboring grid cells. Boundary grid cells of *Phragmitetalia*, *Magnocaricion* with *Phragmites* and *Magnocaricion* can be concluded to share a large proportion of their species with *JS* values higher than 0.5. This is reflected in the modelling results, $P(c_j)_{\max}$ values for these grid cells are generally low because comparable numbers of the $k=1000$ classifiers classify these grid cells as *Phragmitetalia*, *Magnocaricion* with *Phragmites* and *Magnocaricion*. Another conclusion should be drawn for isolated grid cells and small isolated vegetation clusters surrounded by another vegetation type (e.g. as occurs along the western border of the study area, see Fig. 4a). These grid cells are frequently incorrectly classified with high $P(c_j)_{\max}$ values, and are the weak point of the random forest distribution model. The worse performance of the model on boundary grid cells can also be seen in Fig. 5, where ROC curves of classification results computed for boundary grid cells only are lower than those computed for the entire data set. The corresponding AUC_{total} value for model performances in boundary areas equaled 0.92, while being 0.96 for the entire study area.

4.2.3 Performance on independent test data

The use of independent test data allows us to assess the model generalization abilities. Edwards et al. (2006) pointed out that cross-validated model accuracies are frequently different from accuracies assessed with truly independent data. It is easy to conclude that the random forest vegetation distribution model, which was trained on the data set L did not classify data set L_{ev} satisfactory. From the 501 ele-

Table 4. Jaccard index of similarity for boundary grid cells between two vegetation types at the Doode Bemde. Non-adjacent vegetation types are indicated by –.

	<i>Ar</i>	<i>Cp</i>	<i>Ce</i>	<i>Fi</i>	<i>Ph</i>	<i>MP</i>	<i>Ma</i>
<i>Ar</i>	0.59						
<i>Cp</i>	0.38	0.60					
<i>Ce</i>	–	0.45	0.66				
<i>Fi</i>	0.34	0.21	–	0.54			
<i>Ph</i>	–	0.18	0.52	0.27	0.67		
<i>MP</i>	–	0.30	0.36	0.19	0.57	0.65	
<i>Ma</i>	–	0.34	0.39	0.57	0.59	0.53	0.66

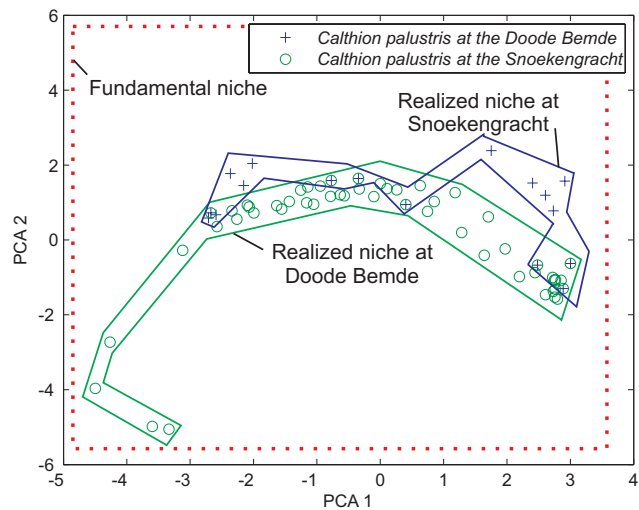


Fig. 7. Conceptual representation of realised niches of *Calthion palustris* at the Doode Beemde and Snoekengracht. The fundamental niche of *Calthion palustris* ranges over all environmental states which would permit to *Calthion palustris* to exist indefinitely (Hutchinson, 1957).

ments included in L_{ev} , only 99 elements were classified correctly (19.8%). This can be explained by the niche concept (Hutchinson, 1957). The fundamental niche of a plant species, and by extension a vegetation type, is defined as an n -dimensional hypervolume (Hutchinson, 1957) in which every point corresponds to a state of the environment which would permit the species to exist and reproduce. Due to interspecific competition species generally occupy only an elementary part of this volume, the realized niche. The niches realized by each of the vegetation types found at the Doode Bemde differ from those realised by the same vegetation types at Snoekengracht. Although similar results were observed for all vegetation types, the example of *Calthion palustris* is given in Fig. 7. Since 13 environmental variables are used in this study, a principle component analysis was performed to reduce dimensions and make results visible. Fig. 7 graphs the component scores of grid cells where *Calthion palustris* was observed on the 2 principle

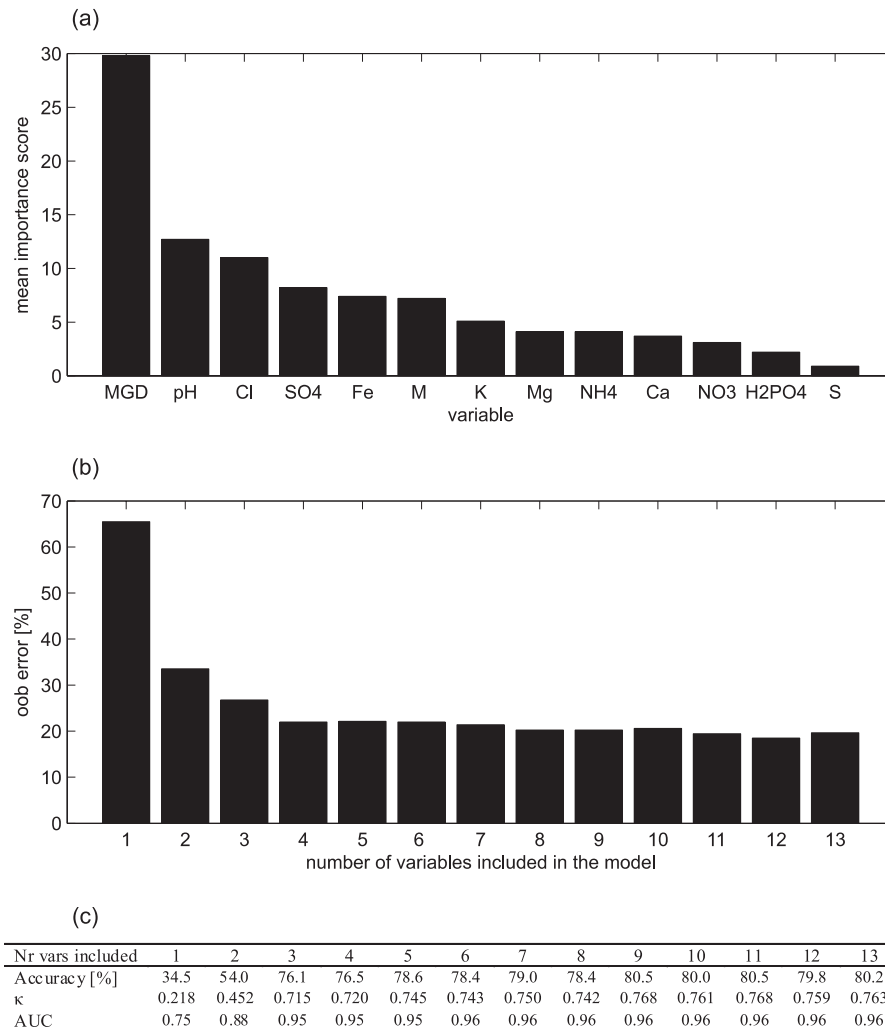


Fig. 8. (a) All variables ranked according to their importance as calculated with the variable importance measure (Appendix A3). M stands for management regime, S represents the variable soil type, and MGD the mean groundwater depth. (b) Oob error of random forest distribution models constructed on data sets with reduced complexity. The model containing only the most important variable (MGD) has an oob error of 65.5%. The oob error decreases gradually when more variables are included. (c) Summarizing table of model performances: accuracy, Cohen's κ and AUC values associated with a decreasing number of variables included.

component axes (cumulatively explaining 70% of variance). Although partly intersecting, two different realized niches can be distinguished. Obviously, a random forest distribution model that is trained on the vegetation distributions at the Doode Bemde and which uses explicit environmental thresholds to compute a classification, cannot perform well on such an independent test data set of an apparently similar ecosystem.

5 Reduction of model complexity

The random forest algorithm includes a procedure to estimate the importance of the independent variables (Appendix A3).

Applying this procedure on data set *L* results in a ranking of all 13 variables according to importance (Fig. 8a). The most important variable is mean groundwater depth. This means that, according to this classification technique, the spatial differences in mean groundwater depths at the Doode Bemde are determinative for the vegetation distributions at the study site. Based on this variable ranking, 13 random forest distribution models were constructed, each on a data set with reduced complexity, i.e. each based on a different number of variables by eliminating the variables in order of importance. Results are summarized in terms of the oob error, and plotted in Fig. 8b. A stable oob error value was found for the models with complexities between 4 and 13 variables. The models constructed on the 3, 2 and 1 most important

variables showed a significant increase in oob error, which is reflected in lower accuracy, κ and AUC values for these models (Fig. 8c).

Based on this result, a simplification of the ecohydrological monitoring scheme for distribution modelling is preliminarily assessed. Since the random forest performances were similar when all 13 or just a part (>3) of these variables were included, there seems to be no need to describe the environmental conditions of the study area by that many variables. Therefore, a simplification of monitoring efforts can be made based on various criteria such as relevance and measurement costs. For similar alluvial ecosystems with groundwater dependent vegetations, the inclusion of groundwater depth together with some – easily measurable – groundwater quality variables such as pH, NO_3^- -N, NH_4^+ -N, and management as environmental variables on which the vegetation distribution modelling is based, is proposed. The independent test data set L_{ev} was redesigned only to include 5 variables: mean groundwater depth, pH, NO_3^- -N, NH_4^+ -N, and management. A random forest distribution model was trained on this data set, and 3-fold cross-validation resulted in an overall accuracy of 72.5% (363 grid cells correctly classified, 138 incorrectly classified), and a κ value of 0.657 and an AUC_{total} value of 0.94 were computed. The reduced random forest distribution model did perform satisfactorily, even when compared to the 3-fold cross-validated results of the random forest model constructed on the entire data set L_{ev} (accuracy=76.6%, κ =0.709, AUC_{total} =0.96).

6 Conclusions

Vegetation patterns arise from the interplay between intraspecific and interspecific biotic interactions and from different abiotic constraints and interacting driving forces and distributions (Schröder, 2006). In this study, we constructed a vegetation distribution model based on spatially distributed environmental variables which were linked with the occurrence of a certain vegetation type. Biotic interactions were only included indirectly, i.e. their effect was included through the observed vegetation distribution pattern, not directly as independent variables underlying the vegetation distribution. As far as classification accuracy of the random forest is concerned, results were satisfactory (AUC_{total} =0.96). Model errors were located in boundary areas ($\text{AUC}_{boundary\ area} = 0.92$) between adjacent vegetation types. A proportion of these errors could be attributed to high similarities between neighboring grid cells. These incorrect predictions were generally based on low probabilities of occurrence of several similar vegetation types. Furthermore, the random forest distribution model cannot be applied beyond the local conditions upon which it was constructed, because realized niches of species/vegetation types do seldom coincide, even between apparently similar sites. This restricts the model's applicability. In order to make it operational on a larger scale

many data would be needed, ranging over the entire ecological amplitude of the modelled attributes. Finally, gradual reductions in model complexity were analysed. Based on these results, a significant reduction of the ecohydrological monitoring scheme could be proposed for a similar groundwater-dependent ecosystem. The random forest distribution model made a reasonably accurate classification (AUC_{total} =0.94) when constructed on spatially distributed measurement of five easily measured environmental variables only.

Appendix A

Random Forest

A1 Growing a random forest

The algorithm for growing a random forest of k classification trees goes as follows:

- (i) for $i = 1$ to k do:
 1. draw a bootstrap subset X_i containing approximately $2/3$ of the elements of the original data set X ;
 2. use X_i to grow an unpruned classification tree to the maximum depth, with the following modification compared to standard classification tree building: at each node, rather than choosing the best split among all variables, randomly select m variables and choose the best split among these variables;
- (ii) predict new data according to the majority vote of the ensemble of k trees.

A2 Out-of-bag error estimate

An unbiased estimate of the generalization error is obtained during the construction of a random forest by:

- (i) for $i = 1$ to k do:
 1. each tree is constructed using a different bootstrap sample X_i from the original data set X . X_i consists of about $2/3$ of the elements of the original data set. The elements not included in X_i , called out-of-bag elements, are not used in the construction of the i -th tree;
 2. these out-of-bag elements are classified by the finalized i -th tree.
- (ii) At the end of the run, on average each element of the original data set X is out-of-bag in one-third of the k tree constructing iterations. Or, each element of the original data set is classified by one-third of the k trees. The proportion of misclassifications [%] over all out-of-bag elements is called the out-of-bag error.

A3 Variable importance

The random forest algorithm can estimate the importance of each variable by using the variable importance measure. Defining variable importances is done by looking at how much the oob error increases when oob data are permuted for one variable while left unchanged for all others. The calculation procedure goes as follows:

- (i) For $i = 1$ to k do (grow a random forest consisting of k classification trees):
 - (1) apply tree i to the n oob elements and count the number of correct classifications over the n oob elements ($C_{i,\text{untouched}}$);
 - (2) for $j = 1$ to p (with p the total number of variables) do:
 - (a) take the n untouched oob elements;
 - (b) randomly permute the values of variable j in the n oob elements;
 - (c) apply tree i to all the j permuted oob elements;
 - (d) count the number of correct classifications ($C_{i,j\text{-permuted}}$);
 - (e) subtract the number of correct classifications of the variable- j -permuted oob elements from the number of correct classifications of the untouched oob elements and divide by the number of oob elements ($\Delta C_{i,j} = (C_{i,\text{untouched}} - C_{i,j\text{-permuted}})/n$);

The results from these iterations are p (number of variables, $j=1$ to p) groups of k (number of trees, $i=1$ to k) $\Delta C_{i,j}$ values. Since trees are independent, correlations among the $\Delta C_{i,j}$ values within the p groups are generally low. Finally:

- (ii) For each of the $j = 1$ to p groups, the mean $\Delta C_{i,j}$ over all $i=1$ to k trees is calculated ($\overline{\Delta C_j} = \sum_{i=1}^k C_{i,j}/k$). The value $\overline{\Delta C_j} \times 100$ is referred to as the “mean importance score” of variable j . The value is positive when $C_{i,\text{untouched}} > C_{i,j\text{-permuted}}$ and negative when $C_{i,\text{untouched}} < C_{i,j\text{-permuted}}$. Mean importance scores have high values when the classification error increases by permuting the values of variable p .
- (iii) Since correlations of the $\Delta C_{i,j}$ scores are generally low within the $j=1$ to p groups, standard errors can be calculated for each of the j groups of $i=1$ to k $\Delta C_{i,j}$ scores. Divide $\overline{\Delta C_j}$ by the standard error to obtain a z -score for variable j , and assign a significance level assuming normality.

Acknowledgements. The authors wish to thank the special research fund (BOF, project nr 011/015/04) of Ghent University, and the Fund for Scientific Research-Flanders (operating and equipment grant 1.5.108.03). We are grateful to W. Huybrechts and

P. De Becker from the Institute of Nature Conservation, Belgium, for providing the data gathered through the Flemish Research Programme on Nature Development (projects VLINA 96/03 and VLINA 00/16).

Edited by: S. Manfreda

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